Exposure to ambient air pollutionand blood lipids in adults: the 33 Communities Chinese Health Study

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53 ABSTRACT

54 Background:Little information exists on the lipidemic effects of air pollution, particularly in 55 developing countries. We aimed toinvestigate the associations oflong-term exposure to 56 ambient air pollutants with lipid levels and dyslipidemias in China.

Methods: In 2009, a total of 15,477 participants aged 18-74 years were recruited from the 33 57 Communities Chinese Health Study conducted in three Northeastern China cities. Total 58 cholesterol (TC), triglycerides (TG), high-density lipoprotein cholesterol (HDL-C), and 59 low-density lipoprotein cholesterol (LDL-C) were measured in participants' blood 60 specimens. Three year (2006-08) average air pollution concentrations were assessed using data 61 from 33 communities (particles with diameters $\leq 1.0 \ \mu m \ (PM_1)$ and $\leq 2.5 \ \mu m \ (PM_{2.5})$ predicted 62 using a spatial statistical model) or 11 air monitoring stations (particles with diameters≤ 10 63 μ m (PM₁₀), sulfur dioxide (SO₂), nitrogen dioxide (NO₂), and ozone (O₃)). Associations were 64 evaluated by two-level logistic and generalized linearregression models. 65

Results: We detected many significant associations between exposure to air pollutants 66 (especially for PM1 and PM2.5) and blood lipid levels. Most of the associations suggested 67 deleterious effects on blood lipid markers (e.g., a 10 μ g/m³ increase in PM₁ was associated 68 with 1.6% (95% confidence interval (CI): 1.1, 2.0), 2.9% (95% CI: -3.3, 9.3), and 3.2% (95% 69 CI: 2.6, 3.9) higher levels of TC, TG, and LDL-C, respectively, but 1.4% (95% CI: -1.8, -0.9) 70 lower HDL-C levels), although beneficial associations were found for O₃. In analysis with 71 dyslipidemias, all the observed associations suggested deleterious lipidemic effects of air 72 pollutants, and no significantbeneficial association was observed for O₃. Stratified analyses 73 showed that the associations were stronger in overweight or obese participants; sex and 74

- agemodified the associations, but the pattern of effects was mixed.
- 76 *Conclusions*:Long-term ambient air pollutionwas associated with both altered lipid profiles
- and dyslipidemias, especially among overweight or obese participants.
- 78 Key words:Particulate matter,Gaseous pollutants,Dyslipidemia, Lipids, Cross-sectional study

79 Abbreviations

80 BMI, body mass index; CI, confidence interval; CVD, cardiovascular diseases; HDL-C, 81 high-density lipoprotein cholesterol; LDL-C, low-density lipoprotein cholesterol; NO₂, 82 nitrogen dioxide; OR, odds ratio; O₃, ozone; PM₁, particles with diameters $\leq 1.0 \mu$ m; PM_{2.5}, 83 particles with diameters $\leq 2.5 \mu$ m; PM₁₀, particles with diameters $\leq 10 \mu$ m; PM_{2.5-10}, particles 84 with diametersranging from 2.5 to 10 μ m; SEPA, the State Environmental Protection 85 Administration of China; SO₂, sulfur dioxide; TC, total cholesterol; TG, triglycerides; 86 33CCHS, the 33 Chinese Community Health Study. 87

88 **1. Introduction**

Cardiovascular diseases(CVD)are responsible for approximately 31% of deaths worldwide 89 (World Health Organization, 2017) and the results of numerous epidemiological studies have 90 supported a causal relation for long-term air pollution exposure with CVD (Brooket al., 2010; 91 Bourdrel et al., 2017). Inhaled air pollutants triggerinflammation, oxidative stress, autonomic 92 imbalance, and epigenetic changes (Brook et al., 2010; Bourdrel et al., 2017). These reactions 93 have been linked to several CVD risk factors, includingatherosclerosis, hypertension, diabetes 94 mellitus, and dyslipidemia(Brook et al., 2010; Thiering and Heinrich, 2015; Rajagopalan and 95 Brook, 2012; Yang et al., 2018). For example, higher levels of particulate matter (PM) have 96 97 been linked to increased systemic inflammation (Brook et al. 2010), which can lead to adverse lipid metabolism and lipid oxidation (Chen et al., 2013). The relationshipsofair pollution with 98 hypertension and diabetes mellitus have been demonstrated bymany epidemiological and 99 100 experimental studies(Brook et al., 2010; Rajagopalan and Brook, 2012; Thiering and Heinrich, 2015), including our own(Dong et al., 2013; Yang et al., 2017; Yang et al., 2018). 101

Dyslipidemia, 102 characterized by hypercholesterolemia, hypertriglyceridemia, hypoalphalipoproteinemia, and/orhyperbetalipoproteinemia, is the foremost cause of 103 atherosclerosis(Stensland-Bugge et al., 2000) and is inextricably related to the development of 104 CVD(Expert Panel on Detection, Evaluation, and Treatment of High Blood Cholesterol in 105 Adults,2001). The global prevalence of dyslipidemias is high and increasing (Cahalin et al., 106 2014). For example, a 2013 report from the American Heart Association suggested that the 107 prevalences hypercholesterolemia, hypoalphalipoproteinemia, 108 of and hyperbetalipoproteinemiain American adults were 43.4%, 21.8%, and 31.1%, respectively 109

