

Metagenomic analyses reveal no differences in genes involved in cellulose degradation under different tillage treatments

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Abstract

Incorporation of plant litter is a frequent agricultural practice to increase nutrient availability in soil and heavily relies on the activity of cellulose degrading microorganisms. Here we address the question how different tillage treatments affect soil microbial communities and their cellulose degrading potential in a long-term agricultural experiment. To identify potential differences in microbial taxonomy and functionality, we generated six soil metagenomes of conventional (CT) and reduced (RT) tillage-treated topsoil samples, which differed in their potential extracellular cellulolytic activity as well as microbial biomass.

Taxonomic analysis of metagenomic data revealed few differences between RT and CT and a dominance of *Proteobacteria* and *Actinobacteria*, whereas eukaryotic phyla were not prevalent. Prediction of cellulolytic enzymes revealed glycoside hydrolase families 1, 3, 5, 94, auxiliary activity family 8 and carbohydrate binding module 2 as the most abundant in soil. These were annotated mainly to the phyla of *Proteobacteria*, *Actinobacteria* and *Bacteroidetes*.

These results suggest that the observed higher cellulolytic activity in RT soils can be explained by a higher microbial biomass or changed expression levels but not by shifts in the soil microbiome. Overall this study reveals stability of soil microbial communities and cellulolytic gene composition under the investigated tillage treatments.

Introduction

The most abundant organic polymer on earth is cellulose. As a key component of plant cell walls it is highly abundant in all plants mostly in combination with hemicellulose and xylan (Varner & Louis 1989). The degradation of cellulose is an essential ecosystem service and it is of key value to gain more quantitative and qualitative understanding of the role of cellulose breakdown within the carbon cycle and in a changing environment. Cellulases are in general hydrolytic and divided into three major types: β -1,4-endoglucanases (EC3.2.1.4), β -1,4-exoglucanases and β -glucosidases (EC3.2.1.21) (Lynd et al. 2002). Exoglucanases are further divided into cellobiohydrolases or, cellobiosidases (EC3.2.1.91, EC3.2.1.176) and cellodextrinases (EC3.2.1.74). These cellulases differ in their binding and cutting sites on the cellulose fibers or oligomers. Besides hydrolysis, other catalytic modes of action to cut cellulose chains have been identified, namely oxidoreduction, mediated by the enzyme cellobiose dehydrogenase (CBH, EC 1.1.99.18, (Langston et al. 2011)) or induced by a quinone- or glycopeptide-mediated Fenton reaction (Baldrian & Valášková 2008), and phosphorolytic degradation (EC 2.4.1.20, EC 2.4.1.49, (Reichenbecher et al. 1997)). Cellulases generally consist of a catalytic module, classified into glycoside hydrolase- (GH-) families, and often harbour a carbohydrate binding module (CBM), whereas cellulases with oxidoreductive catalytic modules are classified as auxiliary activities- (AA-) family of proteins. The classification into modules is based on amino acid sequence similarity (Henrissat 1991) and a comprehensive description of these module families can be found in the CAZy database (Carbohydrate-active enzymes database; www.cazy.org, (Lombard et al. 2014)).

Different agricultural practices like tillage and fertilization strategies influence world-wide carbon cycles in soil (Lal 2004; Schimel 1995; Schlesinger & Andrews 2000). Tillage, a key component of modern agriculture, disperses added nutrients throughout the soil and aerates the soil, but also affects soil aggregates

53 (Abdollahi et al. 2014). There is a strong ongoing debate on how different tillage treatments impact not only
54 yields, but also soil nutrients and diversity and activity of soil microbial communities. To clarify this question
55 several long term agricultural experiments have been established to assess the adaptation of soils to different
56 tillage treatments. In this context it was shown that reduced tillage (RT) leads to a higher carbon content and
57 higher microbial biomass in the topsoil compared to conventional tillage (CT) (van Groenigen et al. 2011;
58 Küstermann et al. 2013; Alvarez et al. 1995; Kandeler et al. 1999).

59 However, it is currently unclear how different tillage treatments influence the phylogenetic structure of
60 the soil microbial community, in particular those which drive the degradation of polymers, like cellulose,
61 which are major constituents of plant residues used to improve soil quality. Only a few studies have
62 compared conventional tillage to no-tillage systems and these studies have detected several differences in
63 taxonomy and functionality of the associated microbial communities (Souza et al. 2013; Carbonetto et al.
64 2014). For example in a long-term (26 years) experiment, deep tillage practice led to a lower microbial
65 diversity in the top 10 cm of the soil (Silva et al. 2013). However, no in-depth analysis focussing on microbes
66 catalysing carbohydrate degradation has been performed so far in agricultural systems.

67 Thus, in this study we assess the impact of different tillage treatments on microbial communities which
68 drive cellulose degradation in soil. Samples were taken from the topsoil (0-10 cm) of a long-term field
69 experiment, where a split plot-based study has been established with conventional (CT; 25 cm working
70 depth) and reduced tillage (RT; 8 cm working depth) treatments (Meyer-Aurich et al. 2009; Küstermann et al.
71 2013). One month before sampling, corn had been harvested and the remaining plant litter had been
72 incorporated into the soil with the different tillage treatments. Due to the different tillage depths the amount
73 of plant residues and consequently the amount of straw in the topsoil differed between the two treatments
74 (Tebrügge et al. 1999). To assess the influence of different tillage treatments on the enzymatic activity of key
75 cellulolytic enzymes in the soil, the potential enzymatic activities of β -glucosidase and cellobiohydrolase were
76 measured using methylumbeliferone-labeled substrates.

77 We explored the influence of long-term conventional and reduced tillage on the taxonomic and functional
78 diversity of soil microbial communities, which are involved in cellulose degradation. As it is well accepted

that cellulases occur ubiquitously in all kingdoms (<http://www.cazy.org>, Lombard et al. 2014) and show a high genetic diversity, we employed a shotgun sequencing approach to identify the different cellulases present and the microbes harbouring these genes in a qualitative (diversity) and semi-quantitative manner. The analysed top soils differed in their amount of total organic carbon (Küstermann et al. 2013). Therefore, we hypothesized to find a higher relative abundance and diversity of genes coding for cellulases in soil samples from RT compared to CT.

