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Non-random processes determine the colonization of groundwater sediments by microbial communities in a pristine porous aquifer

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Keywords: Community assembly, succession, selection, surface colonization, groundwater, sediments.

Originality-significance statement

Groundwater bodies are the largest terrestrial habitat for microorganisms on Earth, where the majority of the microbial biomass lives attached to sediment surfaces. In these unique, low-productivity environments, microbial communities are the drivers of key biogeochemical processes and furthermore provide important services to society like maintaining groundwater quality as one of the most important sources of freshwater and drinking water worldwide. Over the past years, studies on sediment-attached communities in groundwater-surface water mixing zones as well as surface-attached biofilms in other, non-subsurface habitats have provided important insights regarding the ecological processes that drive the assembly of these microbial communities. Compared to most of these environments, however, pristine groundwater is characterized by significantly lower levels of energy and productivity as well as comparatively more stable environmental conditions, which may promote the effect of stochastic processes on community assembly. Moreover, the microbial communities that colonize subsurface sediments typically exhibit much lower cell densities and occur as small, spatially separated micro-colonies rather than dense, coherent biofilms as they are found in other non-subsurface environments. Therefore, our study was motivated by the question whether findings on the processes that govern microbial community assembly and succession of surface-attached communities in those other more dynamic and nutrient-rich environments also apply to sediment-attached microbial communities in pristine groundwater environments. Our study shows intriguing similarities between the community succession on newly-colonized sediments in our investigated porous, pristine aquifer and succession patterns observed for biofilms in other more dynamic aquatic environments, indicating that the assembly of microbial communities on surfaces may be governed by similar underlying mechanisms across a wide range of different habitats. Our results indicate that differences between planktonic and sediment-attached communities often reported for groundwater environments are not the result of purely stochastic events, but that sediment surfaces select for specific groups of microorganisms that assemble over time in a reproducible, non-random way. Furthermore, our data suggest that specific genera, especially within the *Comamonadaceae* and *Oxalobacteraceae*, played a particularly important role in this process.

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1 **Non-random processes determine the colonization of groundwater sediments**
2 **by microbial communities in a pristine porous aquifer**

3
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7
8 **Summary**

9 Sediments accommodate the dominating share of groundwater microbiomes, however the processes that
10 govern the assembly and succession of sediment-attached microbial communities in groundwater aquifers
11 are not well understood. To elucidate these processes, we followed the microbial colonization of sterile
12 sediments in *in situ* microcosms that were exposed to groundwater for almost one year at two distant but
13 hydrologically connected sites of a pristine, shallow, porous aquifer. Our results revealed intriguing
14 similarities between the community succession on the newly-colonized sediments and succession patterns
15 previously observed for biofilms in other more dynamic aquatic environments, indicating that the
16 assembly of microbial communities on surfaces may be governed by similar underlying mechanisms
17 across a wide range of different habitats. Null model simulations on spatiotemporally resolved 16S rRNA
18 amplicon sequencing data further indicated selection of specific OTUs rather than random colonization as
19 the main driver of community assembly. A small fraction of persistent OTUs that had established on the
20 sediments during the first 115 days dominated the final communities (68%-85%), suggesting a key role of
21 these early-colonizing organisms, in particular specific genera within the *Comamonadaceae* and
22 *Oxalobacteraceae*, for community assembly and succession during the colonization of the sediments.
23 Overall, our study suggests that differences between planktonic and sediment-attached communities often
24 reported for groundwater environments are not the result of purely stochastic events, but that sediment

- 25 surfaces select for specific groups of microorganisms that assemble over time in a reproducible, non-
26 random way.

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27

28 **Introduction**

29 The groundwater-saturated zones of the terrestrial subsurface are one of the largest habitats for
30 microorganisms on Earth (Griebler and Lueders, 2009; McMahon and Parnell, 2013). In these unique,
31 low-productivity environments, microbial communities lie at the heart of key biogeochemical processes
32 like the turnover of carbon and other nutrients, mineral cycling, or pollutant degradation (Griebler et al.,
33 2014; Griebler and Avramov, 2015). Sediment-attached communities play a particularly important role in
34 these ecosystems as they represent the bulk of the microbial biomass and activity (Lehman et al., 2001;
35 Griebler et al., 2002; Zhou et al., 2012; McMahon and Parnell, 2013). Previous studies have repeatedly
36 shown that the composition of sediment-attached communities can differ substantially from planktonic
37 communities suspended in the surrounding groundwater (Zhou et al., 2012; Flynn et al., 2013; Hug et al.,
38 2015). However, the ecological processes that give rise to these differences during community assembly
39 and succession are not well understood. Recent studies have suggested a strong link between
40 biogeochemical functions and microbial community composition as well as the underlying ecological
41 assembly processes (Graham et al., 2016a; Graham et al., 2016b; Graham and Stegen, 2017). Therefore, a
42 better understanding of the processes that drive the assembly of sediment-attached microbial communities
43 in groundwater environments is a key step towards a better understanding of the functioning of these
44 ecosystems ~~functioning~~.

45 The question of the influence of deterministic (or niche-based) versus stochastic (or neutral)
46 processes on the assembly, succession, and diversity of microbial communities has increasingly sparked
47 the curiosity of microbial ecologists over the past years (for reviews see Nemergut et al. (2013); Zhou and
48 Ning (2017)). Deterministic theory assumes that environmental factors, both biotic and abiotic, determine
49 the composition and structure of a community by selecting for species with certain traits that enable them
50 to thrive and compete in a given environment (also known as species sorting or environmental filtering).
51 Accordingly, under similar environmental conditions, communities at different locations or points in time
52 are expected to be composed of species with similar traits. In contrast, the stochastic point of view holds

53 that communities are assembled randomly from species with equivalent trait spectra, and that differences
54 in community composition and structure are the result of random events, for example caused by species
55 dispersal or drift due to stochastic birth-death events (Chase and Myers, 2011). Although purely neutral
56 models have been shown to explain observed diversity patterns of microbial communities with surprising
57 accuracy (e.g. Woodcock et al. (2007); Ofițeru et al. (2010); Woodcock and Sloan (2017)), there has been
58 growing consent that both deterministic and stochastic processes can act simultaneously and that the
59 contribution of either process can shift over time and/or with changing environmental conditions
60 (Dumbrell et al., 2009; Ofițeru et al., 2010; Chase and Myers, 2011; Langenheder and Székely, 2011;
61 Ferrenberg et al., 2013; Stegen et al., 2013; Wang et al., 2013; Zhou et al., 2014; Dini-Andreote et al.,
62 2015; Stegen et al., 2015; Stegen et al., 2016b; Veach et al., 2016; Graham et al., 2017).

63 One aspect where the interaction between deterministic and stochastic processes comes into play
64 is the initial assembly of communities and the following species succession in newly-colonized
65 environments (Tilman, 2004; Langenheder and Székely, 2011), for instance during the development of
66 biofilms on initially empty surfaces (Jackson, 2003; Battin et al., 2007). At the initial stage of
67 colonization, the arrival of species in a new environment is often driven by stochastic dispersal (Tilman,
68 2004; Ferrenberg et al., 2013; Dini-Andreote et al., 2015), which can overrule deterministic effects in
69 homogeneous environments and/or if environmental filtering between the environment of the source
70 community and the newly colonized environment is weak (Stegen et al., 2012; Wang et al., 2013; Battin et
71 al., 2016). However, once established, resident species can affect the establishment of newly-arriving
72 species (positively or negatively) during the subsequent succession directly via species interactions or
73 indirectly by modification of their environment (Fukami, 2015). Thus the order and timing of species
74 arrival, although initially stochastic, can determine the composition and functioning of the final
75 community, known as priority effect (Fargione et al., 2003; Tilman, 2004; Fukami et al., 2010; Peay et al.,
76 2012; Tan et al., 2012; Nemergut et al., 2013; Rummens et al., 2018; Svoboda et al., 2018).

77 A general, conceptual model that summarizes the successional stages during biofilm development
78 has been described by Jackson (2003). According to this model, initially empty surfaces offer ample space

79 and resources to allow for the establishment of diverse microorganisms resulting in a rapid increase in
80 species richness and diversity that is fueled by the dispersal of newly-arriving species from a regional
81 species pool like overlaying water during initial biofilm assembly. The steady arrival of new species
82 eventually leads to niche depletion and growing competition between established and newly-arriving
83 species, which more and more suppresses the increase in species richness. As the competition intensifies,
84 less competitive species are lost from the community, which leads to a decline in species richness after the
85 initial phase of community assembly. However, as the biofilm matures further and becomes more
86 heterogeneous, new niches are created that enable specialized species to establish, leading again to an
87 increase in richness and diversity in the mature biofilm. Although Woodcock and Sloan (2017) have
88 demonstrated using a neutral modeling approach that these patterns can be explained based on stochastic
89 processes only, empirical evidence suggests that the assembly of biofilm communities is in fact
90 characterized by a shift from initially stochastic community assembly towards deterministically driven
91 succession at the later stages, for instance caused by species interactions or growing niche space due to
92 increasing spatial and chemical heterogeneity (Martiny et al., 2003; Lyautey et al., 2005; Battin et al.,
93 2016; Veach et al., 2016; Brislawn et al., 2018).

94 To date, most of the studies on ecological processes behind the assembly of microbial
95 communities in groundwater environments have focused on planktonic communities suspended in the
96 groundwater (Stegen et al., 2012; Stegen et al., 2013; Beaton et al., 2016; Danczak et al., 2018), while
97 studies on sediment-attached communities are scarce (Stegen et al., 2016a). In contrast, much insight has
98 been gained over the past years into the assembly of sediment-attached communities in groundwater-
99 surface water mixing zones (hyporheic zone). In these studies, the assembly of planktonic communities
100 generally tended to be more subject to stochastic effects and shifts in assembly processes related to
101 changes in water chemistry, whereas selection had a relatively more pronounced effect on the assembly of
102 sediment-attached communities which, at the same time, was less affected by hydrochemical changes
103 (Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et al., 2018). Compared to the
104 hyporheic zone, pristine groundwater environments (in the absence of surface water impacts) are more

105 stable and only experience little environmental changes (Griebler and Lueders, 2009), which may promote
106 the effect of stochastic processes on community assembly compared to more dynamic environments
107 (Ofițeru et al., 2010; Stegen et al., 2012; Wang et al., 2013; Zhou et al., 2013). Moreover, in contrast to
108 the typically studied biofilms in other environments like surface waters, which form dense, spatially
109 coherent, heterogeneous structures that can reach a thickness in the range of several hundred micrometers
110 (Battin et al., 2016), sediment-attached microbial communities in groundwater aquifers occur as small,
111 patchily distributed micro-colonies that consist of only a few cells (Schmidt et al., 2017), which may be
112 hypothesized to be more prone to stochastic effects than their biofilm counterparts in other environments.

113 In this study we set out to 1) investigate whether the assembly of sediment-attached microbial
114 communities in pristine groundwater environments can be explained by the general patterns observed for
115 surface-attached biofilms in other environments; 2) study the importance of early colonizers for
116 community succession during the colonization of the sediments; and 3) see if the dominating role of
117 selection on community assembly that has been observed for sediment-attached microbial communities in
118 the hyporheic zone also drives community assembly in comparatively stable, pristine groundwater
119 environments. To tackle these goals, we incubated *in situ* microcosms filled with sterilized sediment in
120 monitoring wells at two distant but hydrologically connected sites of a pristine, porous aquifer (Zhou et
121 al., 2012), and followed the succession of the microbial communities as they colonized the sediments over
122 a period of almost one year. We used 16S rRNA amplicon sequencing data to study changes in alpha and
123 beta diversity of the sediment-attached communities incubated at the two sites as well as differences in
124 community composition between sediment-attached and planktonic communities in the surrounding
125 groundwater over the course of the succession. To explore the influence of deterministic and stochastic
126 processes on microbial community assembly and succession, we applied the null model approach
127 developed by Stegen et al. (2012; 2013), which has previously also been used in studies on community
128 assembly in the hyporheic zone (Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et
129 al., 2018) as well as biofilms in other environments (Langenheder et al., 2017; Brislawn et al., 2018), and
130 thus allows us to compare our results to those previous findings.

131

132 **Results**

133 **Site description**

134 The field experiment was conducted over a period of 347 days from March 2010 until February 2011,
135 with intermediate sampling campaigns in May (day 49), July (day 115), and December (day 263). The
136 area with the two monitoring wells used for the incubation of the *in situ* microcosms was located at the
137 foothills of the Bavarian Alps in the upper Isar River valley close to the village of Mittenwald, Germany
138 (Fig. 1). The wells were installed in a pristine, shallow, porous aquifer composed of quaternary sediment
139 mainly consisting of gravel and coarse sand. Well MIT052 was located on a mountain pasture in the
140 forested Riedboden nature reserve 400 m away from the nearby river; well MIT039 was located
141 approximately 2 km away from MIT052 in proximity to the village with a distance of 240 m to the river
142 (for a detailed site description, see Zhou et al. (2012)). Over the course of the experiment, we observed
143 only small fluctuations in physicochemical conditions (Table 1; the temporal dynamics of the individual
144 parameters are shown in Fig. S1).

145

146 **Temporal dynamics of microbial biomass and alpha diversity of sediment-attached and planktonic** 147 **microbial communities**

148 Despite the spatial distance between the two sites, the microbial communities that developed on the
149 initially sterile sediments followed identical trends in alpha diversity and biomass patterns (Fig. 2).
150 Already after the first 49 days, the microbial biomass (measured as prokaryotic cell counts) of attached
151 microbial communities at both sites had reached a plateau of $\sim 10^7$ cells cm^{-3} of sediment followed by a
152 slight decline for the remaining time of the experiment. Although the biomass of sediment-attached
153 microbial communities stayed more or less constant, noticeable changes in the communities still occurred
154 as indicated by OTU richness and diversity which steadily increased by about 50% and 25%, respectively,
155 from May until December, followed by a decline of both parameters in February. Over the same period,
156 community evenness remained relatively high and only changed moderately. The changes in biomass and

157 alpha diversity observed for the newly colonized sediments appeared to be independent from the changes
158 that occurred in the planktonic microbial communities, which were more variable over time and less
159 comparable between the two sites. Microbial biomass was about one to two orders of magnitude lower for
160 the planktonic communities compared to the sediments.

161

162 **Establishment and persistence of newly-arriving OTUs in sediment-attached microbial communities**

163 To assess the impact of early colonizers on microbial community succession, we looked at the number of
164 newly-arriving OTUs that entered the developing sediment-attached communities at each time point over
165 the course of the experiment (Fig. 3). Newly-arriving OTUs are defined here as OTUs that showed an
166 abundance > 0% in the community for the first time at a given time point. At both sites, the number of
167 newly-arriving OTUs declined over time showing that the majority of OTUs had established during the
168 initial phase of the incubation. Despite this declining trend, the fraction of newly-arriving OTUs relative to
169 the total OTU richness at the end of the incubation was still noticeable with ~15-20%. However, looking
170 at the changes in the cumulative relative abundances of the newly-arriving OTUs over time, we saw that
171 the OTUs that had arrived towards the later stages only accounted for a relatively small fraction of the
172 final communities. Even though the cumulative relative abundance of OTUs that had established in the
173 communities within the first 49 days steadily declined, these OTUs still made up 36% and 47% of the
174 final communities at MIT052 and MIT039, respectively. At MIT052, these OTUs together with those that
175 emerged at the following time point in July comprised the majority of the final community at the end of
176 the incubation (together 85%), while OTUs that arrived at the final time point accounted for only 5%. At
177 MIT039, OTUs that had arrived at the first two time points made up for 68% of the final community,
178 whereas OTUs that had arrived in December and February comprised 12% and 20%, respectively.
179 Although these results clearly show the dominance of early-colonizer OTUs in the final communities, a
180 closer look at how many of these OTUs actually persisted until the final time point showed that only
181 ~12% of newly-arriving OTUs from each time point were still present in the final communities (data not
182 shown). Looking at the taxonomies of these newly-arriving OTUs that persisted until the end of the

183 incubation, we again found highly similar patterns for both sites (Fig. S2). At each time point, the most
184 dominant groups were OTUs affiliated to *Comamonadaceae*, mainly *Aquabacterium* and *Polaromonas*
185 spp., in addition to *Oxalobacteraceae*, mainly consisting of *Duganella*, *Massilia*, and *Undibacterium* spp.,
186 as well as *Pseudomonas* spp. and diverse *Caulobacteraceae* and *Sphingomonadaceae*.