(Go et al., 2013). Several previous epidemiological studies explored relationships between 110 ambient air pollutant exposure and dyslipidemia and blood lipid levels, yet the results were 111 inconsistent (Bell et al., 2017; Bind et al., 2016; Cai et al., 2017; Chuang et al., 2011; Jiang et 112 113 al., 2016; Poursafa et al., 2014; Shanley et al., 2016; Sørensen et al., 2015; Wallwork et al., 2017; Yeatts et al., 2007; Yitshak Sade et al., 2016) (see Table S1). Additionally, moststudies 114 investigated effects among specific populations, including asthmatics(Yeatts et al., 2007), 115 patients with chronic diseases (Yitshak Sade et al., 2016), the elderly (Bind et al., 2016; 116 Chuang et al., 2011; Sørensen et al., 2015; Wallwork et al., 2017), and adolescents (Poursafa 117 et al., 2014). However, the lipidemiceffects of air pollution exposurewere rarely evaluated in 118 119 general populations. Moreover, previous studies were mostly conducted in high-income 120 nations or regions(Bell et al., 2017; Bind et al., 2016; Chuang et al., 2011; Shanley et al., 2016; Sørensen et al., 2015; Wallwork et al., 2017; Yeatts et al., 2007; Yitshak Sade et al., 121 2016). There are few data available to characterize the risks of air pollution exposure on lipid 122 levels and dyslipidemia in low-income countries. 123

124 In recent decades, China has experienced a gradual increase in the prevalence of dyslipidemia, although itremains lower than that in many developed countries(Pan et al., 2016). For instance, 125 the 2002 China National Nutrition and Health Survey reported that the prevalences of 126 hypercholesterolemia, hyperbetalipoproteinemia, hypoalphalipoproteinemia, 127 and hypertriglyceridemiawere 2.9%, 2.5%, 7.4%, and 11.9%, respectively (Zhao, 2008), while the 128 corresponding 2013-2014 China Chronic Disease and Risk Factor Surveillance 129 prevalenceswere 6.9%, 8.1%, 20.4%, and 13.8% (Zhang et al. 2018). Simultaneously, air 130 pollution has emerged as a severe environmental problem in China(Guan et al., 2016; Rohde 131

and Muller, 2015). Given temporal increases inboth ambient air pollution and the prevalence
of dyslipidemia, and the scarcity of data, it is of significant public health importance to
explore the relationship between the two. To begin to address the data gap, this study
examinedassociations betweenlong-term residential ambient air pollutionandblood lipid levels
in a large community-based sample of urban adultsparticipating in the33 Chinese Community
Health Study (33CCHS).

138 **2. Methods**

139 **2.1. Study population**

The population of the 33CCHSwas previously described in detail (Dong et al., 2013; Yang et 140 al., 2017). Briefly, in2009, we used a random-number generator coupled to a four-staged, 141 stratified, cluster sampling strategy to recruit study participants. First, to maximize the 142 inter-city gradients of air pollutants, we randomly selected three cities - Shenyang, Anshan, 143 and Jinzhou - from 14 total cities in Liaoning province. There are five districts in Shenyang 144 city and three each inthe cities of Anshan and Jinzhou. Second, we randomly selected three 145 communities from each of the districts, generating a total of 33 study communities. Each 146 study community was approximately 0.25-0.64 km² in area. Third, we randomly selected 147 700-1000 households from each study community. Fourth, from each study household, we 148 randomly selected one adult aged 18 to 74 years for study enrollment. To be included, 149 individuals had to liveat the study address for at least five years, have no severepre-existing 150 diseases (e.g., cancers), and not be pregnant. Based on the sampling frame, 28,830 participants 151 were invited, of whom 24,845 individuals completed the survey, yielding an overall response 152 rate of 86.2%. A total of 9368 individuals were excluded from the present analysis due to 153

refusal to provide a blood sample, leaving a final sample of 15,477 participants (62.3% of the 33CCHS participants).All participants completed informed consent prior to study enrollment, and Sun Yat-Sen University's Human Studies Committee reviewed and approved all study procedures and protocols.

158 **2.2. Health outcomes**

After an overnight fast, peripheral venous blood sampleswere collected from study 159 participants. Total cholesterol (TC), triglycerides (TG), high-density lipoprotein cholesterol 160 (HDL-C), and low-density lipoprotein cholesterol (LDL-C)levels were determinedusing a 161 Hitachi Autoanalyzer (Type 7170A; Hitachi Ltd.; Tokyo, Japan). Hypercholesterolemia was 162 defined as TC \geq 240 mg/dL; hypertriglyceridemia was defined as TG \geq 200 mg/dL; 163 hypoalphalipoproteinemia was defined as HDL-C $\leq 40 \text{ mg/dL}$; and hyperbetalipoproteinemia 164 was defined as LDL-C \geq 160 mg/dL(Joint Committee for Developing Chinese Guidelines on 165 Prevention and Treatment of Dyslipidemia in Adults, 2007). 166