Experimental Procedures

Site description and soil sampling

Soil was sampled from an agricultural field experiment, established in 1992 at a research farm in Scheyern, 40km north of Munich (Germany) (Meyer-Aurich et al. 2009; Küstermann et al. 2013). From this field experiment, two tillage treatments were analysed: conventional tillage with moldboard plough (25 cm working depth) and reduced tillage with rotary harrow (8 cm working depth). The experiment has been set up with three independent replicates (split plot design). Ten-year-average precipitation and temperature are respectively 792.3 mm and 8.3 °C (2000-2010). The soil is a Luvisol with a pH of 6.3, consisting of 2.2% coarse sand, 17.0% fine sand, 55.4% silt and 25.4% clay.

The soil was sampled in November 2012, one month after corn harvest and incorporation of the plant residues into the soil. Of every plot a composite sample of five soil cores to a depth of 10 cm was taken with a soil auger of 5 cm diameter, and sieved with a 3-mm sieve. Afterwards, a subset of the soil was mixed and immediately stored on dry ice and subsequently at -80 °C before DNA extraction. The remaining soil was stored at 4 °C.

Enzymatic activity assays, biomass and chemical measurements

Potential enzymatic activity was measured one day after sampling of soil stored at 4 °C. Potential cellulose degradation activity of the microbial community was assessed by adding methylumbeliferone-complexed (MU-)- β -D-glucoside and MU- β -cellobioside to soil solutions (Pritsch et al. 2005). The optimum substrate

105 concentration (C_{opt}) and incubation time (T_{opt}) for degradation of both substrates were determined during a
106 pilot experiment; C_{opt} =800 μ M MU- β -cellobioside and 500 μ M MU- β -D-glucoside, T_{opt} = 120 min for both.
107 Fluorescence measurement after incubation of the substrate with soil solution was done with a
108 spectrophotometer at excitation wavelength λ =365 nm and emission wavelength of λ =450 nm. Fluorescence
109 values were corrected for soil-autofluorescence or fluorescence inhibition by soil particles (quenching) and
110 background noise.

111 For measurement of microbial carbon, water extractable organic carbon (WEOC) was determined
112 before and after chloroform fumigation (Joergensen 1996). For determination of WEOC, soils were mixed
113 with 0.01M $CaCl_2$ in a 1:5 ratio, incubated in an overhead shaker for 40 min at room temperature, filtered,
114 and stored at -20 °C until measurement with the Dimatoc 100 (DimatecAnalysestechnik GmbH, Germany).

115 For measurement of total carbon (TC) and nitrogen (TN) soil was dried at 40 °C for ca. 5 days and
116 ground to powder shaking in a tissue lyzer at 30 Hz for 3-9 min. 20-25 mg of ground soil powder was
117 weighed in duplicate, wrapped in aluminium and subjected to elemental analysis using an Elementar Vario
118 EL III instrument in combustion mode.

119

120 DNA isolation & Pyrosequencing

121 Soil DNA was extracted from 300 mg of frozen soil samples using the DNA-isolation kit ‘Genomic DNA
122 from soil’ NucleoSpin Soil Kit (Macherey– Nagel, Germany) according to the protocol of the manufacturer.
123 DNA was stored at -20 °C until further processing. Library preparation was performed according to the
124 Roche protocol “Rapid Library Preparation Method Manual” using Roche Molecular Identifier (MID)
125 adapters as barcodes. Amplification of DNA via emulsion PCR was done according to the Roche protocol
126 “emPCR Method Manual - Lib-L Large Volume (LV)”. Pyrosequencing was carried out on the Genome
127 Sequencing (GS) FLX+ instrument, using a GS FLX Titanium sequencing kit XL+. Image- and signal
128 processing was performed by the provided Roche software. The sequences are deposited in SRA under the
129 BioProject ID: PRJNA235154.

130

131

132 Real-time PCR assays

133 Quantitative real-time PCR (qPCR) was performed using primers for amplification of the bacterial 16S
134 ribosomal gene (Bach et al. 2002) and the fungal Internal Transcribed Spacer (ITS) regions (White et al.
135 1990). The reaction mixtures (total volume 25 μ l) for quantification of the 16S ribosomal gene consisted of:
136 12.5 μ l Power SYBR Green master mix (Life Technologies), 5 pmol of each primer, 0.5 μ l 3% bovine serum
137 albumin (BSA) and 2 μ l soil DNA template. The reaction mixtures (total volume 25 μ l) for quantification of
138 the ITS gene consisted of: 12.5 μ l Power SYBR Green master mix (Life Technologies), 10 pmol of each
139 primer and 2 μ l soil DNA template. For quantification, standard curves were calculated using serial dilutions
140 (10^2 to 10^6 gene copies μ l⁻¹) of a plasmid containing *Fusarium oxysporum* DNA (for fungal qPCR) or
141 *Pseudomonas putida* (for bacterial qPCR). PCR detection limit was assessed to 10 gene copies μ l⁻¹. In order
142 to prevent PCR inhibition, the optimal dilution for each amplification assay was determined by dilution series
143 of DNA extracts (data not shown). The qPCR assays were performed in 96-well plates. All PCR runs started
144 with a hot start at 95°C for 10 minutes, then either 40 cycles of 94°C for 30 s, 55°C for 30 s and 72°C for 30 s
145 for ITS samples or 40 cycles of 95°C for 45 s, 58°C for 45 s and 72°C for 45 s for 16S qPCR. Finally melting
146 curve analyses were carried out for all samples by a final cycle of 95°C for 15 s, 60°C for 30 s and 95°C for
147 15 s. The amplification efficiency was calculated from the formula $Eff = [10^{(-1/slope)} - 1] * 100$, and resulted in
148 the following efficiencies: 16S: 102% and ITS: 87%.