187

188 **Microbial community composition and beta diversity**

189 Similar to the alpha diversity patterns, the microbial communities on the newly colonized sediments
190 displayed comparable compositions at the two sites (Fig. 4). Especially during the initial phase of the
191 incubation in May, sediment-attached communities at both sites were dominated by *Oxalobacteraceae* in
192 addition to *Comamonadaceae* and smaller fractions of *Flavobacteriaceae* and *Caulobacteraceae*. Over the
193 further course of the incubation, these taxonomic groups gradually receded and were in part replaced
194 mainly by increasing numbers of *Comamonadaceae* (mainly *Aquabacterium* spp.), *Pseudomonadaceae*,
195 *Nocardiaceae*, and *Rhodocyclaceae* especially at MIT052, in addition to *Sphingomonadaceae*, uncultured
196 *Deltaproteobacteria*, and *Moraxellaceae* at MIT039. Moreover, OTUs affiliated with diverse low-
197 abundant families (with an abundance <10% in the entire dataset; mean = 0.1%; max. = 4.7%) gradually
198 increased in abundance. In contrast, planktonic communities were mainly dominated by members of the
199 *Rhodocyclaceae*, *Comamonadaceae* (mainly *Curvibacter*, *Simplicispira*, and *Rhodoferax* spp.), and
200 *Leptospiraceae*.

201 To get a better picture of the organisms that were responsible for differences between sediment-
202 attached and planktonic communities, we performed similarity percentage (SIMPER) analysis across all
203 samples on relative abundances of OTUs grouped at genus level. Interestingly, we found high agreement
204 between the genera that significantly contributed to the observed dissimilarities between the two types of
205 communities and the taxa identified as successful, persistent colonizers in the previous analysis (Fig. S2).
206 *Aquabacterium*, *Massilia*, and *Duganella* spp. ranked among the genera that contributed most to the
207 dissimilarity (together > 15%; all $p < 0.002$) and were highly differentially abundant in the sediment-
208 attached communities, next to *Flavobacteria* and uncultured members of the *Oxalobacteraceae* (Fig. S3).

209 The changes in microbial community composition over time as well as differences between
210 sediment-attached and planktonic communities in the groundwater were revealed by non-metric
211 multidimensional scaling (NMDS) performed on abundance-weighted β -mean nearest taxon distance (β -
212 MNTD) between communities (Fig. 4). At all measured time points, sediment-attached and planktonic
213 communities clustered separately from each other as reflected by the distinct separation of the two types
214 of communities along the first NMDS axis. Changes in microbial community composition over time were
215 reflected by the separation of data points along the second NMDS axis. Permutational analysis of variance
216 (PERMANOVA) revealed that community type (i.e. sediment-attached vs. planktonic) explained most of
217 the variance in β -MNTD between communities ($R^2 = 0.626$; $p = 0.001$), followed by sampling time point
218 ($R^2 = 0.104$; $p = 0.001$), while site location was not significant ($R^2 = 0$; $p = 1$), showing that communities
219 across sites were similar within each community type and time point. Moreover, fitting environmental
220 variables to the NMDS ordination with stratification of permutations within the community types did not
221 reveal significant correlations between changes in community composition and any of the measured
222 physicochemical parameters (Table 1 and Figure S1) (all $R^2 < 0.32$; $p > 0.1$).

223 Since community type explained most of the variance in beta diversity, we applied partitioning of
224 beta diversity to identify the underlying causes of the differences between sediment-attached and
225 planktonic communities within sites and sampling time points according to Baselga (2012). This approach
226 is based on the additive partitioning of incidence-based Jaccard dissimilarity between two communities
227 into a nestedness and a turnover component. A high contribution of nestedness to the total dissimilarity
228 indicates that two communities are subsets of each other and that differences are caused by differences in
229 species richness (i.e. gain or loss of species). On the other hand, a high contribution of turnover indicates
230 little overlap in species composition, i.e. species in one community have been replaced by other species in
231 the other community. The analysis showed that turnover was the dominating process behind the
232 differences between the two types of communities at each time point (for all time points $> 97\%$) (Fig. S4),
233 showing that both community types were composed of distinct sets of OTUs.

234 We applied the same approach to the dissimilarity between sediment-attached communities at
235 different time points within sites to investigate the degree to which nestedness and turnover contributed to
236 changes in community composition over the course of the sediment colonization. Also in this case
237 turnover dominated over nestedness in all comparisons, especially over long time scales (i.e. comparing
238 communities between May and February; 95-97%), and with a slightly weaker effect over short time
239 scales of the succession (i.e. comparing communities between consecutive time points; 74-90%) (Fig. S5).

240 241 **Impact of stochastic and deterministic processes on community assembly and succession inferred** 242 **from null models**

243 To infer the impact of deterministic and stochastic processes on community assembly, we applied the two-
244 tiered null model approach developed by Stegen et al. (2012; 2013). Briefly, under the assumption that
245 phylogenetically closely related species occupy more similar ecological niches than less closely related
246 species, the strength of deterministic processes is evaluated in tier one based on the β -nearest taxon index
247 (β -NTI). β -NTI < -2 and > +2 indicates that two communities are phylogenetically significantly more or
248 less similar to each other than expected by chance, which is interpreted as homogeneous selection (i.e.
249 selection of similar OTUs) or variable selection (i.e. selection of dissimilar OTUs) in the two
250 communities, respectively. $|\beta$ -NTI| < 2 indicates that two communities are as dissimilar as expected by
251 chance, hinting at stochastic community assembly. In this case, the RC_{bray} index is used in tier two to
252 evaluate the effect of stochastic dispersal. $RC_{\text{bray}} < -0.95$ and $> +0.95$ means that two communities share
253 significantly more or less OTUs than expected by chance, indicating that communities are assembled by
254 homogenizing dispersal or dispersal limitation in combination with drift, respectively. $|RC_{\text{bray}}| < 0.95$
255 indicates that differences between two communities are due to random drift acting alone. We applied this
256 approach to study the role of deterministic and stochastic processes on three levels (Fig. 5): 1) spatial
257 community turnover between the two sites within community types and time points; 2) turnover between
258 sediment-attached and planktonic communities within sites and time points; 3) temporal turnover between
259 communities at consecutive time points within community types and sites.

260 Deterministic and stochastic processes had different influences on the spatial community turnover
261 for sediment-attached and planktonic communities, respectively. In case of planktonic communities,
262 pairwise comparisons between sites for each time point resulted in values for β -NTI that were not
263 significantly different from the null expectation, indicating that differences between the planktonic
264 communities at the two sites were caused by stochastic processes. Pairwise comparisons based on the
265 RC_{bray} index identified homogenizing dispersal as the dominating process (all $RC_{\text{bray}} = -1$; the only
266 exception was observed for December: $RC_{\text{bray}} = +0.99$, indicating dispersal limitation together with drift;
267 data not shown). In contrast to the predominantly stochastic exchange of OTUs between the two sites via
268 dispersal through the groundwater, pairwise comparisons of sediment-attached communities clearly
269 tended towards homogenous selection as main cause of the similarities between the sediment-attached
270 communities (with one exception observed for July where β -NTI was not significant, but a slightly
271 significant $RC_{\text{bray}} = 0.97$ hinted at dispersal limitation together with drift).

272 Since the sediments had to be colonized by microorganisms that were recruited from the
273 surrounding groundwater, even though beta diversity partitioning revealed little overlap between these two
274 communities as discussed above, we tested for the effect of selection processes on the assembly of
275 sediment-attached communities from planktonic microorganisms suspended in the groundwater. The
276 differences between the communities on the newly colonized sediments and the planktonic communities at
277 each site were caused by stochastic process during the first 115 days of the incubation. Pairwise
278 comparisons based on RC_{bray} hinted at dispersal limitation in combination with drift as the processes
279 responsible for these differences (all $RC_{\text{bray}} = +1$; data not shown). This trend changed at the later stage in
280 December after 263 days; at this point significantly positive values for β -NTI hinted at variable selection
281 of phylogenetically distinct OTUs in sediment-attached communities compared to the microorganisms in
282 the surrounding groundwater.

283 Unlike the trends observed for the spatial community turnover, the influence of deterministic and
284 stochastic effects on changes in community composition that occurred over time was much more variable
285 and no clear trends could be observed. Although selection effects appeared to have played a role (both

286 homogenous and variable selection), they mostly did not occur consistently at both sites for neither
287 sediment-attached nor planktonic communities.

288 We used Mantel correlation analysis to investigate whether changes in individual physicochemical
289 parameters in the groundwater had an effect on the changes in assembly processes (based on β -NTI).
290 Similar to the lack of correlations between environmental variables and differences in community
291 composition mentioned above, we did not find significant effects of changes in environmental conditions
292 in this analysis for neither planktonic (all |Spearman's rho| < 0.34; $p > 0.08$) nor sediment-attached
293 communities (all |Spearman's rho| < 0.27; $p > 0.1$).

294

295 Discussion

296 The alpha diversity patterns for the newly-colonized sediments at both sites followed identical trends that
297 closely matched the conceptual model for the formation of biofilms on empty surfaces outlined by Jackson
298 (2003), which describes changes in alpha diversity over three main stages of biofilm development. At the
299 early stage, the large niche space of an initially empty surface allows for the establishment of diverse
300 microorganisms, resulting in a steady increase in alpha diversity, which subsequently levels off and
301 eventually declines due to niche depletion and the loss of less competitive species as the biofilm grows
302 over the course of the succession. However, at the final stage, the mature biofilm becomes increasingly
303 spatially and chemically heterogeneous, which opens new niches for specialized species to thrive and
304 thereby fuels a renewed increase in alpha diversity. Our results only deviated from this model at the final
305 stage of the incubation, where we did not see an increase in species diversity and richness, ~~which~~
306 ~~according to the model should occur due to the presence of specialized niches that are brought about by~~
307 ~~the spatial and chemical heterogeneity in mature biofilms.~~ However, this framework was conceptualized
308 for biofilms in resource-rich, high-productivity environments like activated sludge, wetlands, and lakes
309 (Jackson, 2003). Although we cannot exclude that alpha diversity may have increased again with a
310 prolonged incubation time, we may argue that diverse, specialized niches that develop in mature, spatially
311 heterogeneous biofilms might not form to such an extent in the small, patchily distributed micro-colonies

312 that typically colonize groundwater sediments (Schmidt et al., 2017). Hence, the total niche space in such
313 micro-colonies may be smaller compared to mature biofilms in other environments, similar to what
314 Graham et al. (2016a) have proposed for sediments in the hyporheic zone. Moreover, although the general
315 pattern of decreasing fractions of newly-arriving OTUs was also apparent in our experiment, reflecting the
316 saturation of niche space according to Jackson's biofilm model (Jackson, 2003), we noticed that the
317 fraction of these OTUs at the end of the incubation was still 5-10 times higher compared to findings on
318 biofilms in other environments (e.g. Brislawn et al. (2018)). These deviations of our results from assembly
319 patterns of biofilms, together with the findings made for hyporheic zone sediments (Graham et al., 2016a),
320 might point towards important differences in ecological niche structures between biofilms in resource-rich
321 surface environments and sediment-attached microbial communities in the typically more energy-poor and
322 less productive subsurface.

323 Looking at the abundance changes of newly-arriving OTUs over time, we saw that OTUs
324 colonizing the sediments during the early stage of community assembly (i.e. the first 49 to 115 days)
325 largely dominated the final communities at the end of the experiment. However, at the same time, these
326 dominant OTUs represented only a small fraction of newly-arriving OTUs found at each time point. This
327 was further reflected by the large dominance of OTU turnover over nestedness between successional
328 stages in the sediment-attached communities inferred from beta diversity partitioning, showing that the
329 majority of OTUs that had established at a given time point were in fact replaced by others over the course
330 of the succession. Therefore, in agreement with the findings by Brislawn et al. (2018), the mere timing of
331 OTU arrival did not seem to be a determining factor for the final community structure. Rather, the
332 consistent dominance of specific taxa among these persistent OTUs (mainly genera belonging to the
333 *Oxalobacteraceae*, *Comamonadaceae*, *Caulobacteraceae*, *Sphingomonadaceae*, in addition to
334 *Pseudomonas* spp.) suggests the involvement of certain traits that enable these taxa to sustainably colonize
335 and thrive on sediment surfaces. Interestingly, we also found the same genera among the most important
336 contributors to differences between sediment-attached and planktonic communities and to be highly
337 differentially abundant in the former. The association of these taxa with biofilms and traits that facilitate

338 surface colonization like motility or production of extracellular polysaccharides have been reported before
339 for other environments (Kalmbach et al., 2000; Baldani et al., 2014; Bižić-Ionescu et al., 2014;
340 Niederdorfer et al., 2016; Niederdorfer et al., 2017), supporting the hypothesis about their importance for
341 the development of sediment-attached communities in our study. Over the course of the succession, these
342 dominant OTUs may have facilitated the recruitment of other more diverse taxa that were observed at the
343 later stages of the colonization (Battin et al., 2007; Nemergut et al., 2013; Fukami, 2015).

344 Comparisons of beta diversity patterns revealed that sediment-attached and planktonic
345 communities, respectively, were similar at each time point across the two sampling locations. Using the
346 null model approach developed by Stegen et al. (2012; 2013) revealed that different processes were
347 responsible for the observed similarities. Whereas the spatial turnover of planktonic microbial
348 communities was driven by stochastic processes, mostly homogenizing dispersal (75%), the high
349 similarities between the sediment-attached communities at the two sites were mostly caused by
350 homogenous selection (75%). We are aware that our study consists of only a relatively limited number of
351 observations and therefore the results should be interpreted with the necessary caution. Nevertheless, our
352 results fit observations on assembly processes for communities in the hyporheic zone (Graham et al.,
353 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et al., 2018) as well as ~~for~~ biofilms in surface
354 water streams (Besemer et al., 2012; Veach et al., 2016), suggesting that selection not only plays a
355 determining role in the assembly of surface-attached microbial communities in those dynamic
356 environments but also in pristine groundwater aquifers, despite the comparatively more stable
357 environmental conditions, which have been shown to promote the effect of stochastic over deterministic
358 processes in other environments (Ofițeru et al., 2010; Stegen et al., 2012; Wang et al., 2013; Zhou et al.,
359 2013). Mineral composition has previously been demonstrated to be a driving factor for microbial
360 community composition and assembly (Grösbacher et al., 2016; Stegen et al., 2016a; Jones and Bennett,
361 2017). Since the *in situ* microcosms that we incubated at the two sites in our study were filled with
362 sediment that originated from the same source, it is likely that identical sediment properties selected for
363 the highly similar microbial communities at the two sites.