167 **2.3. Air pollution data**

A detailed description of the exposure assessment was provided in our previous publications(Dong et al., 2013; Yang et al., 2017; Chen et al., 2018). Briefly, daily concentrations of PM with diameters $\leq 1.0 \ \mu m \ (PM_1)$ and $\leq 2.5 \ \mu m \ (PM_{2.5})$ were predicted for the 33 study communities, at a $0.1 \ \infty \ 0.1 \ \circ$ spatial resolution, using PM₁ and PM_{2.5} measurements from air monitoring stations, satellite remote sensing, meteorology, and land use characteristics. Aerosol optical depth data was combined from two types of Moderate Resolution Imaging Spectroradiometer algorithms—Dark Target and Deep Blue. A

generalized additive model was developed to link ground-monitored PM₁ and PM_{2.5} data with 175 aerosol optical depth data and other spatial and temporal predictors. In each of the study 176 districts, there was one air monitoring station, which was located within a 1-kmdistance from 177 the centroid of the community of each study participant's home address (Fig. S1). We 178 collected data for PM with diameters $\leq 10 \ \mu m$ (PM₁₀), sulfur dioxide (SO₂), nitrogen dioxide 179 (NO₂) and ozone (O₃) concentrations from 11 air monitoring stations, according 180 tostandardized procedures set by the State Environmental Protection Administration of China 181 (SEPA)(SEPA, 1992). These air monitoring stations were mandated to be away from main 182 traffic roads, industry sources, or residential sources of emissions from the combustion of coal, 183 184 waste, or oil. Thus, air pollution measurements from these stations were more likely to reflect mixtures from different sources and background levels in urban areas. Daily average 185 concentrations of PM₁₀, SO₂, NO₂, and an eight-hour average of O₃ were calculated using 186 measurements from days with at least 75.0% of one-hour values available. PM_{coarse} (PM_{2.5-10}) 187 was calculated by subtracting $PM_{2.5}$ from PM_{10} . The three-year (2006-08) average 188 concentrations of air pollutants were calculated for the 33 communities and 11 districts (or air 189 monitoring stations), and then assigned to each participant based on proximity of the 190 community or district to his/her residential address, as surrogates of long-term air pollution 191 exposure. 192

193 **2.4. Covariates**

All potential covariates were selected *a priori*. An interviewer administered questionnaire was
used to collect information on age (years), sex (male/female), nationality (Han/others),
household annual income (≤5000 Yuan/5001-10000 Yuan/10001-30000 Yuan/≥30000 Yuan),

highest educational attainment (no school/primary school/middle school/junior college or 197 higher), current smoking (yes/no), alcohol drinking (yes/no), regular exercise (yes/no), 198 controlled diet with low calories and low fat (yes/no), sugar-sweetened soft drink 199 200 consumption (≤ 1 day per week/ 2-4 days per week/ ≥ 5 days per week), family history of dyslipidemia, and district (or community). Height (cm) and weight (kg) were measured 201 without heavy clothes and shoes, and body mass index (BMI, kg/m²) was calculated.Data on 202 203 temperature, humidity, and wind speed for the 11 districts were obtained from Liaoning Provincial Meteorological Bureau. Per-capita gross domestic product and population density 204 in each district were obtained from Shenyang, Jinzhou, and Anshan cities' Statistical 205 206 Yearbooks. Season at the time of blood sampling was also included as a potential cofounder.

207 **2.5. Statistical analysis**

The Shapiro-Wilk and the Bartlett tests were used to examine data normality and homogeneity, respectively. Differences in basic characteristics between men and women, and between participants with and without blood sampling, were tested using Student's t-test, or the Wilcoxon rank sum or chi-square tests. We used the Spearman rank correlation test to assess the relationship between air pollutants.

We applied linear regression models to assess associations between individual air pollutants (per 10 μ g/m³ increase)andblood lipid levels (TC, TG, HDL-C, and LDL-C), which were naturally log-transformed to achieve normal distributions. Effect estimates were then back-transformed from the log scale using 100 × [exp (β) - 1] and presented as percent differences with corresponding 95% confidence intervals (CI). We also used two-level binary

logistic regression models to investigate associations of ambient air pollutants (per $10 \mu g/m^3$ 218 increase) with hypercholesterolemia, hypertriglyceridemia, hypoalphalipoproteinemia and 219 hyperbetalipoproteinemia, whereparticipantswere treated as first-level units and districts or 220 221 communities as second-level units. Detailed descriptions of the two-stage binary logistic regression models were provided in our previous publications(Dong et al., 2013; Yang et al., 222 2017) and in the supplemental material (detailed information on two-level binary logistic 223 regression model). These results are presented as odds ratio (OR) with corresponding 95% CI. 224 All regression models were adjusted for the variables listed in the Covariates section. District 225 or community was incorporated as a random effect, and the remaining covariates were 226 227 incorporated as fixed effects. Multi-pollutant models could not be applied, as all air pollutants 228 were correlated moderately to highly (except NO₂ and SO₂). Therefore, highly correlated pollutants were regressed against each other and the residualswere then incorporated into the 229 230 models for associations between air pollutants and lipid outcomes (Flexeder et al., 2017).