149

150 Analysis of sequencing data and prediction of cellulolytic enzymes

151 Sff-files were separated by MID adapters, using 454/Roche SFF Tools. Trimming of reads was carried out
152 using a modified DynamicTrim (Cox et al. 2010) as supplied by MG-RAST (Meyer et al. 2008) using the
153 following settings: minimum Phred score= 15, maximum number of bases below minimum Phred score=5,
154 minimum read length=50. Remnant adapter sequences and duplicated reads were removed using Biopieces
155 (www.biopieces.org) and cd-hit (Fu et al. 2012), respectively. For taxonomic and functional annotation,
156 filtered reads were blasted against both the National Center for Biotechnology Information (NCBI) non-

157 redundant protein database and the Kyoto Encyclopedia of Genes and Genomes (KEGG)-database (June
158 2011) using Diamond (Buchfink et al. 2014) as well as against the Silva SSU-database using Blastn (expect
159 (e)-value threshold= 10^{-4}). Mapping of the top 25 blast results (i.e. hits with lowest e-value) to taxonomic and
160 functional annotations was carried out with the MetaGenome ANalyzer (MEGAN, Version 5.2.3) software
161 (<http://ab.inf.uni-tuebingen.de/software/megan5/>). During the MEGAN analysis the following settings were
162 used: Min support=1, Min score=50, Top%=10, Min-Complexity Filter=0. Visualization of data was
163 performed in R (R Core Team 2013). To obtain a relative abundance of reads, the number of reads was
164 divided by the total number of filtered reads per sample and multiplied by 100.

165 For prediction of catalytic modules (GH and AA) and carbohydrate binding modules (CBM) in the
166 reads, protein open-reading frames were predicted using FragGeneScan (Rho et al. 2010) and translated to
167 protein sequences. Different families of catalytic modules and CBMs harboured by the key enzymes in
168 cellulose degradation β -1,4-endoglucanases, β -1,4-exoglucanases, β -glucosidases, cellodextrin
169 phosphorylases and cellobiose dehydrogenases, were selected from the CAZy database (www.cazy.org
170 (Lombard et al. 2014)). Protein Hidden Markov Model (HMM)-logos can be used to scan protein sequences
171 using hmmscan (Eddy 2011). For the selected families, several HMM-logos are available in the Protein
172 family (Pfam) A 26.0 database (Punta et al. 2012) and in the DataBase for automated Carbohydrate-active
173 enzyme Annotation (dbCAN, Yin Y et al., 2012), but can also be personally built using hmmbuild (which is
174 contained within the HMMER version 3.0 (March 2010), packaged together with hmmscan;
175 <http://hmmer.org/>) and an alignment-file with sequences containing the corresponding module domain as
176 input (alignment files were generated using sequences from CAZy for each corresponding family).

177 Because the selected module families also contain a varying amount of enzymes other than
178 cellulases, we decided to validate the specificity of these HMM-logos for cellulases by scanning them against
179 a list of positive or negative cellulase sequences. These lists have been assembled by downloading CAZy
180 entries for all glycoside hydrolase, carbohydrate binding and auxiliary activity families that contain
181 cellulolytic proteins. A list of 1283 entries was then classified as positive, having a matching EC number (as
182 mentioned in the introduction) and a list of 1390 entries as negative where the EC number did not match. Of

the positive sequences, 662 were eukaryotic. For the families AA8 and AA9, entries that mentioned activity on cellulose in the name were included in the positive list. Taking all Genbank Ids for these positive and negative entries, the corresponding protein sequences were downloaded from NCBI. Finally, these lists were scanned with the three sets of HMM-logos (from Pfam A, dbCAN or personally built) and the HMM-logos that annotated most of the positive sequences (see supplementary **Table S1** in bold) were used to scan the translated protein sequences from the metagenomes (e-value threshold= 10^{-4}).

To limit the number of false positive predictions we refined the search for cellulases by performing an additional quality control step, i.e. blasting all HMM-annotated protein sequences against the above-mentioned positive sequences database using blastp (e-value threshold= 10^{-5}). To obtain a relative abundance of reads, the number of reads was divided by the total number of filtered reads per sample and multiplied by 100,000.

For taxonomic annotation, all reads with a predicted cellulolytic module, i.e. after scanning against the HMM-logos and the positive sequences database, were blasted against the NCBI non-redundant protein database using Diamond (minscore=50) and mapped with MEGAN5 (parameters as previously announced).

Statistical analysis of sequencing data and diversity analysis

To detect global differences between the metagenomes, principal component analysis was performed on the relative order count data of metagenomic reads annotated to the NCBI non-redundant protein database and the KEGG database and analysed with MEGAN5. Significant differences of soil measurements and annotated read counts between CT and RT soils were determined by unpaired t-test statistics, and adjusted using the Bonferroni correction for metagenomic data, using R (R Core Team 2013). Before testing for differences between soil parameter-measurements, QQ-plots and Shapiro-Wilk tests for normal distribution were performed. For the PCA-plot, the data was square-root transformed, as the QQ-plot showed no normal distribution. Differences were regarded significant when the adjusted *p*-value was smaller than 5 % ($P<0.05$). In addition, only taxa or pathways with a relative abundance of at least 0.05% in one of the replicates were included. Due to the low amount of predicted cellulases, no abundance cut-off was used to

209 detect possible differences between tillage treatments. Shannon indices were calculated using the vegan
210 package in R (Oksanen et al. 2013).

211

212 **Results**

213 Microbial biomass and enzymatic activities

214 As expected, we found higher total soil carbon and nitrogen contents under conditions of reduced tillage
215 ($C=1.62 \pm 0.13\%$, $N= 0.17 \pm 0.01\%$) compared to conventional tillage treatment ($C=1.08 \pm 0.07\%$, $N= 0.12 \pm$
216 0.01% , $P=0.004$ respectively $P=0.004$). In addition, reduced tillage leads to significantly higher microbial
217 biomass compared to soil under conventional tillage, as measured by carbon analysis (0.28 ± 0.03
218 respectively 0.16 ± 0.02 mg g^{-1} dry weight (DW), $P=0.004$, **Fig. 1a**) and by qPCR of both bacterial ($8.37 \pm$
219 0.49×10^{10} respectively $6.82 \pm 0.19 \times 10^{10}$ copies g^{-1} DW, $P=0.007$) and fungal DNA ($5.58 \pm 1.03 \times 10^8$
220 respectively $3.40 \pm 0.40 \times 10^8$ copies g^{-1} DW, $P=0.027$). The high ratio of bacterial 16S rRNA genes to fungal
221 ITS genes was not influenced by the treatments. Potential extracellular enzyme activity measurements
222 showed that CT results in a lower activity than RT for both β -glucosidase (278.7 ± 40.9 respectively $545.1 \pm$
223 86.1 nmol MU $h^{-1} g^{-1}$ DW, $P=0.049$) and cellobiohydrolase (5.0 ± 1.6 respectively 28.8 ± 2.7 nmol MU $h^{-1} g^{-1}$
224 DW, $P=0.002$) (**Fig. 1b and 1c**).