364 Given the high similarities between the sediment-attached communities at both sites throughout
365 the experiment, we would have expected to also find similar patterns regarding the processes that drove
366 the temporal microbial community turnover. However, contrary to this expectation, this was not fully the
367 case as assembly was highly variable without a clearly discernable trend in favor of a single process.
368 Changes in environmental conditions such as nutrient inputs, fluctuating water tables, or surface water-
369 groundwater mixing have been observed to not only affect the composition of (groundwater) microbial
370 communities, but also influence the ecological assembly processes that determine those changes (Lyautey
371 et al., 2005; Stegen et al., 2012; Stegen et al., 2013; Zhou et al., 2014; Dini-Andreote et al., 2015; Stegen
372 et al., 2015; Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017). However, in our case, we did
373 not find indications that changes in physicochemical conditions of the groundwater were related to
374 changes in community composition or shifts in ecological community assembly processes. This could
375 suggest that the changes in community composition over time and the influence of deterministic versus
376 stochastic effects were determined by changes in unmeasured environmental variables (Stegen et al.,
377 2013). Alternatively, the observed lack of correlations between changes in environmental conditions and
378 the processes that determined community assembly can also hint at the impact of endogenous factors like
379 species interactions (Konopka et al., 2015; Battin et al., 2016; Cordero and Datta, 2016). Recently,
380 Danczak et al. (2018) could show that interaction network structures can affect assembly processes of
381 planktonic microbial communities in pristine aquifers. Although our results show that the assembly of
382 sediment-attached communities was mainly deterministic, and that the succession of OTUs was highly
383 reproducible between the two sites, the compositions of the two communities at each time point, and
384 therefore possibly interaction networks, were not totally identical. Hence, the variable patterns of
385 processes that determined the community turnover between successional stages at each site might, at least
386 in part, be attributed to possible differences in interaction networks within the communities between the
387 two sites.

388 An additionally important factor in the assembly and succession of surface-attached communities
389 in aquatic environments is the invasion by species from the surrounding water phase (Battin et al., 2016).

390 The establishment of invading species in a biofilm community depends on both stochastic dispersal as
391 well as interactions with already established species (Battin et al., 2007; Battin et al., 2016). Beta diversity
392 partitioning showed that sediment-attached and planktonic communities were composed of distinct sets of
393 OTUs. We again used the null model approach to test in how far deterministic and stochastic processes
394 contributed to these differences. We found that over the first successional stages the turnover between
395 sediment-attached and planktonic communities was caused by dispersal limitation acting alongside drift
396 and later on shifted towards variable selection. The latter observation could be explained in the light of
397 previous studies which suggested that species with similar ecological niches as resident species have a
398 lower chance of successfully invading a community than species that have less niche overlap with already
399 established species (Fargione et al., 2003; Tilman, 2004; Peay et al., 2012; Tan et al., 2012).

400 The processes that were indicated to have driven community turnover between groundwater and
401 sediment at the earlier stages were however counterintuitive. Unexpectedly, significantly positive values
402 for RC_{bray} suggested dispersal limitation acting alongside drift to have been responsible for the observed
403 differences in community composition, rather than the intuitively more expected scenario of random drift
404 acting alone. Multiple causes could explain these unexpected findings. It has to be noted that the sediment
405 microcosms were incubated in groundwater monitoring wells. It is known that communities found inside
406 monitoring wells may differ from the communities that are actually present in the surrounding
407 groundwater of an aquifer (Griebler et al., 2002; Korbel et al., 2017). In fact, previous analyses of our
408 samples by T-RFLP fingerprinting did indeed reveal some differences between groundwater and well
409 water microbial communities (Zhou et al., 2012). However, Langenheder et al. (2017) have reported
410 identical results for differences between lake biofilms and microbial communities in the overlaying water
411 column, which were not separated by any barrier that could have limited OTU dispersal. They argued in
412 the light of these findings, and based on the arguments provided by Chase et al. (2011), that significantly
413 positive deviations of RC_{bray} from the null expectation may also be caused by strong biotic factors such as
414 competition between species. As niches become more crowded over time, some organisms may try to
415 avoid competition by occupying non-optimal niches, which would not necessarily result in a deviation

416 from the null expectation in phylogenetic null models. Moreover, even though the assumption underlying
417 the β -NTI-based approach about the link between phylogenetic relatedness and niche similarity of
418 microbial species is supported by empirical evidence (Peay et al., 2012; Stegen et al., 2012; Tan et al.,
419 2012; Wang et al., 2013; Dini-Andreote et al., 2015; Martiny et al., 2015), and was also confirmed in our
420 system by a significant phylogenetic signal (Fig. S6), it is known that some species traits are
421 phylogenetically more conserved than others (Martiny et al., 2015). Hence, we may speculate that the
422 significantly positive deviation of RC_{bray} from the null exception at the early stage of the colonization
423 might indicate the involvement of traits that are important for the colonization of sediment surfaces, but
424 which are phylogenetically not well conserved and therefore did not result in a significant signal of β -NTI.
425 Only at the later stage, when the communities on the sediments had matured further, phylogenetically
426 more conserved traits may have gained importance in the turnover between planktonic and attached
427 microbial communities.

428

429 **Conclusion**

430 We have shown that the microbial colonization of sediments in a pristine groundwater aquifer in several
431 aspects follows the general patterns that have also been described for the development of biofilms in other
432 more energy-rich, non-subsurface, aquatic environments (Jackson, 2003), as well as the assembly of
433 sediment-attached communities in highly dynamic hyporheic zones, suggesting that the assembly of
434 microbial communities on surfaces might be governed by similar underlying mechanisms across a wide
435 range of different habitats. Our results indicate that differences between planktonic and sediment-attached
436 communities often reported for groundwater environments are not the result of purely stochastic events,
437 but that sediment surfaces select for specific groups of microorganisms that assemble over time in a
438 reproducible, non-random way, probably determined by sediment properties rather than hydrochemistry.
439 Although we found that early-colonizing OTUs dominated the final communities on the sediments, mere
440 timing OTU of arrival during the succession was likely not a determining factor, as the majority of these
441 early-colonizers were not very persistent. Rather, traits associated with identified key taxa, especially

442 within the *Comamonadaceae* and *Oxalobacteraceae*, seemed to have been a more decisive factor for the
443 persistence of these OTUs. However, the ecological processes behind the temporal succession of OTUs
444 during the colonization still remain unclear and might be influenced by species interaction network
445 structures at a given time point. Moreover, we found indications that different traits with different degrees
446 of phylogenetic conservation may have determined the establishment of OTUs in the developing
447 sediment-attached communities from the surrounding groundwater at different stages of community
448 development. A better understanding of these traits and how they may integrate into species interaction
449 networks will be an important aspect for future research. Computational modelling of microbial
450 communities based on metaomics data, albeit still in its infancy, offers a promising tool to elucidate
451 complex species interactions within microbial communities (Faust and Raes, 2012; Hanemaaijer et al.,
452 2015). If successful, the extra in depth insight gained from such models could be a valuable addition to
453 current approaches that strive for a better understanding of the links between microbial community
454 composition, assembly, and biogeochemical functions (Graham et al., 2016b; Graham and Stegen, 2017) .
455

456 **Experimental procedures**

457 **Experimental setup and sampling**

458 To study the assembly and succession of sediment-attached microbial communities, fresh sediments were
459 taken from the Isar River that drains the investigated aquifer. Sediments were sieved (0.2-0.63 mm) and
460 packed into perforated polyethylene columns with a mesh size of 1-2 mm. Sediment columns were
461 submerged in deionized water and sterilized by autoclaving five times at 121°C for 30 min; after each
462 step, the sediments were rinsed with and again submerged in fresh deionized water. The columns were
463 stored at 4°C submerged in sterile water until the start of the experiment. Replicate sediment columns
464 were incubated in each well; duplicate columns were sampled destructively at each sampling campaign.
465 Samples for DNA extraction were put on dry ice for transport to the lab and were stored at -20°C until
466 DNA extraction according to the method described by Anneser et al. (2010). For the comparison of
467 attached versus planktonic microbial communities, cells from 5 L groundwater were collected on a 0.2 µm

468 polycarbonate filter (Merck Millipore, Darmstadt, Germany) on-site. Filters were shock-frozen on dry ice
469 and stored at -20°C until extraction using the same method as for the sediment samples. For cell counting,
470 0.5 mL groundwater (or 0.5 cm³ sediment) was fixed on-site with glutardialdehyde at a final concentration
471 of 2.5% v/v; samples were stored in the dark at 4°C until further processing according to Bayer et al.
472 (2016). Cells were stained with SYBR-Green I (Invitrogen, Karlsruhe, Germany) at a ratio of 1:10,000
473 and subsequently counted using a LSR II flow cytometer (Becton Dickinson, Heidelberg, Germany). For a
474 description of measurements of physicochemical parameters listed in Table 1 the reader is referred to
475 Zhou et al. (2012).

476

477 **16S rRNA amplicon sequencing**

478 PCR amplification (28 cycles) and subsequent bidirectional 454-pyrosequencing of 16S rRNA gene
479 fragments was done according to Pilloni et al. (2011) using the primer pair Ba27f-Ba519r extended with
480 sequencing adapters and multiplex barcodes. Each of the sample duplicates was amplified again in
481 duplicate; after amplification, all replicates of a given sample were combined before purification using
482 magnetic beads (AMPure-XP; Beckmann Coulter, Brea, CA, USA) according to the manufacturer's
483 instructions. After purification, DNA concentrations were determined using the Quant-iT PicoGreen
484 dsDNA Assay Kit (Invitrogen, Paisley, UK). Barcoded amplicons from all samples were pooled in
485 equimolar amounts before sequencing on a 454 GS FLX pyrosequencer using Titanium chemistry (Roche,
486 Penzberg, Germany). The sequence data was processed with QIIME (version 1.9.0) (Caporaso et al.,
487 2010). Demultiplexing and quality filtering (min./max. sequence length: 250/600 bp; primer mismatches
488 and barcode errors: 0; min. quality score: 25; quality score window size: 50 bp) was done using the
489 'split_libraries.py' command. Chimera filtering was done by mapping reads against the SILVA SSU
490 reference database (release 128) (Quast et al., 2013) using 'identify_chimeric_seqs.py' with usearch61 as
491 detection method. After quality and chimera filtering, the average number of combined forward and
492 reverse reads per sample was 5,709 with an average length of 388 bp. OTUs were clustered by uclust
493 against the SILVA SSU reference database at 97% similarity using the 'pick_open_reference_otus.py'

494 command. After removing low-confidence OTUs (combined abundance of < 0.01% across all samples)
495 and OTUs classified as chloroplasts, a total of 910 OTUs remained in the final OTU table. The total
496 number of reads per sample was rarefied to 2,045 which was the lowest number of reads observed for a
497 single sample. A midpoint-rooted phylogenetic tree was constructed from the alignment of OTU
498 reference sequences using FastTree (Price et al., 2009). Sequence data have been deposited in the NCBI
499 Sequence Read Archive under accession number SRP139256.

500

501 **Data analysis**

502 All analyses were done in R (version 3.5.0) (R Core Team, 2018). Alpha diversity (OTU richness (S),
503 Shannon diversity (H'), Pilon's evenness (J')) was calculated using the vegan package (version 2.5-2)
504 (Oksanen et al., 2018). The number of newly-arriving OTUs (S_{nt}) in sediment samples at a given time
505 point was defined as the number of OTUs that displayed an abundance > 0% for the first time at that time
506 point. Phylogenetic beta diversity was assessed based on β -MNTD that was calculated using the
507 'comdistnt' function of the picante package (version 1.7) (Kembel et al., 2010). Differences in microbial
508 community composition between samples across time, space, and community type were illustrated by
509 NMDS performed on the β -MNTD matrix using the 'metaMDS' function of the vegan package with 40
510 iterations. To test for the effect of physicochemical variables (Table 1) on changes in community
511 composition for sediment-attached and planktonic communities, respectively, variables were standardized
512 to z-scores before fitting to the NDMS ordination using the 'envfit' function of the vegan package with
513 10,000 permutations stratified within community types. PERMANOVA was used to estimate the marginal
514 effects of each of the three categorical variables community type, sampling time point, and site location,
515 respectively, while holding the other two constant using the 'adonis2' function in vegan package with
516 10,000 permutations. For the identification of key organisms that were responsible for the differences
517 between community types, relative OTU abundances were summarized at genus level before SIMPER
518 analysis using the 'simper' function in vegan with 1,000 permutations for significance testing. Beta

519 diversity partitioning based on Jaccard dissimilarity was done using the ‘betapart’ package (Baselga and
520 Orme, 2012).

521 To study the effect of deterministic versus stochastic processes on microbial community assembly,
522 we used the null model approach developed by Stegen et al. (2012; 2013). β -NTI compares the mean
523 phylogenetic distance of OTUs based on β -MNTD between two communities against the distribution of β -
524 MNTD values expected for randomly assembled communities. This distribution is obtained from repeated
525 randomizations in which the OTUs observed in the two communities and their relative abundances are
526 shuffled across the tips of the according phylogenetic tree. The value of β -NTI indicates by how many
527 standard deviations the observed β -MNTD deviates from the mean of the null expectation with $|\beta\text{-NTI}| > 2$
528 indicating significant deviations. β -NTI was calculated with abundance-weighting and 999 randomizations
529 for each pairwise comparison. The assumption of a significant phylogenetic signal was verified using
530 Mantel correlograms as in Dini-Andreote et al. (2015) (see SI and Fig. S6). The RC_{bray} index measures
531 how much the observed Bray-Curtis dissimilarity between two communities differs from the distribution
532 of dissimilarities between probabilistically assembled communities for which the probability of OTUs
533 being drawn is proportional to their respective abundances in the two compared communities and their
534 occurrence frequencies in the regional species pool, while keeping local community richness and the
535 number of individuals constant. RC_{bray} takes values from -1 to +1 where absolute values > 0.95 indicate
536 significant deviations from the null expectation. RC_{bray} was calculated with 999 iterations for each
537 pairwise comparison. Regional species pools for null model simulations were constructed from all OTUs
538 in the full dataset over space and time as in Veitch et al. (2016), because we expected that regional species
539 pools constructed separately for each time point from OTUs at the two sites that only spanned a relatively
540 short transect would have been too conservative to estimate the total regional diversity in the aquifer. To
541 evaluate in how far this large regional species pool may have led to an overestimation of the effects of
542 selection and/or dispersal, we compared these results to simulations where regional species pools were
543 constructed for individual time points for which paired samples of sediment-attached and planktonic
544 communities were available. The outcomes of the null models in both situations were in high agreement

545 with each other, indicating that using the full dataset to construct the regional species pool did not
546 introduce a substantial bias in our analyses (see SI and Fig. S7). To test for the effect of changes in
547 physicochemical conditions on community assembly processes, Mantel tests (Spearman's rank
548 correlation, 10,000 permutations, function 'mantel' in vegan) were performed on the β -NTI matrix and
549 individual Euclidean distance matrices that were calculated for each physicochemical variable separately
550 after standardization.