We performed subgroup analyses by sex (men, women), age group (≥ 50 years, < 50 years), 231 and BMI category ($\geq 25 \text{ kg/m}^2$, $<25 \text{ kg/m}^2$), and across-product term was incorporated into 232 regression models to evaluate the statistical significance of their interactions. Additionally, we 233 repeated theregression analyses of air pollution and lipid levels by excluding participants who 234 took lipid lowering drugs, or those had cardiovascular diseases or diabetes mellitus. Also, we 235 applied multi-annual average concentrations of air pollutants (i.e., one-year average 236 (2008), two-year average (2007-08), and three-year average (2006-08)) as long-term exposures, 237 in order to exclude possible exposure fluctuations over shorter periods. Furthermore, 30-day 238 average air pollutant concentrations before the blood draw were additionally adjusted to 239

explore the potential impact of short-term air pollution exposure on lipids. Similarly, for pollutants in which long-term and short-term levels were highly correlated, these were regressed against each other and the individual residuals were then incorporated into the regression models.

Data analysis was performed using SAS 9.4(SAS Institute, Cary, NC) with a p value less than
0.05 considered as statistically significant for a two-tailed test.

246 **3. Results**

247 3.1. Descriptive statistics

The mean age of study participants was 45.0 years, and 52.7% were men (Table 1). Most 248 participants had a middle school or higher education (84.8%). Thirty percent and 24.6% of 249 them were smokers and drinkers, respectively. Approximately 31.9% reported engaging in 250 regular exercise and 7.9% had a family history of dyslipidemia. The prevalences of 251 hypertriglyceridemia, hypoalphalipoproteinemia, and hypercholesterolemia, 252 hyperbetalipoproteinemiawere 11.1%, 22.6%, 18.3%, and 8.6%, respectively. The median 253 concentrations of TC, TG, HDL-C, and LDL-C were 179.92 mg/dL, 118.58 mg/dL, 50.97 254 mg/dL, and 98.60 mg/dL, respectively. Men and women differed for all sociodemographic and 255 lifestyle variables, with the exception of nationality. The distribution of the main 256 257 characteristics was similar between the analytical sample and those who were excluded from this analysis (Table S2) 258

The PM₁, PM_{2.5}, PM₁₀, SO₂, NO₂, and O₃ concentrations varied greatly across study districts(or communities) with a median of $62 \,\mu\text{g/m}^3$, $73 \,\mu\text{g/m}^3$, $123 \,\mu\text{g/m}^3$, $48 \,\mu\text{g/m}^3$, $33 \,\mu\text{g/m}^3$, and $50\,\mu\text{g/m}^3$, respectively (Table 2).Except for SO₂ and NO₂,moderate to high correlationsbetweenairpollutantswere detected (Spearman correlations ranged from 0.45 to 0.99). In particular, NO₂ and O₃ were positively correlated, which might be caused by high air pollutants levels, especially NO_x and volatile organic compounds(Shi et al., 2015; Zong et al. 2017), at the studysite.

266 3.2. Associations between air pollutants and lipid levels

The associations of air pollutants with blood lipid levels are summarized in Tables 3 and S3. 267 For all participants, higher concentrations of PM1 and PM2.5 consistently showed significant 268 associations with higher levels of TC, TG, and LDL-C, as well as with lower levels of HDL-C. 269 However, associations of PM₁₀ with blood lipids were less consistent, and PM_{2.5-10} was 270 associated only with TG levels (Table S3). Higher NO₂ levels were significantly associated 271 with higher levels of TC and TG and with lower levels of HDL-C. SO₂ was positively 272 associated with TG, but not with the remaining lipid markers. Higher O₃ concentrations were 273 significantly associated with higher levels of TG and HDL-C, but with lower levels of TC and 274 LDL-C.The results remained materially unchanged in sensitivity analyses where 263 275 individuals taking lipid lowering drugs (Table S4) or2222 individuals with CVD or diabetes 276 mellitus (Table S5) were excluded, where multi-annual average concentrations of air pollutants 277 were used (Table S6), and where the models were additionally adjusted for short-term air 278 pollutant levels (Table S7). The only exception in the latter case was the statistical 279 insignificance of the association between O₃ and HDL-C (Table S7). 280

281 We detected statistically significant interactions between air pollutant concentrations and sex

on blood lipids; however, the pattern of effects was mixed in stratified analyses (Fig. 1; Table 282 S8).For example, while associations of all six air pollutants with HDL-C were stronger for 283 men, in the case of TG they werestronger for women. In addition, thepositive associations of 284 LDL-C with PM₁ and PM_{2.5} were stronger in women than in men, and were stronger yet 285 negative for PM₁₀, SO₂, NO₂ and O₃. In stratified analysis by age, the associations between air 286 pollutants and lipid levels were similarly complex, although most interaction terms were not 287 statistically significant (Table S8). In another stratified analysis, BMI significantly modified 288 the associations of PM₁ and PM_{2.5} with HDL-C and LDL-C, with stronger associations among 289 overweight/obese participants (Table S8). 290

291 3.3. Associations between air pollutants and dyslipidemias

For all participants, we detected statistically significant associations between: (1)higher PM₁, 292 PM_{2.5}concentrations with higher odds 293 and for hypercholesterolemia, hypoalphalipoproteinemia, and hyperbetalipoproteinemia; (2)higher 294 PM₁₀, SO₂, and O₃concentrations with higher odds for hypertriglyceridemia; and (3) higher 295 NO₂concentrations with higher odds for hypercholesterolemia (Table 4). In stratified analyses 296 by sex and age, we detected statistically significant interactions of air pollutant concentrations 297 with sex and age for severaldyslipidemia associations, but the pattern was mixed (Table S9).In 298 stratified analyses by BMI, associations of air pollutants with dyslipidemias (particularly for 299 hypoalphalipoproteinemia) were consistently greater in participants who were overweight or 300 obese (Table S9). 301