225

226 Taxonomic and functional annotation of metagenomes

227 Shotgun sequencing of the six DNA libraries (3 independent replicates per tillage treatment) resulted in an
228 overall amount of 0.5 Gigabases of data, which corresponds to a mean of 157,106 filtered reads per replicate
229 with an average length of 410 bps (**Table S2**). To account for the different number of reads between the
230 biological replicates, relative abundances per replicate were calculated. Taxonomic annotation of the
231 metagenomes was performed by blasting all filtered reads against the Silva SSU database and assigning the
232 taxonomic rank with MEGAN5. Due to the low amount of ribosomal sequences in the metagenomes (0.12%),
233 only a classification at kingdom level was performed and revealed a dominance of Bacteria (86.15%)
234 followed by Eukaryota (13.39%) and Archaea (0.45%) (data not shown). 61% of the reads could be

235 taxonomically annotated to at least super-kingdom level with the NCBI non-redundant protein database.
236 Therefore, all further taxonomic annotations refer to the NCBI database. At super-kingdom level the majority
237 of reads was mapped to Bacteria (98.03%) followed by Eukaryota (1.08%), Archaea (0.73%) and Viruses
238 (0.05%). Exploratory rarefaction analysis at each taxonomic level revealed an acceptable coverage of
239 diversity on the order level for all six samples (**Fig. 2**). Taxonomic analysis showed that all soil metagenomes
240 are dominated by *Proteobacteria* and *Actinobacteria* that together make up for 26.3% of all reads (**Fig. S1a**).
241 Further prevalent bacterial phyla were *Bacteroidetes* (3.3%), *Acidobacteria* (3.3%), *Verrucomicrobia* (1.7%),
242 *Gemmatimonadetes* (1.4%), *Planctomycetes* (1.3%) and *Chloroflexi* (1.1%). At the order level
243 *Actinomycetales*, *Rhizobiales*, *Myxococcales*, *Burkholderiales* and *Planctomycetales* accounted for 13.5% of
244 all annotated reads (**Fig. S1b**).

245 The principal component analysis performed on the relative order count data from the metagenome
246 reads (**Fig. S2**) revealed no clear clustering of replicates from conventional and reduced tilled soils. In
247 accordance, no significant change in alpha-diversity (Shannon indices: conventional tillage: 3.58 ± 0.02 ,
248 reduced tillage: 3.55 ± 0.04) between the metagenomes was detected. Significantly different abundances of
249 taxonomically annotated metagenome reads between CT and RT soils are shown in Table S3. Genes
250 annotated to originate from the order *Ktedonobacteraceae* were more abundant in conventional tillage-treated
251 soil compared to reduced tillage-treated soil ($P=0.003$).

252 For functional annotation, the metagenomes were mapped to the KEGG database, to which 35.21 %
253 of the reads could be annotated, and visualized with MEGAN5. Two-component system, Purine metabolism,
254 ABC transporters, Nitrogen metabolism and Pyrimidine metabolism were among the most abundant
255 pathways and account for 6.14 % of all reads (**Fig. S3**). Soils under conventional tillage contained
256 significantly more genes involved in Carbohydrate metabolism ($P=0.008$), Xenobiotics Biodegradation and
257 metabolism ($P=0.0006$), Arachidonic acid metabolism ($P=0.006$), Tyrosine Metabolism ($P=0.010$) and Drug
258 Metabolism ($P=0.0006$, **Table S3**).

259
260

Prediction of cellulolytic enzymes from metagenomes of conventional and reduced tillage-treated soils

When focusing on the enzymes involved in cellulose degradation within the starch and sucrose metabolism pathway of KEGG, the β -glucosidase (EC 3.2.1.21) and endoglucanase (EC 3.2.1.4) were most abundant, with 270 reads (mean relative abundance $0.071\% \pm 0.008$) for conventional and 410 reads (mean relative abundance $0.073\% \pm 0.003$) for reduced tillage annotated to β -glucosidase. 123 reads (mean relative abundance $0.032\% \pm 0.004$) for conventional and 186 reads (mean relative abundance $0.033\% \pm 0.001$) for reduced tillage were annotated to endoglucanase. These numbers revealed no significant difference between conventional and reduced tillage. Much fewer reads were annotated to exoglucanase (EC 3.2.1.91, CT: $0.001\% \pm 0.0006$ and RT: $0.002\% \pm 0.0018$) and cellobiose phosphorylase (EC 2.4.1.20, CT: $0.002\% \pm 0.0014$ and RT: $0.001\% \pm 0.0005$) and none to cellobiose dehydrogenase (EC 1.1.99.18). In the KEGG database there is no reference KEGG Orthology groups for cellobiosidase (EC 3.2.1.176) or cellodextrinase (EC 3.2.1.74).

More extensive screening of the metagenomes for cellulases was performed using hmmscan. To identify the most sensitive HMMs for cellulolytic proteins we benchmarked publically available as well as personally built HMMs against a set of positive and negative CAZy proteins (see Experimental Procedures). This revealed that most HMMs detected also many negative CAZy proteins. Using the most sensitive HMMs (**Table S1** in bold), 5,906 reads could be annotated to a catalytic or carbohydrate binding module. To reduce the number of false positive predictions using hmmscan, an additional filtering step was performed by scanning the predicted cellulases against the same set of positive CAZy proteins (**Table S4**). After this quality control step 2,021 reads remained with a predicted catalytic or carbohydrate binding function, which corresponds to 0.214 % of the total amount of metagenome reads. The percentage of total annotated cellulases was not different between treatments. A large variety of catalytic modules and CBMs that are involved in cellulose degradation were found in the metagenomes of this agricultural soil; 18 glycoside hydrolases, 4 auxiliary activities and 14 carbohydrate binding modules (**Fig. 3**). After quality control, most hits were found for GH1, GH3, GH94, AA8 and CBM2. Statistical analysis revealed significantly more reads annotated to CBM11 in conventional tillage-treated soil than in reduced tillage-treated soil ($P=0.020$).

286 Conversely, reduced tillage-treated soil contained more reads annotated to GH48 than conventionally tilled
287 soil ($P=0.032$).