551

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561

562 **References**

- 563
564 Anneser, B., Piloni, G., Bayer, A., Lueders, T., Griebler, C., Einsiedl, F., and Richters, L. (2010) High
565 Resolution Analysis of Contaminated Aquifer Sediments and Groundwater—What Can be Learned in
566 Terms of Natural Attenuation? *Geomicrobiology Journal* **27**: 130-142.
567
- 568 Baldani, I., Rouws, L., Cruz, L., Olivares, F., Schmid, M., and Hartmann, A. (2014) The Family
569 Oxalobacteraceae. In *The Prokaryotes, Alphaproteobacteria and Betaproteobacteria*. Rosenberg, E.,
570 DeLong, E.F., Lory, S., Stackebrandt, E., and Thompson, F. (eds). Heidelberg, Germany: Springer, pp. 919-
571 974.
572
- 573 Baselga, A. (2012) The relationship between species replacement, dissimilarity derived from nestedness,
574 and nestedness. *Global Ecology and Biogeography* **21**: 1223-1232.
575
- 576 Baselga, A., and Orme, C.D.L. (2012) betapart: an R package for the study of beta diversity. *Methods in*
577 *Ecology and Evolution* **3**: 808-812.
578
- 579 Battin, T.J., Besemer, K., Bengtsson, M.M., Romani, A.M., and Packmann, A.I. (2016) The ecology and
580 biogeochemistry of stream biofilms. *Nature Reviews Microbiology* **14**: 251-263.
581
- 582 Battin, T.J., Sloan, W.T., Kjelleberg, S., Daims, H., Head, I.M., Curtis, T.P., and Eberl, L. (2007) Microbial
583 landscapes: new paths to biofilm research. *Nature Reviews Microbiology* **5**: 76-81.
584
- 585 Bayer, A., Drexel, R., Weber, N., and Griebler, C. (2016) Quantification of aquatic sediment prokaryotes—
586 A multiple-steps optimization testing sands from pristine and contaminated aquifers. *Limnologica -*
587 *Ecology and Management of Inland Waters* **56**: 6-13.
588
- 589 Beaton, E.D., Stevenson, B.S., King-Sharp, K.J., Stamps, B.W., Nunn, H.S., and Stuart, M. (2016) Local and
590 Regional Diversity Reveals Dispersal Limitation and Drift as Drivers for Groundwater Bacterial
591 Communities from a Fractured Granite Formation. *Frontiers in Microbiology* **7**: 1933.
592
- 593 Besemer, K., Peter, H., Logue, J.B., Langenheder, S., Lindström, E.S., Tranvik, L.J., and Battin, T.J. (2012)
594 Unraveling assembly of stream biofilm communities. *The ISME Journal* **6**: 1459-1468.
595
- 596 Bižić-Ionescu, M., Zeder, M., Ionescu, D., Orlić, S., Fuchs, B.M., Grossart, H.-P., and Amann, R. (2014)
597 Comparison of bacterial communities on limnic versus coastal marine particles reveals profound
598 differences in colonization. *Environmental Microbiology* **17**: 3500-3514.
599
- 600 Brislawn, C.J., Graham, E.B., Dana, K., Ihardt, P., Fansler, S.J., Chirlser, W.B. et al. (2018) Forfeiting the
601 founder effect: turnover defines biofilm community succession. *bioRxiv pre-print; article version: March*
602 **15, 2018**.
603
- 604 Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K. et al. (2010) QIIME
605 allows analysis of high-throughput community sequencing data. *Nature Methods* **7**: 335-336.
606

- 607 Chase, J.M., and Myers, J.A. (2011) Disentangling the importance of ecological niches from stochastic
608 processes across scales. *Philosophical Transactions of the Royal Society B: Biological Sciences* **366**: 2351-
609 2363.
- 610
- 611 Chase, J.M., Kraft, N.J.B., Smith, K.G., Vellend, M., and Inouye, B.D. (2011) Using null models to
612 disentangle variation in community dissimilarity from variation in α -diversity. *Ecosphere* **2**: 1-11.
- 613
- 614 Cordero, O.X., and Datta, M.S. (2016) Microbial interactions and community assembly at microscales.
615 *Current Opinion in Microbiology* **31**: 227-234.
- 616
- 617 Danczak, R.E., Johnston, M.D., Kenah, C., Slattery, M., and Wilkins, M.J. (2018) Microbial Community
618 Cohesion Mediates Community Turnover in Unperturbed Aquifers. *mSystems* **3**: e00066-00018.
- 619
- 620 Dini-Andreote, F., Stegen, J.C., van Elsas, J.D., and Salles, J.F. (2015) Disentangling mechanisms that
621 mediate the balance between stochastic and deterministic processes in microbial succession.
622 *Proceedings of the National Academy of Sciences* **112**: E1326-E1332.
- 623
- 624 Dumbrell, A.J., Nelson, M., Helgason, T., Dytham, C., and Fitter, A.H. (2009) Relative roles of niche and
625 neutral processes in structuring a soil microbial community. *The Isme Journal* **4**: 337.
- 626
- 627 Fargione, J., Brown, C.S., and Tilman, D. (2003) Community assembly and invasion: An experimental test
628 of neutral versus niche processes. *Proceedings of the National Academy of Sciences* **100**: 8916-8920.
- 629
- 630 Faust, K., and Raes, J. (2012) Microbial interactions: from networks to models. *Nature Reviews*
631 *Microbiology* **10**: 538.
- 632
- 633 Ferrenberg, S., O'Neill, S.P., Knelman, J.E., Todd, B., Duggan, S., Bradley, D. et al. (2013) Changes in
634 assembly processes in soil bacterial communities following a wildfire disturbance. *The Isme Journal* **7**:
635 1102.
- 636
- 637 Flynn, T.M., Sanford, R.A., Ryu, H., Bethke, C.M., Levine, A.D., Ashbolt, N.J., and Santo Domingo, J.W.
638 (2013) Functional microbial diversity explains groundwater chemistry in a pristine aquifer. *BMC*
639 *Microbiology* **13**: 146.
- 640
- 641 Fukami, T. (2015) Historical Contingency in Community Assembly: Integrating Niches, Species Pools, and
642 Priority Effects. *Annual Review of Ecology, Evolution, and Systematics* **46**: 1-23.
- 643
- 644 Fukami, T., Dickie, I.A., Paula Wilkie, J., Paulus, B.C., Park, D., Roberts, A. et al. (2010) Assembly history
645 dictates ecosystem functioning: evidence from wood decomposer communities. *Ecology Letters* **13**: 675-
646 684.
- 647
- 648 Graham, B.E., and Stegen, C.J. (2017) Dispersal-Based Microbial Community Assembly Decreases
649 Biogeochemical Function. *Processes* **5**.
- 650
- 651 Graham, E.B., Crump, A.R., Resch, C.T., Fansler, S., Arntzen, E., Kennedy, D.W. et al. (2016a) Coupling
652 Spatiotemporal Community Assembly Processes to Changes in Microbial Metabolism. *Frontiers in*
653 *Microbiology* **7**: 1949.
- 654

- 655 Graham, E.B., Crump, A.R., Resch, C.T., Fansler, S., Arntzen, E., Kennedy, D.W. et al. (2017) Deterministic
656 influences exceed dispersal effects on hydrologically-connected microbiomes. *Environmental*
657 *Microbiology* **19**: 1552-1567.
- 658
- 659 Graham, E.B., Knelman, J.E., Schindlbacher, A., Siciliano, S., Breulmann, M., Yannarell, A. et al. (2016b)
660 Microbes as Engines of Ecosystem Function: When Does Community Structure Enhance Predictions of
661 Ecosystem Processes? *Frontiers in Microbiology* **7**: 214.
- 662
- 663 Griebler, C., and Lueders, T. (2009) Microbial biodiversity in groundwater ecosystems. *Freshwater*
664 *Biology* **54**: 649-677.
- 665
- 666 Griebler, C., and Avramov, M. (2015) Groundwater ecosystem services: a review. *Freshwater Science* **34**:
667 355-367.
- 668
- 669 Griebler, C., Malard, F., and Lefébure, T. (2014) Current developments in groundwater ecology — from
670 biodiversity to ecosystem function and services. *Current Opinion in Biotechnology* **27**: 159-167.
- 671
- 672 Griebler, C., Mindl, B., Slezak, D., and Geiger-Kaiser, M. (2002) Distribution patterns of attached and
673 suspended bacteria in pristine and contaminated shallow aquifers studied with an in situ sediment
674 exposure microcosm. *Aquatic Microbial Ecology* **28**: 117-129.
- 675
- 676 Grösbacher, M., Spicher, C., Bayer, A., Obst, M., Karwautz, C., Pilloni, G. et al. (2016) Organic
677 contamination versus mineral properties: competing selective forces shaping bacterial community
678 assembly in aquifer sediments. *Aquatic Microbial Ecology* **76**: 243-255.
- 679
- 680 Hanemaaijer, M., Röling, W.F.M., Olivier, B.G., Khandelwal, R.A., Teusink, B., and Bruggeman, F.J. (2015)
681 Systems modeling approaches for microbial community studies: from metagenomics to inference of the
682 community structure. *Frontiers in Microbiology* **6**: 213.
- 683
- 684 Hug, L.A., Thomas, B.C., Brown, C.T., Frischkorn, K.R., Williams, K.H., Tringe, S.G., and Banfield, J.F. (2015)
685 Aquifer environment selects for microbial species cohorts in sediment and groundwater. *The ISME*
686 *Journal* **9**: 1846–1856.
- 687
- 688 Jackson, C.R. (2003) Changes in community properties during microbial succession. *Oikos* **101**: 444-448.
- 689
- 690 Jones, A.A., and Bennett, P.C. (2017) Mineral Ecology: Surface Specific Colonization and Geochemical
691 Drivers of Biofilm Accumulation, Composition, and Phylogeny. *Frontiers in Microbiology* **8**: 491.
- 692
- 693 Kalmbach, S., Manz, W., Bendinger, B., and Szewzyk, U. (2000) In situ probing reveals *Aquabacterium*
694 *commune* as a widespread and highly abundant bacterial species in drinking water biofilms. *Water*
695 *Research* **34**: 575-581.
- 696
- 697 Kembel, S.W., Cowan, P.D., Helmus, M.R., Cornwell, W.K., Morlon, H., Ackerly, D.D. et al. (2010) Picante:
698 R tools for integrating phylogenies and ecology. *Bioinformatics* **26**: 1463-1464.
- 699
- 700 Konopka, A., Lindemann, S., and Fredrickson, J. (2015) Dynamics in microbial communities: unraveling
701 mechanisms to identify principles. *The ISME Journal* **9**: 1488-1495.
- 702

- 703 Korbek, K., Chariton, A., Stephenson, S., Greenfield, P., and Hose, G.C. (2017) Wells provide a distorted
704 view of life in the aquifer: implications for sampling, monitoring and assessment of groundwater
705 ecosystems. *Scientific Reports* **7**: 40702.
706
- 707 Langenheder, S., and Székely, A.J. (2011) Species sorting and neutral processes are both important
708 during the initial assembly of bacterial communities. *The Isme Journal* **5**: 1086–1094.
709
- 710 Langenheder, S., Wang, J., Karjalainen, S.M., Laamanen, T.M., Tolonen, K.T., Vilmi, A., and Heino, J.
711 (2017) Bacterial metacommunity organization in a highly connected aquatic system. *FEMS Microbiology*
712 *Ecology* **93**: fiw225-fiw225.
713
- 714 Lehman, R.M., Colwell, F.S., and Bala, G.A. (2001) Attached and Unattached Microbial Communities in a
715 Simulated Basalt Aquifer under Fracture- and Porous-Flow Conditions. *Applied and Environmental*
716 *Microbiology* **67**: 2799-2809.
717
- 718 Lyautey, E., Jackson, C.R., Cayrou, J., Rols, J.-L., and Garabétian, F. (2005) Bacterial Community
719 Succession in Natural River Biofilm Assemblages. *Microbial Ecology* **50**: 589-601.
720
- 721 Martiny, A.C., Jørgensen, T.M., Albrechtsen, H.-J., Arvin, E., and Molin, S. (2003) Long-Term Succession of
722 Structure and Diversity of a Biofilm Formed in a Model Drinking Water Distribution System. *Applied and*
723 *Environmental Microbiology* **69**: 6899-6907.
724
- 725 Martiny, J.B.H., Jones, S.E., Lennon, J.T., and Martiny, A.C. (2015) Microbiomes in light of traits: A
726 phylogenetic perspective. *Science* **350**: aac9323.
727
- 728 McMahon, S., and Parnell, J. (2013) Weighing the deep continental biosphere. *FEMS Microbiology*
729 *Ecology* **87**: 113-120.
730
- 731 Nemergut, D.R., Schmidt, S.K., Fukami, T., O'Neill, S.P., Bilinski, T.M., Stanish, L.F. et al. (2013) Patterns
732 and Processes of Microbial Community Assembly. *Microbiology and Molecular Biology Reviews : MMBR*
733 **77**: 342-356.
734
- 735 Niederdorfer, R., Peter, H., and Battin, T.J. (2016) Attached biofilms and suspended aggregates are
736 distinct microbial lifestyles emanating from differing hydraulics. *Nature Microbiology* **1**: 16178.
737
- 738 Niederdorfer, R., Besemer, K., Battin, T.J., and Peter, H. (2017) Ecological strategies and metabolic trade-
739 offs of complex environmental biofilms. *npj Biofilms and Microbiomes* **3**: 21.
740
- 741 Ofițeru, I.D., Lunn, M., Curtis, T.P., Wells, G.F., Criddle, C.S., Francis, C.A., and Sloan, W.T. (2010)
742 Combined niche and neutral effects in a microbial wastewater treatment community. *Proceedings of the*
743 *National Academy of Sciences* **107**: 15345-15350.
744
- 745 Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlenn, D. et al. (2018) vegan:
746 Community Ecology Package (R package version 2.5-2).
747
- 748 Peay, K.G., Belisle, M., and Fukami, T. (2012) Phylogenetic relatedness predicts priority effects in nectar
749 yeast communities. *Proceedings of the Royal Society B: Biological Sciences* **279**: 749-758.
750

- 751 Pilloni, G., von Netzer, F., Engel, M., and Lueders, T. (2011) Electron acceptor-dependent identification of
752 key anaerobic toluene degraders at a tar-oil-contaminated aquifer by Pyro-SIP. *FEMS Microbiology*
753 *Ecology* **78**: 165-175.
- 754
- 755 Price, M.N., Dehal, P.S., and Arkin, A.P. (2009) FastTree: Computing Large Minimum Evolution Trees with
756 Profiles instead of a Distance Matrix. *Molecular Biology and Evolution* **26**: 1641-1650.
- 757
- 758 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P. et al. (2013) The SILVA ribosomal RNA
759 gene database project: improved data processing and web-based tools. *Nucleic Acids Research* **41**: D590-
760 D596.
- 761
- 762 R Core Team (2018) R: A Language and Environment for Statistical Computing. In. Austria: R Foundation
763 for Statistical Computing.
- 764
- 765 Rummens, K., De Meester, L., and Souffreau, C. (2018) Inoculation history affects community
766 composition in experimental freshwater bacterioplankton communities. *Environmental Microbiology* **20**:
767 1120-1133.
- 768
- 769 Schmidt, S.I., Cuthbert, M.O., and Schwientek, M. (2017) Towards an integrated understanding of how
770 micro scale processes shape groundwater ecosystem functions. *Science of The Total Environment* **592**:
771 215-227.
- 772
- 773 Stegen, J.C., Lin, X., Konopka, A.E., and Fredrickson, J.K. (2012) Stochastic and deterministic assembly
774 processes in subsurface microbial communities. *The Isme Journal* **6**: 1653-1664.
- 775
- 776 Stegen, J.C., Lin, X., Fredrickson, J.K., and Konopka, A.E. (2015) Estimating and mapping ecological
777 processes influencing microbial community assembly. *Frontiers in Microbiology* **6**: 370.
- 778
- 779 Stegen, J.C., Lin, X., Fredrickson, J.K., Chen, X., Kennedy, D.W., Murray, C.J. et al. (2013) Quantifying
780 community assembly processes and identifying features that impose them. *The Isme Journal* **7**: 2069-
781 2079.
- 782
- 783 Stegen, J.C., Konopka, A., McKinley, J.P., Murray, C., Lin, X., Miller, M.D. et al. (2016a) Coupling among
784 Microbial Communities, Biogeochemistry, and Mineralogy across Biogeochemical Facies. *Scientific*
785 *Reports* **6**: 30553.
- 786
- 787 Stegen, J.C., Fredrickson, J.K., Wilkins, M.J., Konopka, A.E., Nelson, W.C., Arntzen, E.V. et al. (2016b)
788 Groundwater–surface water mixing shifts ecological assembly processes and stimulates organic carbon
789 turnover. *Nature Communications* **7**: 11237.
- 790
- 791 Stegen, J.C., Johnson, T., Fredrickson, J.K., Wilkins, M.J., Konopka, A.E., Nelson, W.C. et al. (2018)
792 Influences of organic carbon speciation on hyporheic corridor biogeochemistry and microbial ecology.
793 *Nature Communications* **9**: 585.
- 794
- 795 Svoboda, P., Lindström, E.S., Ahmed Osman, O., and Langenheder, S. (2018) Dispersal timing determines
796 the importance of priority effects in bacterial communities. *The Isme Journal* **12**: 644-646.
- 797
- 798 Tan, J., Pu, Z., Ryberg, W.A., and Jiang, L. (2012) Species phylogenetic relatedness, priority effects, and
799 ecosystem functioning. *Ecology* **93**: 1164-1172.