302 **4. Discussion**

303 4.1. Key findings

To our knowledge, this is the largest population-based epidemiological study to date, to 304 explore associations between ambient air pollution and blood lipids in a developing country. We 305 306 detected many statistically significant associations between exposure to long-term ambient air pollutants (particularly PM1 and PM2.5) and blood lipid levels and the prevalence of 307 dyslipidemias. Most of the associations suggested deleterious effects on blood lipid levels (i.e., 308 associated with higher levels of TC, TG, and LDL-C, and lower HDL-C levels), though 309 beneficial associations were detected for O₃.However, all observed associations suggested 310 deleterious effects of air pollutants on dyslipidemias; nobeneficial significant association was 311 312 observed for O₃ with dyslipidemia. Stratified analyses showed that associations between air 313 pollutants and lipids were generally stronger among overweight and obese participants; sex and agealso modified associations, but the pattern of effects was complicated. 314

4.2. Comparison with other studies and interpretations

Several previous studies investigated associationsbetween air pollution exposure and blood 316 lipid levels or dyslipidemias, but the findings have been inconsistent(Bell et al., 2017; Bind et 317 al., 2016; Cai et al., 2017; Chuang et al., 2011; Jiang et al., 2016; Poursafa et al., 2014; 318 Shanley et al., 2016; Sørensen et al., 2015; Wallwork et al., 2017; Yeatts et al., 2007; Yitshak 319 Sade et al., 2016). Consistent with our current findings, a cross-sectional study from the USA 320 revealed that an 11.1 $\mu g/m^3$ increase in PM₁₀ was associated with 2.42% higher TG 321 levels(Shanleyet al., 2016). Another cross-sectional study, from Denmark, reported positive 322 association for PM2.5 exposure with TC levels. Similarly, a retrospective cohort study of Israeli 323 adults reported a statistically significant association between higher PM2.5 concentrations and 324

lower HDL-C levels(Yitshak Sade et al., 2016).Chuang et al.(2011) found thatPM₁₀ and 325 NO₂were significantly associated with higher TC levels in Taiwan, but there were no 326 associations with TGor HDL-C levels. In addition, a panel study among American adults 327 328 showed no significant association between PM₂ sexposure and TC levels (Yeatts et al., 2007). The specific reasons for the inconsistent results across studies of ambient air pollution 329 exposure and blood lipid levels are not clear. They may be related to the differences in 330 population characteristics (e.g., age, genetic background, lifestyles, and health status) and/or 331 local or regional differences in the physical and chemical properties of the air pollutants (e.g., 332 concentrations, chemical constituents, and sources). 333

The biological mechanisms underlying links between air pollutants and lipid metabolism are 334 not fully understood. However, several possible biological pathways have been proposed. One 335 hypothesis is that inhaled air pollution elicits systemic inflammation and oxidativestress 336 (Lodovici and Bigagli, 2011; Shanley et al., 2016; Thompson et al., 2010), which can induce 337 adverse lipid metabolism and lipid oxidation (Chen et al., 2013). Air pollutants could also 338 cause aberrant DNA methylation by decreasing activity of DNA methyltransferases. Several 339 studies have linked air pollution exposure to abnormalities in global DNA methylation as well 340 asto methylation at specific genes related to lipid metabolism (Chen et al., 2016; Bind et al., 341 2014). The associations between air pollutant concentrations, PM in particular, and blood lipid 342 levels in our current study are consistent with these hypothesized biologicalmechanisms. 343

In stratified analyses, we found stronger associations between air pollutants and blood lipids
among overweight and obese participants, which are partially consistent with S ørensen et al's

work. (2015). Existing evidence shows that both air pollution exposure 346 and overweight/obesity are associated with higher systemic inflammation (Bastard et al., 2006; 347 Rajagopalan and Brook, 2012). Overweight and obese participants might therefore be more 348 vulnerable to adverse health effects from air pollution, which act in part through an 349 inflammation pathway. We also found that sex and age modified the effects of air pollution on 350 lipid levels and dyslipidemias, but the pattern was mixed. To thebest of our knowledge, only 351 two prior studies (Shanley et al., 2016; Sørensen et al., 2015)investigated sex- and 352 age-specific associations between air pollutants and blood lipids. In line with our linear 353 regression findings, Shanley et al. (2016) reported that age did not significantly modify the 354 355 associations of PM₁₀ with TC and TG. However, they also observed a stronger association between PM₁₀ and TC among women, whereas for PM₁₀ and TG a stronger association was 356 observedamong men, which contradicts our results. In another study, Sørensen et al. (2015) 357 observed that both age and sex did not modify the associations of NO_2 and $PM_{2.5}$ with 358 TC.Collectively, there is limited and inconsistent epidemiological evidenceto characterize the 359 role of age and sex in modifying air pollution-lipid associations at present, and so further 360 361 investigation is merited.