288 The taxonomic affiliation of the pooled cellulase-annotated sequences (from Figure 3) is shown in
289 **Fig. 4** on phylum and order level, and reflects the overall abundance of phyla in the metagenomes (**Fig. S1a**).
290 18.56% of all cellulase reads mapped to the *Proteobacteria*, 11.23% to *Actinobacteria* and fewer to
291 *Bacteroidetes* (8.61%), *Cyanobacteria* (2.23%), *Gemmatimonadetes* (2.18%), *Verrucomicrobia* (2.13%),
292 *Acidobacteria* (1.73%), *Firmicutes* (1.53%) and *Chloroflexi* (1.53%). *Actinobacteria*, *Bacteroidetes*,
293 *Cyanobacteria* and *Firmicutes* show a high abundance here, whereas they are not so abundant in the total
294 metagenome (respectively 7.1%, 3.3%, 0.7% and 0.8% of total reads, **Fig. S1a**). Results on order level
295 support the importance of the role which *Actinobacteria* (*Actinomycetales*), *Proteobacteria* (*Rhizobiales* (α),
296 *Burkholderiales* (β), *Xanthomonadales* (γ), *Myxococcales* (δ)), *Bacteroidetes* (*Cytophagales*,
297 *Sphingobacteriales*) and *Verrucomicrobia* (*Verrucomicrobiales*) play in cellulose degradation. Of the reads
298 that were predicted to contain cellulolytic modules 0.71% mapped to fungi (*Ascomycota* and *Basidiomycota*).

299 Finally, the taxonomic affiliation for the individual most abundant cellulolytic families (GH1, 3, 5,
300 94, AA8 and CBM2) and the difference between conventional and reduced tillage is shown separately in
301 Figure S4. The low number of reads leads to a high variation in the amount of annotated reads among
302 replicates and treatments.

303

304 **Discussion**

305 Relative abundance and diversity of genes coding for cellulolytic enzymes

306 We predicted a multitude of cellulolytic enzymes in our study, which reflects the complex nature of cellulose
307 degradation in soil. The functional annotation of cellulase genes agrees with other studies searching for
308 catalytic modules and CBMs in aerobic environments, where also genes of GH family 3, 5 and 9 have been
309 identified (DeAngelis et al. 2010; Anderson et al. 2012; Duan & Feng 2010; Nyysönen et al. 2013). This is
310 also true for their taxonomic annotation, as cellulolytic microorganisms have been found in the phyla of
311 *Proteobacteria*, *Firmicutes*, *Bacteroidetes* and *Actinobacteria* (Woo et al. 2013; Anderson et al. 2012;

312 Nyssönen et al. 2013; Berlemont & Martiny 2013). Also the order of *Actinomycetales* contains well known
313 cellulose degraders (Větrovský et al. 2014; Solans & Vobis 2003), as does the order of *Myxococcales*
314 (Reichenbach et al. 2006).

315 To our knowledge, a similar refining analysis on HMM-annotated reads as presented here (**Fig. 3**)
316 has not previously been done. We find a highly variable specificity of each HMM-motif for cellulolytic
317 function, which ranges from 0.0 to 100.0 % of specificity (**Table S4**). Our results show that such an analysis
318 is necessary to get a more exact prediction of gene function, as results can depend heavily on the nature of the
319 HMM. However, sequence similarity scoring may not be the optimal method to annotate diverging protein-
320 coding sequences. To define the actual activity of the predicted cellulases, other approaches like cloning and
321 subsequent functional characterization, transcriptomics and proteomics are necessary in addition.

322 To further assess the question what percentage of predicted cellulases can be more reliably annotated,
323 we compared the sequences predicted to harbour cellulolytic domains (GH and AA) to the sequences that
324 mapped to KEGG modules. This revealed that 19.5% of all sequences predicted by the HMMs before
325 filtering and 28.8% after filtering with the curated database were annotated by the KEGG database as well.
326 This indicates that at least a part of the identified sequences is potentially functional in the soil.

327 In the overall metagenome, 1.08 % (NCBI nucleotide database) or 13.39 % (Silva SSU database) was
328 annotated to eukaryotes. Indeed, the qPCR on the soil DNA showed a very low abundance of fungal ITS
329 sequences, especially compared to bacterial 16S sequences. These results are comparable to earlier measured
330 ITS-sequence abundance in agricultural soil (Ng et al. 2012). In accordance with the overall metagenome,
331 0.93% of the predicted cellulases in the metagenome were annotated to eukaryotes (0.71% was annotated to
332 fungi) when using the NCBI non-redundant protein database. However, 12% of the predicted cellulases were
333 annotated to a CAZy module which contain mostly eukaryotic sequences (GH7, CBM1 and 49, AA3, 8 and
334 9). Discrepancies between the annotation results of different databases (NCBI versus Silva or NCBI versus
335 CAZy) clearly show the database biases towards sequences from often-studied organisms. In a high-
336 throughput approach, numbers of reads annotated to taxa or functions are only realistic if a sufficient
337 database-representation of the corresponding genes exists. Unfortunately, this is not the case, as databases are

338 biased towards culturable bacteria (Wooley et al. 2010) and care should be taken when assessing results
339 based on similarity to database contents. In addition, the discrepancy between the annotated abundance of
340 fungi by the Silva SSU database and the qPCR results presented here, might be explained by the detection of
341 a different fungal sequence (i.e. 18S in the Silva SSU database, and ITS by the PCR assay). Besides database
342 annotation biases, fungal genes contain many intronic sequences, which require long reads for accurate
343 annotation. A transcriptomic approach might resolve the hidden contribution of fungi in the soil metabolic
344 function, as is evidenced by other studies targeting soil transcripts (Kellner et al. 2010; Baldrian et al. 2012).

345 The low amount of fungi annotated in the overall metagenome and the cellulase-annotated sequences
346 is striking when one considers the role of fungi in biomass degradation. In forest soils, fungal cellulases are
347 known to be important for cellulose degradation (Bailey et al. 2002; Buée et al. 2009). It is however known
348 that agricultural soils harbour lower amounts of fungi than undisturbed soils, as tillage disrupts soil fungal
349 communities (Miller & Lodge 1997). Furthermore, the yearly application of mineral fertilizers and fungicides
350 negatively influences the amount of fungi in the sampled soil.