- 800
801 Tilman, D. (2004) Niche tradeoffs, neutrality, and community structure: A stochastic theory of resource
802 competition, invasion, and community assembly. *Proceedings of the National Academy of Sciences* **101**:
803 10854-10861.
804
805 Veach, A.M., Stegen, J.C., Brown, S.P., Dodds, W.K., and Jumpponen, A. (2016) Spatial and successional
806 dynamics of microbial biofilm communities in a grassland stream ecosystem. *Molecular Ecology* **25**:
807 4674-4688.
808
809 Wang, J., Shen, J., Wu, Y., Tu, C., Soininen, J., Stegen, J.C. et al. (2013) Phylogenetic beta diversity in
810 bacterial assemblages across ecosystems: deterministic versus stochastic processes. *The Isme Journal* **7**:
811 1310-1321.
812
813 Woodcock, S., and Sloan, W.T. (2017) Biofilm community succession: a neutral perspective. *Microbiology*
814 **163**: 664-668.
815
816 Woodcock, S., van der Gast, C.J., Bell, T., Lunn, M., Curtis, T.P., Head, I.M., and Sloan, W.T. (2007) Neutral
817 assembly of bacterial communities. *FEMS Microbiology Ecology* **62**: 171-180.
818
819 Zhou, J., and Ning, D. (2017) Stochastic Community Assembly: Does It Matter in Microbial Ecology?
820 *Microbiology and Molecular Biology Reviews* **81**: e00002-00017.
821
822 Zhou, J., Liu, W., Deng, Y., Jiang, Y.-H., Xue, K., He, Z. et al. (2013) Stochastic Assembly Leads to
823 Alternative Communities with Distinct Functions in a Bioreactor Microbial Community. *mBio* **4**: e00584-
824 00512.
825
826 Zhou, J., Deng, Y., Zhang, P., Xue, K., Liang, Y., Van Nostrand, J.D. et al. (2014) Stochasticity, succession,
827 and environmental perturbations in a fluidic ecosystem. *Proceedings of the National Academy of*
828 *Sciences* **111**: E836-E845.
829
830 Zhou, Y., Kellermann, C., and Griebler, C. (2012) Spatio-temporal patterns of microbial communities in a
831 hydrologically dynamic pristine aquifer. *FEMS Microbiology Ecology* **81**: 230-242.

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851 **Figure legends**

852 **Figure 1:** Schematic illustration of the main geographical features of the study site, the locations of the
853 two monitoring wells, and general groundwater flow directions.

854
855 **Figure 2:** Temporal dynamics of biomass and alpha diversity for sediment-attached communities and
856 planktonic communities in the groundwater at MIT052 and MIT039, respectively. Biomass is given as the
857 number of prokaryotic cells per mL groundwater or the equivalent volume (cm³) of sediment. The time in
858 days for each time point is given in parentheses.

859
860 **Figure 3:** (A) Changes over time in the fraction of newly-arriving OTUs (S_n) relative to the total number
861 of OTUs (S) in sediment-attached communities. (B) Changes over time in the cumulative relative
862 abundance of newly-arriving OTUs that first occurred at the time point indicated in the upper right corner
863 of each graph.

864
865 **Figure 4:** (A) Taxonomic microbial community composition of sediment-attached communities and
866 planktonic communities in the groundwater per time point. Taxonomic groups are summarized at family

867 level. Families with an abundance < 10% in the entire dataset were grouped as ‘Diverse others’ for clarity
868 of display (mean abundance: 0.1%; max: 4.7%). (B) NMDS plot showing differences in microbial
869 community composition based on abundance-weighted β -MNTD (stress: 0.09).

870
871 **Figure 5:** Values for β -NTI from pairwise microbial community comparisons. The range of β -NTI
872 indicating stochastic community turnover is shaded in grey. (A) Spatial community turnover of planktonic
873 communities in the groundwater and sediment-attached communities, respectively, between MIT052 and
874 MIT039 per time point. (B) Community turnover between planktonic and sediment-attached communities
875 within sites per time point (note: the bar corresponding to the comparison of communities at MIT052 in
876 July is not visible; β -NTI = 0.04). (C) Temporal community turnover of planktonic and sediment-attached
877 communities, respectively, between consecutive time points within sites (note: bar for the comparison of
878 July vs. December for planktonic communities at MIT052 is not visible; β -NTI = -0.005). NA: data not
879 available.

880
881 **Figure S1:** Changes in physicochemical parameters over time measured in the groundwater at the two
882 sites. DOC: dissolved organic carbon; AOC: assimilable organic carbon; DO: dissolved oxygen.

883
884 **Figure S2:** Occurrence frequencies of the most dominant families and most dominant single genera within
885 those families among newly-arriving OTUs (*Sn*) in the sediment-attached communities that first occurred
886 at the indicated time point and were still detected in the final communities at MIT052 and MIT039,
887 respectively. Families with an individual occurrence frequency < 3% are grouped as ‘Diverse others’ for
888 clarity of display. Bars representing each family are ordered by occurrence frequency in descending order
889 from top to bottom.

890
891 **Figure S3:** Differential abundances of genera that contributed most to the dissimilarity between sediment-
892 attached communities and planktonic communities in the groundwater identified by SIMPER analysis.

893 Only the genera with the highest significant contribution to the dissimilarity are shown ($> 0.1\%$; $p < 0.05$);
894 the average contribution of each displayed genus is indicated by the color intensity of the bars. (A) Log_{10} -
895 ratios of differential average abundances in planktonic communities over sediment-attached communities
896 for genera found in both community types. (B) Average relative abundances of genera exclusively found
897 in one community type.

898
899 **Figure S4:** Relative contributions of turnover and nestedness to the total Jaccard dissimilarity between
900 sediment-attached and planktonic communities at each site per time point inferred from beta diversity
901 partitioning.

902
903 **Figure S5:** Relative contributions of turnover and nestedness to the total Jaccard dissimilarity between
904 sediment-attached communities within sites across time points inferred from beta diversity partitioning.

905
906 **Figure S6:** Phylogenetic signal inferred from Mantel correlograms showing Pearson correlation between
907 phylogenetic distances and differences in environmental optima between OTUs within phylogenetic
908 distance classes evaluated at distance steps of 0.01 for (A) sediment-attached and (B) planktonic
909 communities. Filled symbols indicate significant correlations ($p < 0.05$).

910
911 **Figure S7:** Comparison of the outcomes of null model simulations to estimate (A) $\beta\text{-NTI}$ and (B) RC_{bray}
912 based on different regional species pools for pairwise community comparisons shown in Figure 5 and
913 discussed in the main text. The horizontal axes represent results based on regional species pools
914 constructed from OTUs found in the full data set; vertical axes show results based on regional species
915 pools constructed from OTUs found in subsets of samples within time points, or in case of comparisons to
916 estimate temporal community turnover, from two consecutive time points (only for $\beta\text{-NTI}$). Colors
917 represent the different investigated turnover processes shown in Figure 5 (see main text). Dashed lines
918 mark significance thresholds for each index (see main text). Linear regression slopes of the straight line \pm

919 0.95 confidence intervals and Pearson correlation coefficients are indicated in the figures. Flags indicate
920 pairwise comparisons for which the outcomes of the null models did not agree between the two strategies
921 for constructing regional species pools.

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923 **Tables**

924 **Table 1:** Mean values and standard deviation (SD) of physicochemical parameters averaged over the two investigated sites and all time points (n =
 925 10). Individual values for each site and time point are shown in Figure S1.

	Water table (m bs) ^a	DOC ^b (mg L ⁻¹)	AOC ^c (µg L ⁻¹)	pH	Temperature (°C)	Electrical conductivity (µS cm ⁻¹)	DO ^d (mg L ⁻¹)	PO ₄ ³⁻ (mg L ⁻¹)	SO ₄ ²⁻ (mg L ⁻¹)	NO ₃ ⁻ (mg L ⁻¹)	Cl ⁻ (mg L ⁻¹)	K ⁺ (mg L ⁻¹)	Na ⁺ (mg L ⁻¹)	Ca ²⁺ (mg L ⁻¹)	Mg ²⁺ (mg L ⁻¹)
Mean	8.9	1.10	11.4	7.81	6.88	297	10.21	0.011	14.5	2.49	1.80	0.30	1.03	46.1	12.1
SD	5.0	0.37	14.7	0.10	0.59	41	0.82	0.015	8.8	0.78	1.13	0.07	0.62	5.5	1.0

926 ^a meter below surface927 ^b dissolved organic carbon928 ^c assimilable organic carbon929 ^d dissolved oxygen

Non-random processes determine the colonization of groundwater sediments by microbial communities in a pristine porous aquifer

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Running title: Microbial colonization of groundwater sediments.

Keywords: Community assembly, succession, selection, surface colonization, groundwater, sediments.

Originality-significance statement

Groundwater bodies are the largest terrestrial habitat for microorganisms on Earth, where the majority of the microbial biomass lives attached to sediment surfaces. In these unique, low-productivity environments, microbial communities are the drivers of key biogeochemical processes and furthermore provide important services to society like maintaining groundwater quality as one of the most important sources of freshwater and drinking water worldwide. Over the past years, studies on sediment-attached communities in groundwater-surface water mixing zones as well as surface-attached biofilms in other, non-subsurface habitats have provided important insights regarding the ecological processes that drive the assembly of these microbial communities. Compared to most of these environments, however, pristine groundwater is characterized by significantly lower levels of energy and productivity as well as comparatively more stable environmental conditions, which may promote the effect of stochastic processes on community assembly. Moreover, the microbial communities that colonize subsurface sediments typically exhibit much lower cell densities and occur as small, spatially separated micro-colonies rather than dense, coherent biofilms as they are found in other non-subsurface environments. Therefore, our study was motivated by the question whether findings on the processes that govern microbial community assembly and succession of surface-attached communities in those other more dynamic and nutrient-rich environments also apply to sediment-attached microbial communities in pristine groundwater environments. Our study shows intriguing similarities between the community succession on newly-colonized sediments in our investigated porous, pristine aquifer and succession patterns observed for biofilms in other more dynamic aquatic environments, indicating that the assembly of microbial communities on surfaces may be governed by similar underlying mechanisms across a wide range of different habitats. Our results indicate that differences between planktonic and sediment-attached communities often reported for groundwater environments are not the result of purely stochastic events, but that sediment surfaces select for specific groups of microorganisms that assemble over time in a reproducible, non-random way. Furthermore, our data suggest that specific genera, especially within the *Comamonadaceae* and *Oxalobacteraceae*, played a particularly important role in this process.

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1 **Non-random processes determine the colonization of groundwater sediments**
2 **by microbial communities in a pristine porous aquifer**

3

4 Lucas Fillinger^{a, 1}, Yuxiang Zhou^{a, 1}, Claudia Kellermann^a, Christian Griebler^{a*}

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8 **Summary**

9 Sediments accommodate the dominating share of groundwater microbiomes, however the processes that
10 govern the assembly and succession of sediment-attached microbial communities in groundwater aquifers
11 are not well understood. To elucidate these processes, we followed the microbial colonization of sterile
12 sediments in *in situ* microcosms that were exposed to groundwater for almost one year at two distant but
13 hydrologically connected sites of a pristine, shallow, porous aquifer. Our results revealed intriguing
14 similarities between the community succession on the newly-colonized sediments and succession patterns
15 previously observed for biofilms in other more dynamic aquatic environments, indicating that the
16 assembly of microbial communities on surfaces may be governed by similar underlying mechanisms
17 across a wide range of different habitats. Null model simulations on spatiotemporally resolved 16S rRNA
18 amplicon sequencing data further indicated selection of specific OTUs rather than random colonization as
19 the main driver of community assembly. A small fraction of persistent OTUs that had established on the
20 sediments during the first 115 days dominated the final communities (68%-85%), suggesting a key role of
21 these early-colonizing organisms, in particular specific genera within the *Comamonadaceae* and
22 *Oxalobacteraceae*, for community assembly and succession during the colonization of the sediments.
23 Overall, our study suggests that differences between planktonic and sediment-attached communities often
24 reported for groundwater environments are not the result of purely stochastic events, but that sediment

25 surfaces select for specific groups of microorganisms that assemble over time in a reproducible, non-
26 random way.

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28 **Introduction**

29 The groundwater-saturated zones of the terrestrial subsurface are one of the largest habitats for
30 microorganisms on Earth (Griebler and Lueders, 2009; McMahon and Parnell, 2013). In these unique,
31 low-productivity environments, microbial communities lie at the heart of key biogeochemical processes
32 like the turnover of carbon and other nutrients, mineral cycling, or pollutant degradation (Griebler et al.,
33 2014; Griebler and Avramov, 2015). Sediment-attached communities play a particularly important role in
34 these ecosystems as they represent the bulk of the microbial biomass and activity (Lehman et al., 2001;
35 Griebler et al., 2002; Zhou et al., 2012; McMahon and Parnell, 2013). Previous studies have repeatedly
36 shown that the composition of sediment-attached communities can differ substantially from planktonic
37 communities suspended in the surrounding groundwater (Zhou et al., 2012; Flynn et al., 2013; Hug et al.,
38 2015). However, the ecological processes that give rise to these differences during community assembly
39 and succession are not well understood. Recent studies have suggested a strong link between
40 biogeochemical functions and microbial community composition as well as the underlying ecological
41 assembly processes (Graham et al., 2016a; Graham et al., 2016b; Graham and Stegen, 2017). Therefore, a
42 better understanding of the processes that drive the assembly of sediment-attached microbial communities
43 in groundwater environments is a key step towards a better understanding of the functioning of these
44 ecosystems.