362 4.3. Implications for policy makers

Dyslipidemias are well-documented risk factors for CVD (Zhang et al., 2003).Randomized trials have shown that lipid-lowering treatment could significantly decrease the risk of CVD (Fulcher et al., 2015). For example, the Asia Pacific Cohort Studies Collaboration reported 35% and 25% increased risksfor coronary death and incident stroke in Asians, respectively, per 1-mmol/L higher serum TC (Zhang et al. 2003).Yet, a meta-analysis of 22 trials found that a 368 1-mmol/L LDL-C reduction could decrease major CVD events by 21% (Fulcher et al., 2015). 369 In the current study, we found that a $10-\mu g/m^3$ increase in air pollutants was associated with an 370 approximately 1-2% increase in blood lipids levels. Although the observed difference was 371 relatively small and thus of uncertain clinical impact, our findings have certain public health 372 implications for helpingpolicy makers to develop intervention policies, given the high levels 373 of air pollution (Guan et al., 2016) and high prevalence of CVD in China (Peters et al. 2017).

374 4.4. Strengths and limitations

This study has several strengths. First, our analysis was based on a large sample of 375 northeastern Chinese with a high response rate, using standardized protocols and instruments, 376 which ensured sufficient statistical powerto detect modest effects, and generalizability of our 377 results.Second, unlike most previous studies, which focused on specific populations, our 378 present study provides valuable evidence regarding a general population in a developing 379 nation. Third, in addition to measuring exposure to the traditional ambient 'criteria'air 380 pollutants (i.e., PM_{2.5}, PM₁₀, SO₂, NO₂, and O₃), we, for the first time, report effects for PM₁ 381 on blood lipids. In addition, all air pollutants levels are high in our study settings, thus our 382 results can provide a valuable reference for other developing countries, such as India.Finally, 383 a combination of objectively measured lipid levelscoupled to a rich set of covariate data 384 allowed for a comprehensive data analysis, including adjustment for BMI, physical activity, 385 diet, and meteorological variables to minimize the impact of confounding. 386

387 Despite the novel nature of our results, several limitations should also be acknowledged. First,
388 the cross-sectional study design precludedassessment of temporality, and we are thus unable

infera causal association between air pollution exposure and blood lipid 389 to levels.Second, exposure levels were assigned using data from the nearest air monitor or 390 community rather than using personal air pollution exposure data, which means that only 33 391 (for PM₁ and PM_{2.5}) and 11 (for PM₁₀, NO₂, SO₂, and O₃) unique air pollution values were 392 available for the 15,477 participants. The values may have misclassified some participants, by 393 randomly underestimating exposure in some and overestimating 394 exposure in others.Nevertheless, such exposure misclassification is likely to bias the results towards null 395 (Hutcheon et al., 2010). This indicates that if we had individual-level data on air pollution 396 levels, our estimated effects of air pollution on blood lipids would have been stronger than the 397 398 current estimates. Furthermore, our exposure assessment did not capture specific emissions 399 known to adversely affect health, such as traffic-related sources, that are likely to show large variation across both space and time. Third, participants'baseline characteristics differed 400 significantly among thestudy districts and communities (Yang et al., 2017). Although we 401 collected rich covariate data to adjust for confounding, it is possible that the observed 402 significant associations were biased by unmeasured confounding factors that differed across 403 404 the study districts or communities, including health-care access, available green-space, noise, and household environments. Unfortunately, these data were not collected by the 33CCHS.In 405 addition, conditions such as acute infection and inflammation, hyperthyroidism, and nephrotic 406 407 syndrome, may affect lipid status (Nigam, 2011). These data were also not available in our current study, which may have compromised our estimates. Fourth, 62.3% of the 33CCHS 408 participants with blood specimens were included in the current analysis, and so a selection 409 bias was not impossible. However, the distribution of baseline characteristics was similar 410

between participants with and without a blood specimen, and so any effect is likely to be modest. Fifth, we used a questionnaire to collect self-reported information on demographic and lifestyle characteristics; thus,recall bias and misclassification is possible. Finally, correlations between air pollutants were generally moderate or high, which limited our ability to assess the health effects of multiple pollutantssimultaneously.However, we performed regression analyses on air pollutants that highly correlated with each other, and then adjusted theindividual residuals in order to accommodate the co-exposures.

418 **5.** Conclusions

419 Our findings suggest that long-term exposure to ambient air pollution is associated with 420 altered lipid levels and the prevalence of dyslipidemias, especially among overweight and 421 obese people.However, considering the limitations of our study, future well-designed 422 longitudinal studies are warranted to more definitively evaluate the effects of ambient air 423 pollution on lipid metabolism.

424 **Declaration of interests**

425 None

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569 **Figure legend**

- **Fig. 1.** Associations between air pollutants and blood lipid levels (A: total cholesterol, TC; B:
- triglycerides, TG; C: high-density lipoprotein cholesterol, HDL-C; D: low-density lipoprotein
- 572 cholesterol, LDL-C) by sex.Stars represent statistically significant interactions.