351

352 Influence of tillage treatment on soil metagenomes

353 We hypothesized to find a higher abundance and diversity of cellulase genes in RT soil compared to CT soil,
354 based on the assumption of higher substrate availability under RT. Few significant differences were found
355 between the metagenomes, e.g. genes related to GH48 were exclusively found in RT soil, whereas CT soil
356 harboured more CBM11 genes (**Fig. 3**). GH48-family proteins are generally harboured by bacteria and
357 anaerobic fungi, and are always present in cellulosomes. Furthermore, GH48 cellulases are considered the
358 key component of various cellulolytic systems (Devillard et al. 2004; Olson et al. 2010; Izquierdo et al.
359 2010). CBM11-family proteins are only known to be present in bacterial cellulases. However, so far not
360 much is known of the function of cellulase diversity or specific groups of cellulases in the degradation
361 process in soil. Future research will have to elucidate the specific contribution of each family. Another
362 finding is that more reads were annotated to the class *Ktedonobacteria* in CT soil than in RT soil (**Table S3**).
363 *Ktedonobacteria* are gram-positive, aerobic organisms that produce branched vegetative mycelia (Yokota et

364 al. 2012). However, not much is known of their physiology and growth conditions or ecological importance.
365 Using the KEGG database for functional annotation of the metagenomes we detect significantly more reads
366 for carbohydrate metabolism and biodegradation of xenobiotics in CT soil compared to RT soil. This is an
367 interesting finding but difficult to interpret as there are many downstream pathways and none of them differ
368 between tillage treatments.

369 Most detected differences were however based on low read counts, questioning their ecological
370 relevance. For this a metagenomic assessment with higher resolution or more biological replicates can be
371 helpful to perform more powerful statistics. In general, this in-depth analysis of soil metagenomes from RT
372 and CT soils reveals no global differences in taxonomic and functional content, and none in alpha-diversity.

373 This is to our knowledge the first metagenomic analysis of the same agricultural soils under different
374 tillage practices. It is therefore plausible that a change in tillage practice has far less impact on a functioning
375 microbial community than a comparison of tillage versus no-tillage, for which several differences in
376 microbiome structure and taxonomic composition were detected (Carbonetto et al. 2014; Souza et al. 2013).
377 In this respect we can hypothesize that conditions below 10 cm in the RT soil in this study might resemble
378 conditions under no-tillage. Indeed, Angers *et al.* showed that the surface layer and deeper soil layers can be
379 differentially affected by tillage treatment (Angers et al. 1997). In the subsoil of RT, no extra carbon is
380 incorporated. Accordingly, no differences in soil organic carbon between CT and RT were observed at deeper
381 soil levels (18-25 cm) at this experimental site (Küstermann et al. 2013). To confirm whether the subsoil of
382 RT is affected in a similar way as no-tillage soil, future research is required.

383 Our results indicate that tillage intensity does not strongly influence microbial community structure
384 in the timeframe of 20 years, and implies that these microbial communities in agricultural soils are stable to
385 some extent in response to long-term differences in tillage treatments. The impact of different tillage
386 treatments might prove less selective for soil microbial community members than other environmental factors
387 such as the regional climate. This theory is in agreement with a recent meta-analysis of reduced tillage
388 systems in organic farming (Cooper, personal communication).

389 In contrast to the similarity between metagenomes, microbial biomass, fungal and bacterial signature
390 gene copies and cellulase enzyme activity, which were determined per gram soil, were clearly lower in CT
391 than RT topsoil. In agreement with these observations, a higher fungal biomass in reduced compared to
392 conventionally tilled soil has been previously reported (Murugan et al. 2013; van Groenigen et al. 2010). Due
393 to a difference in microbial biomass, a higher bacterial 16S- and fungal ITS-abundance and enzymatic
394 activity in RT compared to CT soils might be an indirect result of tillage. Alternatively, the higher enzymatic
395 activity might be due to a higher abundance of cellulase enzymes in RT soil.

396 Several other explanations for the discrepancy between the enzyme activity measurements and the
397 metagenomic analysis exist. One potential explanation is that the soil microbiome might primarily react on
398 differences in tillage treatments on a transcriptional level. The metagenomic approach does not reveal the
399 active members of a given microbial community and it is likely that these will be different under
400 conventional and reduced tillage practice. A metatranscriptomic or metaproteomic approach would be
401 required to reveal the effect of tillage on the expression or translation of specific cellulase genes. The
402 advantage of these approaches is that the active organisms are addressed directly, but the disadvantage is that
403 such measurements can fluctuate considerably over time. Furthermore, it is known that sequencing depth
404 influences the detection of rare or abundant species (Wooley et al. 2010), so it is conceivable that in this
405 study many rare microbial species have been missed that contribute to cellulose degradation. However, if
406 these species have a very low abundance, it is unlikely that their role in degradation is ecologically
407 significant.

408 Another explanation for the observed discrepancy is the sampling time. The incorporation of plant
409 litter leads to a high input of carbon into the soil. It is conceivable that at the time of sampling after one
410 month, the whole microbial community might be profiting from an increased labile carbon pool and there
411 would be little competition between microorganisms. This is in agreement with findings in a field study by
412 Fu *et al.*, which showed that soil respiration was still increasing after 30 days of crop residue incorporation,
413 both under no-tillage and conventional tillage (Fu et al. 2000). Sampling at a later time point (e.g. half a year)

414 might reveal stronger differences between degradation potential of microbial communities of CT and RT
415 soils due to the lower amount of carbon left in the topsoil of CT compared to RT.

416

417 **Outlook**

418 Cellulose degradation replenishes the labile carbon pool in soil and is an important ecosystem service. In this
419 study we addressed the question how different tillage treatments affect cellulose degrading communities in
420 soil in an agricultural setting. As expected, we identified a high diversity of cellulases in soil and several
421 glycoside hydrolase families with high abundances, but did not find evidence that tillage treatment strongly
422 influences cellulose degrading communities at the level of DNA.