45 The question of the influence of deterministic (or niche-based) versus stochastic (or neutral)
46 processes on the assembly, succession, and diversity of microbial communities has increasingly sparked
47 the curiosity of microbial ecologists over the past years (for reviews see Nemergut et al. (2013); Zhou and
48 Ning (2017)). Deterministic theory assumes that environmental factors, both biotic and abiotic, determine
49 the composition and structure of a community by selecting for species with certain traits that enable them
50 to thrive and compete in a given environment (also known as species sorting or environmental filtering).
51 Accordingly, under similar environmental conditions, communities at different locations or points in time
52 are expected to be composed of species with similar traits. In contrast, the stochastic point of view holds

53 that communities are assembled randomly from species with equivalent trait spectra, and that differences
54 in community composition and structure are the result of random events, for example caused by species
55 dispersal or drift due to stochastic birth-death events (Chase and Myers, 2011). Although purely neutral
56 models have been shown to explain observed diversity patterns of microbial communities with surprising
57 accuracy (e.g. Woodcock et al. (2007); Ofițeru et al. (2010); Woodcock and Sloan (2017)), there has been
58 growing consent that both deterministic and stochastic processes can act simultaneously and that the
59 contribution of either process can shift over time or with changing environmental conditions (Dumbrell et
60 al., 2009; Ofițeru et al., 2010; Chase and Myers, 2011; Langenheder and Székely, 2011; Ferrenberg et al.,
61 2013; Stegen et al., 2013; Wang et al., 2013; Zhou et al., 2014; Dini-Andreote et al., 2015; Stegen et al.,
62 2015; Stegen et al., 2016b; Veach et al., 2016; Graham et al., 2017).

63 One aspect where the interaction between deterministic and stochastic processes comes into play
64 is the initial assembly of communities and the following species succession in newly-colonized
65 environments (Tilman, 2004; Langenheder and Székely, 2011), for instance during the development of
66 biofilms on initially empty surfaces (Jackson, 2003; Battin et al., 2007). At the initial stage of
67 colonization, the arrival of species in a new environment is often driven by stochastic dispersal (Tilman,
68 2004; Ferrenberg et al., 2013; Dini-Andreote et al., 2015), which can overrule deterministic effects in
69 homogeneous environments or if environmental filtering between the environment of the source
70 community and the newly colonized environment is weak (Stegen et al., 2012; Wang et al., 2013; Battin et
71 al., 2016). However, once established, resident species can affect the establishment of newly-arriving
72 species (positively or negatively) during the subsequent succession directly via species interactions or
73 indirectly by modification of their environment (Fukami, 2015). Thus the order and timing of species
74 arrival, although initially stochastic, can determine the composition and functioning of the final
75 community, known as priority effect (Fargione et al., 2003; Tilman, 2004; Fukami et al., 2010; Peay et al.,
76 2012; Tan et al., 2012; Nemergut et al., 2013; Rummens et al., 2018; Svoboda et al., 2018).

77 A general, conceptual model that summarizes the successional stages during biofilm development
78 has been described by Jackson (2003). According to this model, initially empty surfaces offer ample space

79 and resources to allow for the establishment of diverse microorganisms resulting in a rapid increase in
80 species richness and diversity that is fueled by the dispersal of newly-arriving species from a regional
81 species pool like overlaying water during initial biofilm assembly. The steady arrival of new species
82 eventually leads to niche depletion and growing competition between established and newly-arriving
83 species, which more and more suppresses the increase in species richness. As the competition intensifies,
84 less competitive species are lost from the community, which leads to a decline in species richness after the
85 initial phase of community assembly. However, as the biofilm matures further and becomes more
86 heterogeneous, new niches are created that enable specialized species to establish, leading again to an
87 increase in richness and diversity in the mature biofilm. Although Woodcock and Sloan (2017) have
88 demonstrated using a neutral modeling approach that these patterns can be explained based on stochastic
89 processes only, empirical evidence suggests that the assembly of biofilm communities is in fact
90 characterized by a shift from initially stochastic community assembly towards deterministically driven
91 succession at the later stages, for instance caused by species interactions or growing niche space due to
92 increasing spatial and chemical heterogeneity (Martiny et al., 2003; Lyautey et al., 2005; Battin et al.,
93 2016; Veach et al., 2016; Brislawn et al., 2018).

94 To date, most of the studies on ecological processes behind the assembly of microbial
95 communities in groundwater environments have focused on planktonic communities suspended in the
96 groundwater (Stegen et al., 2012; Stegen et al., 2013; Beaton et al., 2016; Danczak et al., 2018), while
97 studies on sediment-attached communities are scarce (Stegen et al., 2016a). In contrast, much insight has
98 been gained over the past years into the assembly of sediment-attached communities in groundwater-
99 surface water mixing zones (hyporheic zone). In these studies, the assembly of planktonic communities
100 generally tended to be more subject to stochastic effects and shifts in assembly processes related to
101 changes in water chemistry, whereas selection had a relatively more pronounced effect on the assembly of
102 sediment-attached communities which, at the same time, was less affected by hydrochemical changes
103 (Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et al., 2018). Compared to the
104 hyporheic zone, pristine groundwater environments (in the absence of surface water impacts) are more

105 stable and only experience little environmental changes (Griebler and Lueders, 2009), which may promote
106 the effect of stochastic processes on community assembly compared to more dynamic environments
107 (Ofițeru et al., 2010; Stegen et al., 2012; Wang et al., 2013; Zhou et al., 2013). Moreover, in contrast to
108 the typically studied biofilms in other environments like surface waters, which form dense, spatially
109 coherent, heterogeneous structures that can reach a thickness in the range of several hundred micrometers
110 (Battin et al., 2016), sediment-attached microbial communities in groundwater aquifers occur as small,
111 patchily distributed micro-colonies that consist of only a few cells (Schmidt et al., 2017), which may be
112 hypothesized to be more prone to stochastic effects than their biofilm counterparts in other environments.

113 In this study we set out to 1) investigate whether the assembly of sediment-attached microbial
114 communities in pristine groundwater environments can be explained by the general patterns observed for
115 surface-attached biofilms in other environments; 2) study the importance of early colonizers for
116 community succession during the colonization of the sediments; and 3) see if the dominating role of
117 selection on community assembly that has been observed for sediment-attached microbial communities in
118 the hyporheic zone also drives community assembly in comparatively stable, pristine groundwater
119 environments. To tackle these goals, we incubated *in situ* microcosms filled with sterilized sediment in
120 monitoring wells at two distant but hydrologically connected sites of a pristine, porous aquifer (Zhou et
121 al., 2012), and followed the succession of the microbial communities as they colonized the sediments over
122 a period of almost one year. We used 16S rRNA amplicon sequencing data to study changes in alpha and
123 beta diversity of the sediment-attached communities incubated at the two sites as well as differences in
124 community composition between sediment-attached and planktonic communities in the surrounding
125 groundwater over the course of the succession. To explore the influence of deterministic and stochastic
126 processes on microbial community assembly and succession, we applied the null model approach
127 developed by Stegen et al. (2012; 2013), which has previously also been used in studies on community
128 assembly in the hyporheic zone (Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et
129 al., 2018) as well as biofilms in other environments (Langenheder et al., 2017; Brislawn et al., 2018), and
130 thus allows us to compare our results to those previous findings.

131

132 **Results**

133 **Site description**

134 The field experiment was conducted over a period of 347 days from March 2010 until February 2011,
135 with intermediate sampling campaigns in May (day 49), July (day 115), and December (day 263). The
136 area with the two monitoring wells used for the incubation of the *in situ* microcosms was located at the
137 foothills of the Bavarian Alps in the upper Isar River valley close to the village of Mittenwald, Germany
138 (Fig. 1). The wells were installed in a pristine, shallow, porous aquifer composed of quaternary sediment
139 mainly consisting of gravel and coarse sand. Well MIT052 was located on a mountain pasture in the
140 forested Riedboden nature reserve 400 m away from the nearby river; well MIT039 was located
141 approximately 2 km away from MIT052 in proximity to the village with a distance of 240 m to the river
142 (for a detailed site description, see Zhou et al. (2012)). Over the course of the experiment, we observed
143 only small fluctuations in physicochemical conditions (Table 1; the temporal dynamics of the individual
144 parameters are shown in Fig. S1).

145

146 **Temporal dynamics of microbial biomass and alpha diversity of sediment-attached and planktonic** 147 **microbial communities**

148 Despite the spatial distance between the two sites, the microbial communities that developed on the
149 initially sterile sediments followed identical trends in alpha diversity and biomass patterns (Fig. 2).
150 Already after the first 49 days, the microbial biomass (measured as prokaryotic cell counts) of attached
151 microbial communities at both sites had reached a plateau of $\sim 10^7$ cells cm^{-3} of sediment followed by a
152 slight decline for the remaining time of the experiment. Although the biomass of sediment-attached
153 microbial communities stayed more or less constant, noticeable changes in the communities still occurred
154 as indicated by OTU richness and diversity which steadily increased by about 50% and 25%, respectively,
155 from May until December, followed by a decline of both parameters in February. Over the same period,
156 community evenness remained relatively high and only changed moderately. The changes in biomass and

157 alpha diversity observed for the newly colonized sediments appeared to be independent from the changes
158 that occurred in the planktonic microbial communities, which were more variable over time and less
159 comparable between the two sites. Microbial biomass was about one to two orders of magnitude lower for
160 the planktonic communities compared to the sediments.

161

162 **Establishment and persistence of newly-arriving OTUs in sediment-attached microbial communities**

163 To assess the impact of early colonizers on microbial community succession, we looked at the number of
164 newly-arriving OTUs that entered the developing sediment-attached communities at each time point over
165 the course of the experiment (Fig. 3). Newly-arriving OTUs are defined here as OTUs that showed an
166 abundance $> 0\%$ in the community for the first time at a given time point. At both sites, the number of
167 newly-arriving OTUs declined over time showing that the majority of OTUs had established during the
168 initial phase of the incubation. Despite this declining trend, the fraction of newly-arriving OTUs relative to
169 the total OTU richness at the end of the incubation was still noticeable with $\sim 15\text{-}20\%$. However, looking
170 at the changes in the cumulative relative abundances of the newly-arriving OTUs over time, we saw that
171 the OTUs that had arrived towards the later stages only accounted for a relatively small fraction of the
172 final communities. Even though the cumulative relative abundance of OTUs that had established in the
173 communities within the first 49 days steadily declined, these OTUs still made up 36% and 47% of the
174 final communities at MIT052 and MIT039, respectively. At MIT052, these OTUs together with those that
175 emerged at the following time point in July comprised the majority of the final community at the end of
176 the incubation (together 85%), while OTUs that arrived at the final time point accounted for only 5%. At
177 MIT039, OTUs that had arrived at the first two time points made up for 68% of the final community,
178 whereas OTUs that had arrived in December and February comprised 12% and 20%, respectively.
179 Although these results clearly show the dominance of early-colonizer OTUs in the final communities, a
180 closer look at how many of these OTUs actually persisted until the final time point showed that only
181 $\sim 12\%$ of newly-arriving OTUs from each time point were still present in the final communities (data not
182 shown). Looking at the taxonomies of these newly-arriving OTUs that persisted until the end of the

183 incubation, we again found highly similar patterns for both sites (Fig. S2). At each time point, the most
184 dominant groups were OTUs affiliated to *Comamonadaceae*, mainly *Aquabacterium* and *Polaromonas*
185 spp., in addition to *Oxalobacteraceae*, mainly consisting of *Duganella*, *Massilia*, and *Undibacterium* spp.,
186 as well as *Pseudomonas* spp. and diverse *Caulobacteraceae* and *Sphingomonadaceae*.

187

188 **Microbial community composition and beta diversity**

189 Similar to the alpha diversity patterns, the microbial communities on the newly colonized sediments
190 displayed comparable compositions at the two sites (Fig. 4). Especially during the initial phase of the
191 incubation in May, sediment-attached communities at both sites were dominated by *Oxalobacteraceae* in
192 addition to *Comamonadaceae* and smaller fractions of *Flavobacteriaceae* and *Caulobacteraceae*. Over the
193 further course of the incubation, these taxonomic groups gradually receded and were in part replaced
194 mainly by increasing numbers of *Comamonadaceae* (mainly *Aquabacterium* spp.), *Pseudomonadaceae*,
195 *Nocardiaceae*, and *Rhodocyclaceae* especially at MIT052, in addition to *Sphingomonadaceae*, uncultured
196 *Deltaproteobacteria*, and *Moraxellaceae* at MIT039. Moreover, OTUs affiliated with diverse low-
197 abundant families (with an abundance <10% in the entire dataset; mean = 0.1%; max. = 4.7%) gradually
198 increased in abundance. In contrast, planktonic communities were mainly dominated by members of the
199 *Rhodocyclaceae*, *Comamonadaceae* (mainly *Curvibacter*, *Simplicispira*, and *Rhodoferax* spp.), and
200 *Leptospiraceae*.

201 To get a better picture of the organisms that were responsible for differences between sediment-
202 attached and planktonic communities, we performed similarity percentage (SIMPER) analysis across all
203 samples on relative abundances of OTUs grouped at genus level. Interestingly, we found high agreement
204 between the genera that significantly contributed to the observed dissimilarities between the two types of
205 communities and the taxa identified as successful, persistent colonizers in the previous analysis (Fig. S2).
206 *Aquabacterium*, *Massilia*, and *Duganella* spp. ranked among the genera that contributed most to the
207 dissimilarity (together > 15%; all $p < 0.002$) and were highly differentially abundant in the sediment-
208 attached communities, next to *Flavobacteria* and uncultured members of the *Oxalobacteraceae* (Fig. S3).

209 The changes in microbial community composition over time as well as differences between
210 sediment-attached and planktonic communities in the groundwater were revealed by non-metric
211 multidimensional scaling (NMDS) performed on abundance-weighted β -mean nearest taxon distance (β -
212 MNTD) between communities (Fig. 4). At all measured time points, sediment-attached and planktonic
213 communities clustered separately from each other as reflected by the distinct separation of the two types
214 of communities along the first NMDS axis. Changes in microbial community composition over time were
215 reflected by the separation of data points along the second NMDS axis. Permutational analysis of variance
216 (PERMANOVA) revealed that community type (i.e. sediment-attached vs. planktonic) explained most of
217 the variance in β -MNTD between communities ($R^2 = 0.626$; $p = 0.001$), followed by sampling time point
218 ($R^2 = 0.104$; $p = 0.001$), while site location was not significant ($R^2 = 0$; $p = 1$), showing that communities
219 across sites were similar within each community type and time point. Moreover, fitting environmental
220 variables to the NMDS ordination with stratification of permutations within the community types did not
221 reveal significant correlations between changes in community composition and any of the measured
222 physicochemical parameters (Table 1 and Figure S1) (all $R^2 < 0.32$; $p > 0.1$).