	Value (mean ±SD, n(%), or median (Q1, Q3))							
Characteristics	Total (n=15,477)	Men (n=8156)	Women (n=7321)					
Age (years) ^a	44.97 ± 13.45	44.44 ± 14.20	45.56 ± 12.55					
Ethnicity								
Han	14,554 (94.0%)	7670 (94.0%)	6884 (94.0%)					
Other	923 (6.0%)	486 (6.0%)	437 (6.0%)					
Education ^a								
Junior college or higher	3579 (23.1%)	2250 (27.6%)	1329 (18.2%)					
Middle school	9554 (61.7%)	5008 (61.4%)	4546 (62.1%)					
Primary school	1863 (12.0%)	782 (9.6%)	1081 (14.8%)					
No school	481 (3.1%)	116 (1.4%)	365 (5.0%)					
Annual family income ^a								
≤5000 Yuan	1167 (7.5%)	618 (7.6%)	549 (7.5%)					
5001-10,000 Yuan	1977 (12.8%)	846 (10.4%)	1131 (15.5%)					
10,001-30,000 Yuan	7869 (50.8%)	4198 (51.5%)	3671 (50.1%)					
≥30,000 Yuan	4464 (28.8%)	2494 (30.6%)	1970 (26.9%)					
Tobacco smoking status ^a								
Nonsmoker	10,837 (70.0%)	4004 (49.1%)	6833 (93.3%)					
Smoker	4640 (30.0%)	4152 (50.9%)	488 (6.7%)					
Alcohol drinking status ^a								
Nondrinker	11,668 (75.4%)	4562 (55.9%)	7106 (97.1%)					
Drinker	3809 (24.6%)	3594 (44.1%)	215 (2.9%)					
Regular exercise ^a								
Yes	4932 (31.9%)	2724 (33.4%)	2208 (30.2%)					
No	10,545 (68.1%)	5432 (66.6%)	5113 (69.8%)					
Control diet with								
low calorie and fat ^a								
Yes	3861 (24.9%)	1828 (22.4%)	2033 (27.8%)					
No	11,616 (75.1%)	6328 (77.6%)	5288 (72.2%)					
Sugar-sweetened	, , , ,							
soft drink intake ^a								
≤1 day per week	13,621 (88.0%)	6996 (85.8%)	6625 (90.5%)					
2-4 days per week	1286 (8.3%)	818 (10.0%)	468 (6.4%)					
>5 davs per week	570 (3.7%)	342 (4.2%)	228 (3.1%)					
BMI ^a			- (- · · · ·)					
$>25 \text{ kg/m}^2$	6271 (40.5%)	3660 (44.9%)	2611 (35.7%)					
$<25 \text{ kg/m}^2$	9206 (59.5%)	4496 (55.1%)	4710 (64.3%)					
Family history of dyslipidemia ^a								
Yes	1228 (7.9%)	492 (6.0%)	736 (10.1%)					
No	14.249 (92.1%)	7664 (94.0%)	6585 (89.9%)					
Blood lipids& dyslinidemias	, (///)							
TC (mg/dL)	179.92 (155.98, 205.41)	180.70 (157.53, 204.83)	179.15 (154.44, 206.18)					
$TG (mg/dI)^{a}$	118 58 (81 //2 176 00)	130.09 (88.50, 200.80)	106 19 (75 22 160 18)					

Table 1Study population characteristics (n = 15,477).

HDL-C (mg/dL) ^a	50.97 (43.63, 60.61)	48.26 (40.84, 58.30)	54.05 (47.10, 62.93)
LDL-C (mg/dL)	98.60 (75.67, 122.31)	98.60 (75.87, 122.29)	98.74 (75.50, 122.41)
Hypercholesterolemia	1717 (11.1%)	910 (11.2%)	807 (11.0%)
Hypertriglyceridemia ^a	3494 (22.6%)	2336 (28.6%)	1158 (15.8%)
Hypoalphalipoproteinemia ^a	2836 (18.3%)	2064 (25.3%)	772 (10.6%)
Hyperbetalipoproteinemia	1333 (8.6%)	698 (8.6%)	635 (8.7%)

Abbreviations: BMI, body mass index; HDL, high-density lipoprotein cholesterol; LDL, low-density lipoprotein cholesterol; Q1, 25thpercentile; Q3, 75th percentile; SD, standard deviation; TC, total cholesterol; TG, triglycerides.

^a Statistically significant difference between men and women (p<.05).

	Summary statistics						Spearman correlation coefficients						
Exposure	Mean	Median	Minimum	Maximum	IQR	>WHO guideline (%) ^c	PM_1	PM _{2.5}	PM_{10}	SO_2	NO_2	O ₃	
$PM_1 (\mu g/m^3)^a$	65.97	62	50	82	15	none ^d	1.00	0.99 ^e	0.73 ^e	0.52	0.67 ^e	0.47	
$PM_{2.5} (\mu g/m^3)^a$	82.02	73	64	104	26	100		1.00	$0.72^{\rm e}$	0.51	$0.63^{\rm e}$	0.45^{e}	
$PM_{10} (\mu g/m^3)^b$	123.06	123	93	145	19	100			1.00	0.81^{e}	0.65^{e}	0.81^{e}	
$SO_2(\mu g/m^3)^b$	54.42	48	36	78	20	100				1.00	0.25	0.84^{e}	
$NO_2(\mu g/m^3)^b$	35.28	33	27	45	9	18.2					1.00	0.45	
$O_3(\mu g/m^3)^b$	49.40	50	27	71	22	0.0						1.00	

578 **Table 2**Summary statistics and pairwise Spearman correlations of air pollutants.