423 We detect only few differences in the microbial communities of conventionally versus reduced
424 tillage-treated soils as well as in the abundance of cellulolytic genes. The differences observed in potential
425 enzymatic activities could not be explained by changes in diversity and might be driven by the increased
426 microbial biomass present at RT sites or different expression patterns. However, as the present study is purely
427 based on functional predictions, a confirmation of the results by more targeted methods that confirm the
428 cellulolytic activities are needed, e.g. by using isotopic labelled cellulose as substrate and subsequent stable
429 isotope probing analysis, or a transcriptomic approach further combined with degenerate primer-gene
430 capture. Finally, further quantification of genomic or transcribed cellulases in response to different biotic and
431 abiotic factors, as well as studies that improve our understanding on the dynamics of these microbial
432 communities in time and space, will be a crucial step to better understand to what extent the large diversity of
433 cellulolytic enzymes is relevant for efficient cellulose degradation.

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437

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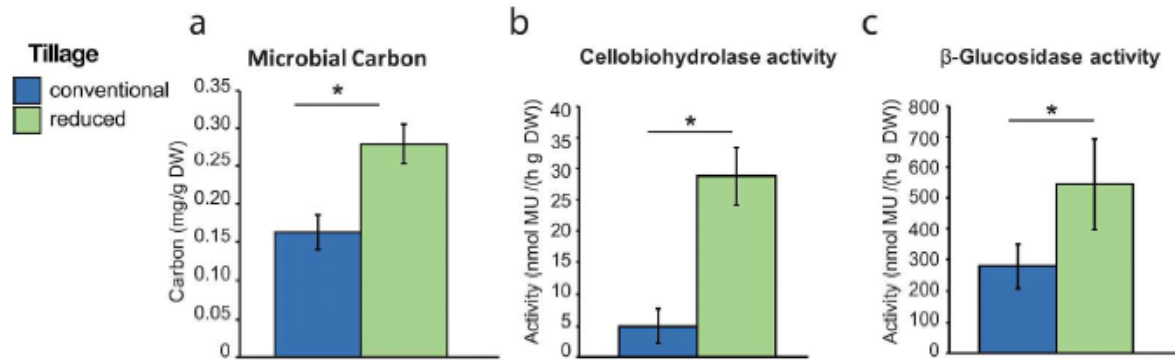
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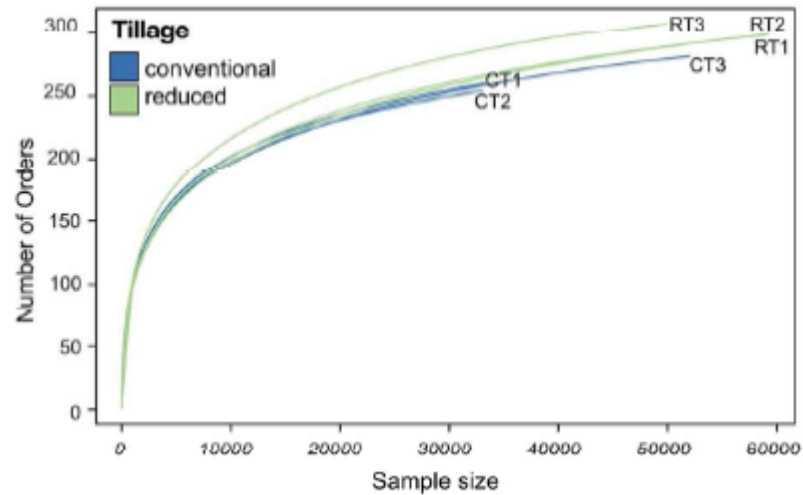
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573
574 **Figure Legends**



575
576 **Figure 1: Reduced tillage increases microbial carbon and potential cellobiohydrolase and β-glucosidase**
577 **activities in the topsoil:** a) Depicted is the amount of microbial carbon detected in soils with conventional
578 and reduced tillage treatments, in mg per gram dry weight soil. b,c) Shown is the amount of produced
579 methylumbelliferone (MU) in nmol per hour and gram dry weight soil, that has been released by the
580 enzymatic degradation of MU-β-cellobioside (b) or MU-β-D-glucoside (c). Significant differences between
581 conventional and reduced tillage were determined by unpaired t-test statistics (* = $P < 0.05$, $n = 3$).



582

583 **Figure 2: Rarefaction curves for the metagenomes of conventional and reduced tillage-treated soil:**

584 Depicted is the number of detected orders (Blastx (Diamond) of all filtered metagenome reads against the
 585 NCBI non-redundant protein database, e-value cut-off = 10^{-4} , assignment of taxonomic rank with MEGAN5)
 586 as a function of the sequencing depth, i.e. a randomly chosen amount of reads from the metagenomes for each
 587 replicate.

588

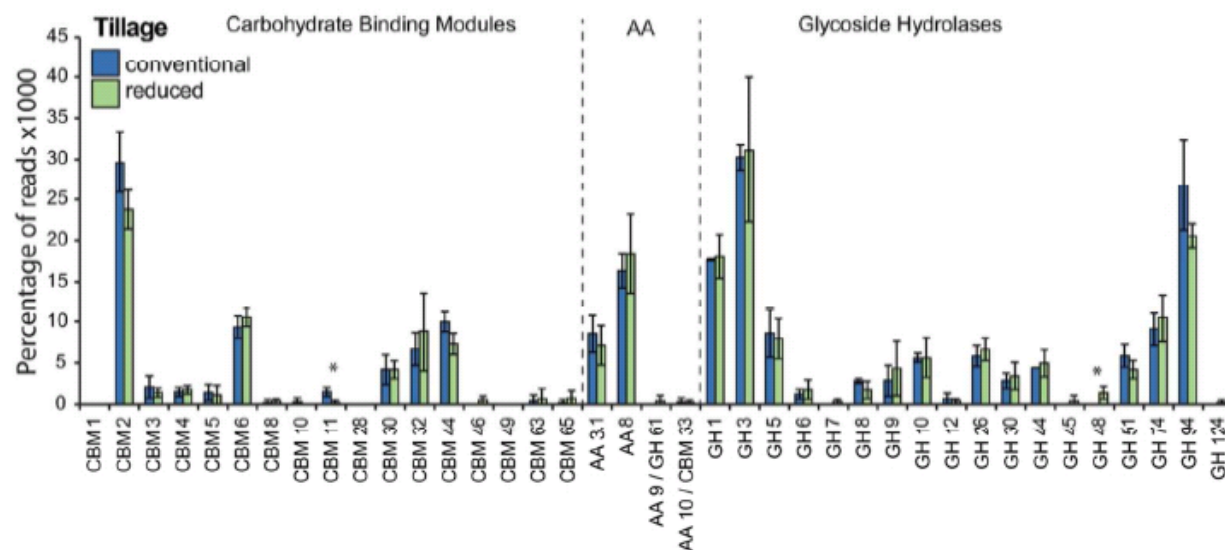


Figure 3: Prediction of cellulases and carbohydrate binding modules from metagenomes of conventional and reduced tillage-treated soil: Depicted is the percentage of reads multiplied by 1000 annotated to catalytic and carbohydrate binding modules found in the metagenomes of conventional and reduced tillage-treated soil using HMM modules and after an additional quality control using a curated database. Significant differences between conventional and reduced tillage were determined by unpaired t-test statistics (* <0.05, n=3).