223 Since community type explained most of the variance in beta diversity, we applied partitioning of
224 beta diversity to identify the underlying causes of the differences between sediment-attached and
225 planktonic communities within sites and sampling time points according to Baselga (2012). This approach
226 is based on the additive partitioning of incidence-based Jaccard dissimilarity between two communities
227 into a nestedness and a turnover component. A high contribution of nestedness to the total dissimilarity
228 indicates that two communities are subsets of each other and that differences are caused by differences in
229 species richness (i.e. gain or loss of species). On the other hand, a high contribution of turnover indicates
230 little overlap in species composition, i.e. species in one community have been replaced by other species in
231 the other community. The analysis showed that turnover was the dominating process behind the
232 differences between the two types of communities at each time point (for all time points $> 97\%$) (Fig. S4),
233 showing that both community types were composed of distinct sets of OTUs.

234 We applied the same approach to the dissimilarity between sediment-attached communities at
235 different time points within sites to investigate the degree to which nestedness and turnover contributed to
236 changes in community composition over the course of the sediment colonization. Also in this case
237 turnover dominated over nestedness in all comparisons, especially over long time scales (i.e. comparing
238 communities between May and February; 95-97%), and with a slightly weaker effect over short time
239 scales of the succession (i.e. comparing communities between consecutive time points; 74-90%) (Fig. S5).

240
241 **Impact of stochastic and deterministic processes on community assembly and succession inferred**
242 **from null models**

243 To infer the impact of deterministic and stochastic processes on community assembly, we applied the two-
244 tiered null model approach developed by Stegen et al. (2012; 2013). Briefly, under the assumption that
245 phylogenetically closely related species occupy more similar ecological niches than less closely related
246 species, the strength of deterministic processes is evaluated in tier one based on the β -nearest taxon index
247 (β -NTI). β -NTI < -2 and $> +2$ indicates that two communities are phylogenetically significantly more or
248 less similar to each other than expected by chance, which is interpreted as homogeneous selection (i.e.
249 selection of similar OTUs) or variable selection (i.e. selection of dissimilar OTUs) in the two
250 communities, respectively. $|\beta$ -NTI < 2 indicates that two communities are as dissimilar as expected by
251 chance, hinting at stochastic community assembly. In this case, the RC_{bray} index is used in tier two to
252 evaluate the effect of stochastic dispersal. $RC_{\text{bray}} < -0.95$ and $> +0.95$ means that two communities share
253 significantly more or less OTUs than expected by chance, indicating that communities are assembled by
254 homogenizing dispersal or dispersal limitation in combination with drift, respectively. $|RC_{\text{bray}}| < 0.95$
255 indicates that differences between two communities are due to random drift acting alone. We applied this
256 approach to study the role of deterministic and stochastic processes on three levels (Fig. 5): 1) spatial
257 community turnover between the two sites within community types and time points; 2) turnover between
258 sediment-attached and planktonic communities within sites and time points; 3) temporal turnover between
259 communities at consecutive time points within community types and sites.

260 Deterministic and stochastic processes had different influences on the spatial community turnover
261 for sediment-attached and planktonic communities, respectively. In case of planktonic communities,
262 pairwise comparisons between sites for each time point resulted in values for β -NTI that were not
263 significantly different from the null expectation, indicating that differences between the planktonic
264 communities at the two sites were caused by stochastic processes. Pairwise comparisons based on the
265 RC_{bray} index identified homogenizing dispersal as the dominating process (all $RC_{\text{bray}} = -1$; the only
266 exception was observed for December: $RC_{\text{bray}} = +0.99$, indicating dispersal limitation together with drift;
267 data not shown). In contrast to the predominantly stochastic exchange of OTUs between the two sites via
268 dispersal through the groundwater, pairwise comparisons of sediment-attached communities clearly
269 tended towards homogenous selection as main cause of the similarities between the sediment-attached
270 communities (with one exception observed for July where β -NTI was not significant, but a slightly
271 significant $RC_{\text{bray}} = 0.97$ hinted at dispersal limitation together with drift).

272 Since the sediments had to be colonized by microorganisms that were recruited from the
273 surrounding groundwater, even though beta diversity partitioning revealed little overlap between these two
274 communities as discussed above, we tested for the effect of selection processes on the assembly of
275 sediment-attached communities from planktonic microorganisms suspended in the groundwater. The
276 differences between the communities on the newly colonized sediments and the planktonic communities at
277 each site were caused by stochastic process during the first 115 days of the incubation. Pairwise
278 comparisons based on RC_{bray} hinted at dispersal limitation in combination with drift as the processes
279 responsible for these differences (all $RC_{\text{bray}} = +1$; data not shown). This trend changed at the later stage in
280 December after 263 days; at this point significantly positive values for β -NTI hinted at variable selection
281 of phylogenetically distinct OTUs in sediment-attached communities compared to the microorganisms in
282 the surrounding groundwater.

283 Unlike the trends observed for the spatial community turnover, the influence of deterministic and
284 stochastic effects on changes in community composition that occurred over time was much more variable
285 and no clear trends could be observed. Although selection effects appeared to have played a role (both

286 homogenous and variable selection), they mostly did not occur consistently at both sites for neither
287 sediment-attached nor planktonic communities.

288 We used Mantel correlation analysis to investigate whether changes in individual physicochemical
289 parameters in the groundwater had an effect on the changes in assembly processes (based on β -NTI).
290 Similar to the lack of correlations between environmental variables and differences in community
291 composition mentioned above, we did not find significant effects of changes in environmental conditions
292 in this analysis for neither planktonic (all |Spearman's rho| < 0.34; $p > 0.08$) nor sediment-attached
293 communities (all |Spearman's rho| < 0.27; $p > 0.1$).

294

295 **Discussion**

296 The alpha diversity patterns for the newly-colonized sediments at both sites followed identical trends that
297 closely matched the conceptual model for the formation of biofilms on empty surfaces outlined by Jackson
298 (2003), which describes changes in alpha diversity over three main stages of biofilm development. At the
299 early stage, the large niche space of an initially empty surface allows for the establishment of diverse
300 microorganisms, resulting in a steady increase in alpha diversity, which subsequently levels off and
301 eventually declines due to niche depletion and the loss of less competitive species as the biofilm grows
302 over the course of the succession. However, at the final stage, the mature biofilm becomes increasingly
303 spatially and chemically heterogeneous, which opens new niches for specialized species to thrive and
304 thereby fuels a renewed increase in alpha diversity. Our results only deviated from this model at the final
305 stage of the incubation, where we did not see an increase in species diversity and richness. However, this
306 framework was conceptualized for biofilms in resource-rich, high-productivity environments like
307 activated sludge, wetlands, and lakes (Jackson, 2003). Although we cannot exclude that alpha diversity
308 may have increased again with a prolonged incubation time, we may argue that diverse, specialized niches
309 that develop in mature, spatially heterogeneous biofilms might not form to such an extent in the small,
310 patchily distributed micro-colonies that typically colonize groundwater sediments (Schmidt et al., 2017).
311 Hence, the total niche space in such micro-colonies may be smaller compared to mature biofilms in other

312 environments, similar to what Graham et al. (2016a) have proposed for sediments in the hyporheic zone.
313 Moreover, although the general pattern of decreasing fractions of newly-arriving OTUs was also apparent
314 in our experiment, reflecting the saturation of niche space according to Jackson's biofilm model (Jackson,
315 2003), we noticed that the fraction of these OTUs at the end of the incubation was still 5-10 times higher
316 compared to findings on biofilms in other environments (e.g. Brislawn et al. (2018)). These deviations of
317 our results from assembly patterns of biofilms, together with the findings made for hyporheic zone
318 sediments (Graham et al., 2016a), might point towards important differences in ecological niche structures
319 between biofilms in resource-rich surface environments and sediment-attached microbial communities in
320 the typically more energy-poor and less productive subsurface.

321 Looking at the abundance changes of newly-arriving OTUs over time, we saw that OTUs
322 colonizing the sediments during the early stage of community assembly (i.e. the first 49 to 115 days)
323 largely dominated the final communities at the end of the experiment. However, at the same time, these
324 dominant OTUs represented only a small fraction of newly-arriving OTUs found at each time point. This
325 was further reflected by the large dominance of OTU turnover over nestedness between successional
326 stages in the sediment-attached communities inferred from beta diversity partitioning, showing that the
327 majority of OTUs that had established at a given time point were in fact replaced by others over the course
328 of the succession. Therefore, in agreement with the findings by Brislawn et al. (2018), the mere timing of
329 OTU arrival did not seem to be a determining factor for the final community structure. Rather, the
330 consistent dominance of specific taxa among these persistent OTUs (mainly genera belonging to the
331 *Oxalobacteraceae*, *Comamonadaceae*, *Caulobacteraceae*, *Sphingomonadaceae*, in addition to
332 *Pseudomonas* spp.) suggests the involvement of certain traits that enable these taxa to sustainably colonize
333 and thrive on sediment surfaces. Interestingly, we also found the same genera among the most important
334 contributors to differences between sediment-attached and planktonic communities and to be highly
335 differentially abundant in the former. The association of these taxa with biofilms and traits that facilitate
336 surface colonization like motility or production of extracellular polysaccharides have been reported before
337 for other environments (Kalmbach et al., 2000; Baldani et al., 2014; Bižić-Ionescu et al., 2014;

338 Niederdorfer et al., 2016; Niederdorfer et al., 2017), supporting the hypothesis about their importance for
339 the development of sediment-attached communities in our study. Over the course of the succession, these
340 dominant OTUs may have facilitated the recruitment of other more diverse taxa that were observed at the
341 later stages of the colonization (Battin et al., 2007; Nemergut et al., 2013; Fukami, 2015).

342 Comparisons of beta diversity patterns revealed that sediment-attached and planktonic
343 communities, respectively, were similar at each time point across the two sampling locations. Using the
344 null model approach developed by Stegen et al. (2012; 2013) revealed that different processes were
345 responsible for the observed similarities. Whereas the spatial turnover of planktonic microbial
346 communities was driven by stochastic processes, mostly homogenizing dispersal (75%), the high
347 similarities between the sediment-attached communities at the two sites were mostly caused by
348 homogenous selection (75%). We are aware that our study consists of only a relatively limited number of
349 observations and therefore the results should be interpreted with the necessary caution. Nevertheless, our
350 results fit observations on assembly processes for communities in the hyporheic zone (Graham et al.,
351 2016a; Stegen et al., 2016b; Graham et al., 2017; Stegen et al., 2018) as well as biofilms in surface water
352 streams (Besemer et al., 2012; Veach et al., 2016), suggesting that selection not only plays a determining
353 role in the assembly of surface-attached microbial communities in those dynamic environments but also in
354 pristine groundwater aquifers, despite the comparatively more stable environmental conditions, which
355 have been shown to promote the effect of stochastic over deterministic processes in other environments
356 (Ofițeru et al., 2010; Stegen et al., 2012; Wang et al., 2013; Zhou et al., 2013). Mineral composition has
357 previously been demonstrated to be a driving factor for microbial community composition and assembly
358 (Grösbacher et al., 2016; Stegen et al., 2016a; Jones and Bennett, 2017). Since the *in situ* microcosms that
359 we incubated at the two sites in our study were filled with sediment that originated from the same source,
360 it is likely that identical sediment properties selected for the highly similar microbial communities at the
361 two sites.

362 Given the high similarities between the sediment-attached communities at both sites throughout
363 the experiment, we would have expected to also find similar patterns regarding the processes that drove

364 the temporal microbial community turnover. However, contrary to this expectation, this was not fully the
365 case as assembly was highly variable without a clearly discernable trend in favor of a single process.
366 Changes in environmental conditions such as nutrient inputs, fluctuating water tables, or surface water-
367 groundwater mixing have been observed to not only affect the composition of (groundwater) microbial
368 communities, but also influence the ecological assembly processes that determine those changes (Lyautey
369 et al., 2005; Stegen et al., 2012; Stegen et al., 2013; Zhou et al., 2014; Dini-Andreote et al., 2015; Stegen
370 et al., 2015; Graham et al., 2016a; Stegen et al., 2016b; Graham et al., 2017). However, in our case, we did
371 not find indications that changes in physicochemical conditions of the groundwater were related to
372 changes in community composition or shifts in ecological community assembly processes. This could
373 suggest that the changes in community composition over time and the influence of deterministic versus
374 stochastic effects were determined by changes in unmeasured environmental variables (Stegen et al.,
375 2013). Alternatively, the observed lack of correlations between changes in environmental conditions and
376 the processes that determined community assembly can also hint at the impact of endogenous factors like
377 species interactions (Konopka et al., 2015; Battin et al., 2016; Cordero and Datta, 2016). Recently,
378 Danczak et al. (2018) could show that interaction network structures can affect assembly processes of
379 planktonic microbial communities in pristine aquifers. Although our results show that the assembly of
380 sediment-attached communities was mainly deterministic, and that the succession of OTUs was highly
381 reproducible between the two sites, the compositions of the two communities at each time point, and
382 therefore possibly interaction networks, were not totally identical. Hence, the variable patterns of
383 processes that determined the community turnover between successional stages at each site might, at least
384 in part, be attributed to possible differences in interaction networks within the communities between the
385 two sites.

386 An additionally important factor in the assembly and succession of surface-attached communities
387 in aquatic environments is the invasion by species from the surrounding water phase (Battin et al., 2016).
388 The establishment of invading species in a biofilm community depends on both stochastic dispersal as
389 well as interactions with already established species (Battin et al., 2007; Battin et al., 2016). Beta diversity

390 partitioning showed that sediment-attached and planktonic communities were composed of distinct sets of
391 OTUs. We again used the null model approach to test in how far deterministic and stochastic processes
392 contributed to these differences. We found that over the first successional stages the turnover between
393 sediment-attached and planktonic communities was caused by dispersal limitation acting alongside drift
394 and later on shifted towards variable selection. The latter observation could be explained in the light of
395 previous studies which suggested that species with similar ecological niches as resident species have a
396 lower chance of successfully invading a community than species that have less niche overlap with already
397 established species (Fargione et al., 2003; Tilman, 2004; Peay et al., 2012; Tan et al., 2012).

398 The processes that were indicated to have driven community turnover between groundwater and
399 sediment at the earlier stages were however counterintuitive. Unexpectedly, significantly positive values
400 for RC_{bray} suggested dispersal limitation acting alongside drift to have been responsible for the observed
401 differences in community composition, rather than the intuitively more expected scenario of random drift
402 acting alone. Multiple causes could explain these unexpected findings. It has to be noted that the sediment
403 microcosms were incubated in groundwater monitoring wells. It is known that communities found inside
404 monitoring wells may differ from the communities that are actually present in the surrounding
405 groundwater of an aquifer (Griebler et al., 2002; Korbel et al., 2017). In fact, previous analyses of our
406 samples by T-RFLP fingerprinting did indeed reveal some differences between groundwater and well
407 water microbial communities (Zhou et al., 2012). However, Langenheder et al. (2017) have reported
408 identical results for differences between lake biofilms and microbial communities in the overlaying water
409 column, which were not separated by any barrier that could have limited OTU dispersal. They argued in
410 the light of these findings, and based on the arguments provided by Chase et al. (2011), that significantly
411 positive deviations of RC_{bray} from the null expectation may also be caused by strong biotic factors such as
412 competition between species. As niches become more crowded over time, some organisms may try to
413 avoid competition by occupying non-optimal niches, which would not necessarily result in a deviation
414 from the null expectation in phylogenetic null models. Moreover, even though the assumption underlying
415 the β -NTI-based approach about the link between phylogenetic relatedness and niche similarity of

416 microbial species is supported by empirical evidence (Peay et al., 2012; Stegen et al., 2012; Tan et al.,
417 2012; Wang et al., 2013; Dini-Andreote et al., 2015; Martiny et al., 2015), and was also confirmed in our
418 system by a significant phylogenetic signal (Fig. S6), it is known that some species traits are
419 phylogenetically more conserved than others (Martiny et al., 2015). Hence, we may speculate that the
420 significantly positive deviation of RC_{bray} from the null exception at the early stage of the colonization
421 might indicate the involvement of traits that are important for the colonization of sediment surfaces, but
422 which are phylogenetically not well conserved and therefore did not result in a significant signal of β -NTI.
423 Only at the later stage, when the communities on the sediments had matured further, phylogenetically
424 more conserved traits may have gained importance in the turnover between planktonic and attached
425 microbial communities.