579 Abbreviations: IQR, interquartile range; NO₂, nitrogen dioxide;O₃, ozone; PM₁, particle with aerodynamic diameter $\leq 1.0 \ \mu m$; PM_{2.5}, particle with aerodynamic diameter $\leq 2.5 \ \mu m$; PM₁₀, particle with aerodynamic diameter $\leq 10 \ \mu m$; SD, standard deviation; SO₂, sulfur dioxide.

^aBased on values from 33 communities.

^bBased on values from 11 districts.

^cWorld Health Organization (WHO) air quality guidelines (2005).

584 ^dNo guideline for PM_1 .

^eStatistically significant correlation (p < .05).

	%changes (95% Confidence Interval) ^a				
Pollutant	TC	TG	HDL-C	LDL-C	
PM ₁	1.6 (1.1, 2.0) ^b	2.9 (-3.3, 9.3)	-1.4 (-1.8, -0.9) ^b	3.2 (2.6, 3.9) ^b	
PM _{2.5}	$1.1 (0.8, 1.4)^{b}$	1.1 (0.4, 1.8) ^b	-1.1 (-1.4, -0.8) ^b	2.9 (2.4, 3.5) ^b	
PM_{10}	-0.2 (-0.5, 0.1)	4.7 (3.6, 5.9) ^b	-0.2 (-0.7, 0.2)	-0.9 (-1.3, -0.4) ^b	
SO_2	-0.2 (-0.7, 0.1)	5.1 (3.9, 6.3) ^b	-0.1 (-0.6, 0.4)	-0.1 (-0.7, 0.5)	
NO_2	0.7 (0.0, 1.4) ^b	6.0 (3.5, 8.6) ^b	-1.6 (-2.3, -1.0) ^b	-0.1 (-1.2, 1.1)	
O ₃	-1.2 (-1.6, -0.8) ^b	5.6 (4.5, 6.7) ^b	0.6 (0.2, 1.0) ^b	-2.7 (-3.2, -2.2) ^b	

Table 3 Associations between per $10 - \mu g/m^3$ increment in air pollutants and blood lipid levels (n = 15,477).

Abbreviations: HDL-C, high-density lipoprotein cholesterol; LDL-C, low-density lipoprotein cholesterol; NO₂, nitrogen dioxide; O₃, ozone; PM₁, particle with aerodynamic diameter $\leq 1.0 \ \mu m$; PM_{2.5}, particle with aerodynamic diameter $\leq 2.5 \ \mu m$; PM₁₀, particle with aerodynamic diameter $\leq 10 \ \mu m$; SO₂, sulfur dioxide; TC, total cholesterol; TG, triglycerides.

^aAdjusted for age, sex, body mass index, education, family income, smoking, alcohol drinking, exercise, diet, sugary drink intake, family
 history of dyslipidemia, temperature, humidity, wind speed, season, gross domestic product, population density, residuals from regression
 model of highly correlated pollutants, and study district (or community).

^bStatistically significant association (p < .05).

	Odds Ratio (95% Confidence Interval) ^a				
Pollutant	Hypercholesterolemia	Hypertriglyceridemia	Hypoalphalipoproteinemia	Hyperbetalipoproteinemia	
PM_1	1.26 (1.02, 1.57) ^b	1.03(0.91, 1.17)	1.27 (1.06, 1.52) ^b	$1.29 (1.02, 1.64)^{b}$	
PM _{2.5}	1.18 (1.01, 1.37) ^b	1.07 (0.95, 1.19)	1.15 (1.02, 1.30) ^b	$1.28 (1.05, 1.57)^{b}$	
PM_{10}	1.05 (0.92, 1.22)	1.14 (1.01, 1.29) ^b	1.08 (0.88, 1.32)	1.06 (0.89, 1.27)	
SO_2	1.11 (0.79, 1.55)	1.16 (1.00, 1.39) ^b	1.01 (0.79, 1.29)	0.98 (0.77, 1.25)	
NO_2	$1.23(1.02, 1.48)^{b}$	1.21 (0.76, 1.90)	1.27 (0.68, 2.38)	1.25 (0.74, 2.11)	
O_3	0.95 (0.77, 1.19)	1.17 (1.01, 1.36) ^b	0.97 (0.78, 1.20)	0.98 (0.84, 1.15)	

Table 4Associations between per $10-\mu g/m^3$ increment in air pollutants and dyslipidemias (n = 15,477).

Abbreviations: NO₂, nitrogen dioxide; O₃, ozone; PM₁, particle with aerodynamic diameter $\leq 1.0 \ \mu m$; PM_{2.5}, particle with aerodynamic diameter $\leq 2.5 \ \mu m$; PM₁₀, particle with aerodynamic diameter $\leq 10 \ \mu m$; SO₂, sulfur dioxide.

^aAdjusted for age, sex, body mass index, education, family income, smoking, alcohol drinking, exercise, diet, sugary drink intake, family
 history of dyslipidemia, temperature, humidity, wind speed, season, gross domestic product, population density, residuals from regression
 model of highly correlated pollutants, and study district (or community).

 b Statistically significant association (p<.05).