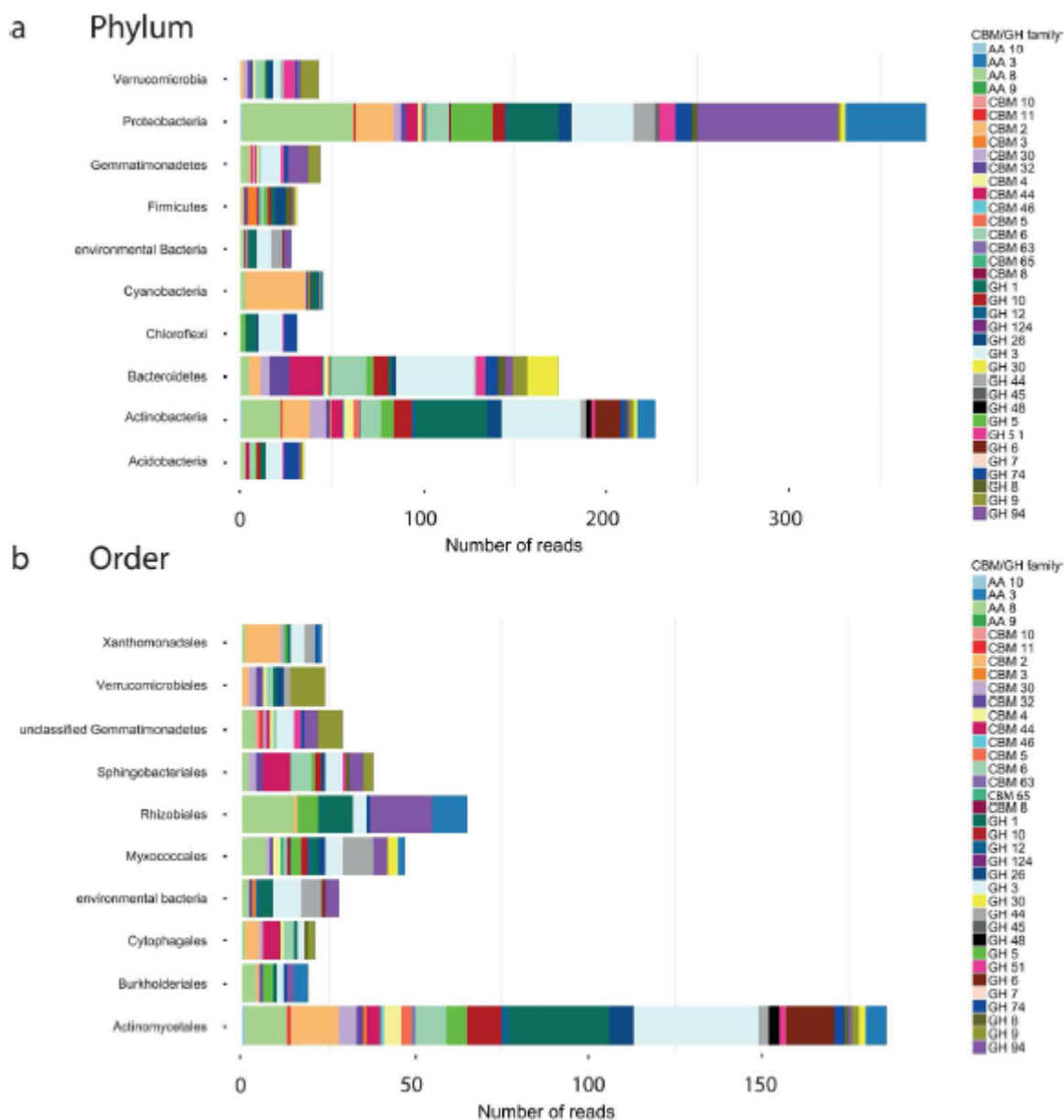


Figure 4:Taxonomic analysis of predicted catalytic and carbohydrate binding modules: Shown is the number of reads annotated to different phyla **a)** or orders **b)** (Blastx (Diamond) against NCBI non-redundant protein database, assignment of taxonomic ranks using MEGAN5) for each catalytic or carbohydrate binding module. Presented are the most abundant taxa.

601

602 **Supplementary Figure Captions**

603 **Figure S1: Taxonomic analysis of metagenomes of conventional and reduced tillage-treated soil:**

604 Depicted are the percentage of reads of the thirty most abundant phyla (a) and orders (b), according to the
605 NCBI non-redundant protein database (n=3). Taxonomic ranks were assigned using MEGAN5.

606

607 **Figure S2: Principal component analysis of metagenomes of conventional and reduced tillage- treated**

608 **soil:** Depicted are the first two components of the principle component analysis based on the relative
609 abundance of orders identified by Blastx (Diamond) against the NCBI non-redundant protein database and
610 assignment of taxonomic ranks using MEGAN5. There is no clear clustering of the metagenomes indicating
611 no strong differences between the metagenomes at this global level.

612

613 **Figure S3: Functional analysis of metagenomes of conventional and reduced tillage-treated soil:** Shown

614 is the percentage of reads annotated (Blastx (Diamond) against the KEGG database, e-value cut-off= 10^{-5} and
615 assignment of functionality using MEGAN5) to different KEGG-pathways, from metagenomes of
616 conventional and reduced tillage-treated soil (n=3). Presented are the thirty most abundant pathways.

617

618 **Figure S4: Taxonomic analysis of most abundant catalytic and carbohydrate binding modules:** For the

619 six most abundant HMM families from Figure 3, a detailed taxonomic annotation with the most abundant
620 phyla is shown. Significant differences between conventional and reduced tillage were determined by
621 unpaired t-test statistics.

622