426

427 **Conclusion**

428 We have shown that the microbial colonization of sediments in a pristine groundwater aquifer in several
429 aspects follows the general patterns that have also been described for the development of biofilms in other
430 more energy-rich, non-subsurface, aquatic environments (Jackson, 2003), as well as the assembly of
431 sediment-attached communities in highly dynamic hyporheic zones, suggesting that the assembly of
432 microbial communities on surfaces might be governed by similar underlying mechanisms across a wide
433 range of different habitats. Our results indicate that differences between planktonic and sediment-attached
434 communities often reported for groundwater environments are not the result of purely stochastic events,
435 but that sediment surfaces select for specific groups of microorganisms that assemble over time in a
436 reproducible, non-random way, probably determined by sediment properties rather than hydrochemistry.
437 Although we found that early-colonizing OTUs dominated the final communities on the sediments, mere
438 timing OTU of arrival during the succession was likely not a determining factor, as the majority of these
439 early-colonizers were not very persistent. Rather, traits associated with identified key taxa, especially
440 within the *Comamonadaceae* and *Oxalobacteraceae*, seemed to have been a more decisive factor for the
441 persistence of these OTUs. However, the ecological processes behind the temporal succession of OTUs

442 during the colonization still remain unclear and might be influenced by species interaction network
443 structures at a given time point. Moreover, we found indications that different traits with different degrees
444 of phylogenetic conservation may have determined the establishment of OTUs in the developing
445 sediment-attached communities from the surrounding groundwater at different stages of community
446 development. A better understanding of these traits and how they may integrate into species interaction
447 networks will be an important aspect for future research. Computational modelling of microbial
448 communities based on metaomics data, albeit still in its infancy, offers a promising tool to elucidate
449 complex species interactions within microbial communities (Faust and Raes, 2012; Hanemaaijer et al.,
450 2015). If successful, the extra in depth insight gained from such models could be a valuable addition to
451 current approaches that strive for a better understanding of the links between microbial community
452 composition, assembly, and biogeochemical functions (Graham et al., 2016b; Graham and Stegen, 2017).

453

454 **Experimental procedures**

455 **Experimental setup and sampling**

456 To study the assembly and succession of sediment-attached microbial communities, fresh sediments were
457 taken from the Isar River that drains the investigated aquifer. Sediments were sieved (0.2-0.63 mm) and
458 packed into perforated polyethylene columns with a mesh size of 1-2 mm. Sediment columns were
459 submerged in deionized water and sterilized by autoclaving five times at 121°C for 30 min; after each
460 step, the sediments were rinsed with and again submerged in fresh deionized water. The columns were
461 stored at 4°C submerged in sterile water until the start of the experiment. Replicate sediment columns
462 were incubated in each well; duplicate columns were sampled destructively at each sampling campaign.
463 Samples for DNA extraction were put on dry ice for transport to the lab and were stored at -20°C until
464 DNA extraction according to the method described by Anneser et al. (2010). For the comparison of
465 attached versus planktonic microbial communities, cells from 5 L groundwater were collected on a 0.2 µm
466 polycarbonate filter (Merck Millipore, Darmstadt, Germany) on-site. Filters were shock-frozen on dry ice
467 and stored at -20°C until extraction using the same method as for the sediment samples. For cell counting,

468 0.5 mL groundwater (or 0.5 cm³ sediment) was fixed on-site with glutardialdehyde at a final concentration
469 of 2.5% v/v; samples were stored in the dark at 4°C until further processing according to Bayer et al.
470 (2016). Cells were stained with SYBR-Green I (Invitrogen, Karlsruhe, Germany) at a ratio of 1:10,000
471 and subsequently counted using a LSR II flow cytometer (Becton Dickinson, Heidelberg, Germany). For a
472 description of measurements of physicochemical parameters listed in Table 1 the reader is referred to
473 Zhou et al. (2012).

474

475 **16S rRNA amplicon sequencing**

476 PCR amplification (28 cycles) and subsequent bidirectional 454-pyrosequencing of 16S rRNA gene
477 fragments was done according to Pilloni et al. (2011) using the primer pair Ba27f-Ba519r extended with
478 sequencing adapters and multiplex barcodes. Each of the sample duplicates was amplified again in
479 duplicate; after amplification, all replicates of a given sample were combined before purification using
480 magnetic beads (AMPure-XP; Beckmann Coulter, Brea, CA, USA) according to the manufacturer's
481 instructions. After purification, DNA concentrations were determined using the Quant-iT PicoGreen
482 dsDNA Assay Kit (Invitrogen, Paisley, UK). Barcoded amplicons from all samples were pooled in
483 equimolar amounts before sequencing on a 454 GS FLX pyrosequencer using Titanium chemistry (Roche,
484 Penzberg, Germany). The sequence data was processed with QIIME (version 1.9.0) (Caporaso et al.,
485 2010). Demultiplexing and quality filtering (min./max. sequence length: 250/600 bp; primer mismatches
486 and barcode errors: 0; min. quality score: 25; quality score window size: 50 bp) was done using the
487 'split_libraries.py' command. Chimera filtering was done by mapping reads against the SILVA SSU
488 reference database (release 128) (Quast et al., 2013) using 'identify_chimeric_seqs.py' with usearch61 as
489 detection method. After quality and chimera filtering, the average number of combined forward and
490 reverse reads per sample was 5,709 with an average length of 388 bp. OTUs were clustered by uclust
491 against the SILVA SSU reference database at 97% similarity using the 'pick_open_reference_otus.py'
492 command. After removing low-confidence OTUs (combined abundance of < 0.01% across all samples)
493 and OTUs classified as chloroplasts, a total of 910 OTUs remained in the final OTU table. The total

494 number of reads per sample was rarefied to 2,045 which was the lowest number of reads observed for a
495 single sample. A midpoint-rooted phylogenetic tree was constructed from the alignment of OTU
496 reference sequences using FastTree (Price et al., 2009). Sequence data have been deposited in the NCBI
497 Sequence Read Archive under accession number SRP139256.

498

499 **Data analysis**

500 All analyses were done in R (version 3.5.0) (R Core Team, 2018). Alpha diversity (OTU richness (S),
501 Shannon diversity (H'), Pilon's evenness (J')) was calculated using the vegan package (version 2.5-2)
502 (Oksanen et al., 2018). The number of newly-arriving OTUs (S_n) in sediment samples at a given time
503 point was defined as the number of OTUs that displayed an abundance $> 0\%$ for the first time at that time
504 point. Phylogenetic beta diversity was assessed based on β -MNTD that was calculated using the
505 'comdistnt' function of the picante package (version 1.7) (Kembel et al., 2010). Differences in microbial
506 community composition between samples across time, space, and community type were illustrated by
507 NMDS performed on the β -MNTD matrix using the 'metaMDS' function of the vegan package with 40
508 iterations. To test for the effect of physicochemical variables (Table 1) on changes in community
509 composition for sediment-attached and planktonic communities, respectively, variables were standardized
510 to z-scores before fitting to the NDMS ordination using the 'envfit' function of the vegan package with
511 10,000 permutations stratified within community types. PERMANOVA was used to estimate the marginal
512 effects of each of the three categorical variables community type, sampling time point, and site location,
513 respectively, while holding the other two constant using the 'adonis2' function in vegan package with
514 10,000 permutations. For the identification of key organisms that were responsible for the differences
515 between community types, relative OTU abundances were summarized at genus level before SIMPER
516 analysis using the 'simper' function in vegan with 1,000 permutations for significance testing. Beta
517 diversity partitioning based on Jaccard dissimilarity was done using the 'betapart' package (Baselga and
518 Orme, 2012).

519 To study the effect of deterministic versus stochastic processes on microbial community assembly,
520 we used the null model approach developed by Stegen et al. (2012; 2013). β -NTI compares the mean
521 phylogenetic distance of OTUs based on β -MNTD between two communities against the distribution of β -
522 MNTD values expected for randomly assembled communities. This distribution is obtained from repeated
523 randomizations in which the OTUs observed in the two communities and their relative abundances are
524 shuffled across the tips of the according phylogenetic tree. The value of β -NTI indicates by how many
525 standard deviations the observed β -MNTD deviates from the mean of the null expectation with $|\beta\text{-NTI}| > 2$
526 indicating significant deviations. β -NTI was calculated with abundance-weighting and 999 randomizations
527 for each pairwise comparison. The assumption of a significant phylogenetic signal was verified using
528 Mantel correlograms as in Dini-Andreote et al. (2015) (see SI and Fig. S6). The RC_{bray} index measures
529 how much the observed Bray-Curtis dissimilarity between two communities differs from the distribution
530 of dissimilarities between probabilistically assembled communities for which the probability of OTUs
531 being drawn is proportional to their respective abundances in the two compared communities and their
532 occurrence frequencies in the regional species pool, while keeping local community richness and the
533 number of individuals constant. RC_{bray} takes values from -1 to +1 where absolute values > 0.95 indicate
534 significant deviations from the null expectation. RC_{bray} was calculated with 999 iterations for each
535 pairwise comparison. Regional species pools for null model simulations were constructed from all OTUs
536 in the full dataset over space and time as in Veach et al. (2016), because we expected that regional species
537 pools constructed separately for each time point from OTUs at the two sites that only spanned a relatively
538 short transect would have been too conservative to estimate the total regional diversity in the aquifer. To
539 evaluate in how far this large regional species pool may have led to an overestimation of the effects of
540 selection and/or dispersal, we compared these results to simulations where regional species pools were
541 constructed for individual time points for which paired samples of sediment-attached and planktonic
542 communities were available. The outcomes of the null models in both situations were in high agreement
543 with each other, indicating that using the full dataset to construct the regional species pool did not
544 introduce a substantial bias in our analyses (see SI and Fig. S7). To test for the effect of changes in

545 physicochemical conditions on community assembly processes, Mantel tests (Spearman's rank
546 correlation, 10,000 permutations, function 'mantel' in vegan) were performed on the β -NTI matrix and
547 individual Euclidean distance matrices that were calculated for each physicochemical variable separately
548 after standardization.

549

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559

560 **References**

- 561
562 Anneser, B., Piloni, G., Bayer, A., Lueders, T., Griebler, C., Einsiedl, F., and Richters, L. (2010) High
563 Resolution Analysis of Contaminated Aquifer Sediments and Groundwater—What Can be Learned in
564 Terms of Natural Attenuation? *Geomicrobiology Journal* **27**: 130-142.
565
- 566 Baldani, I., Rouws, L., Cruz, L., Olivares, F., Schmid, M., and Hartmann, A. (2014) The Family
567 Oxalobacteraceae. In *The Prokaryotes, Alphaproteobacteria and Betaproteobacteria*. Rosenberg, E.,
568 DeLong, E.F., Lory, S., Stackebrandt, E., and Thompson, F. (eds). Heidelberg, Germany: Springer, pp. 919-
569 974.
570
- 571 Baselga, A. (2012) The relationship between species replacement, dissimilarity derived from nestedness,
572 and nestedness. *Global Ecology and Biogeography* **21**: 1223-1232.
573
- 574 Baselga, A., and Orme, C.D.L. (2012) betapart: an R package for the study of beta diversity. *Methods in*
575 *Ecology and Evolution* **3**: 808-812.
576
- 577 Battin, T.J., Besemer, K., Bengtsson, M.M., Romani, A.M., and Packmann, A.I. (2016) The ecology and
578 biogeochemistry of stream biofilms. *Nature Reviews Microbiology* **14**: 251-263.
579
- 580 Battin, T.J., Sloan, W.T., Kjelleberg, S., Daims, H., Head, I.M., Curtis, T.P., and Eberl, L. (2007) Microbial
581 landscapes: new paths to biofilm research. *Nature Reviews Microbiology* **5**: 76-81.
582
- 583 Bayer, A., Drexel, R., Weber, N., and Griebler, C. (2016) Quantification of aquatic sediment prokaryotes—
584 A multiple-steps optimization testing sands from pristine and contaminated aquifers. *Limnologica -*
585 *Ecology and Management of Inland Waters* **56**: 6-13.
586
- 587 Beaton, E.D., Stevenson, B.S., King-Sharp, K.J., Stamps, B.W., Nunn, H.S., and Stuart, M. (2016) Local and
588 Regional Diversity Reveals Dispersal Limitation and Drift as Drivers for Groundwater Bacterial
589 Communities from a Fractured Granite Formation. *Frontiers in Microbiology* **7**: 1933.
590
- 591 Besemer, K., Peter, H., Logue, J.B., Langenheder, S., Lindström, E.S., Tranvik, L.J., and Battin, T.J. (2012)
592 Unraveling assembly of stream biofilm communities. *The ISME Journal* **6**: 1459-1468.
593
- 594 Bižić-Ionescu, M., Zeder, M., Ionescu, D., Orlić, S., Fuchs, B.M., Grossart, H.-P., and Amann, R. (2014)
595 Comparison of bacterial communities on limnic versus coastal marine particles reveals profound
596 differences in colonization. *Environmental Microbiology* **17**: 3500-3514.
597
- 598 Brislawn, C.J., Graham, E.B., Dana, K., Ihardt, P., Fansler, S.J., Chirlser, W.B. et al. (2018) Forfeiting the
599 founder effect: turnover defines biofilm community succession. *bioRxiv pre-print; article version: March*
600 **15, 2018**.
601
- 602 Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K. et al. (2010) QIIME
603 allows analysis of high-throughput community sequencing data. *Nature Methods* **7**: 335-336.
604

- 605 Chase, J.M., and Myers, J.A. (2011) Disentangling the importance of ecological niches from stochastic
606 processes across scales. *Philosophical Transactions of the Royal Society B: Biological Sciences* **366**: 2351-
607 2363.
- 608
- 609 Chase, J.M., Kraft, N.J.B., Smith, K.G., Vellend, M., and Inouye, B.D. (2011) Using null models to
610 disentangle variation in community dissimilarity from variation in α -diversity. *Ecosphere* **2**: 1-11.
- 611
- 612 Cordero, O.X., and Datta, M.S. (2016) Microbial interactions and community assembly at microscales.
613 *Current Opinion in Microbiology* **31**: 227-234.
- 614
- 615 Danczak, R.E., Johnston, M.D., Kenah, C., Slattery, M., and Wilkins, M.J. (2018) Microbial Community
616 Cohesion Mediates Community Turnover in Unperturbed Aquifers. *mSystems* **3**: e00066-00018.
- 617
- 618 Dini-Andreote, F., Stegen, J.C., van Elsas, J.D., and Salles, J.F. (2015) Disentangling mechanisms that
619 mediate the balance between stochastic and deterministic processes in microbial succession.
620 *Proceedings of the National Academy of Sciences* **112**: E1326-E1332.
- 621
- 622 Dumbrell, A.J., Nelson, M., Helgason, T., Dytham, C., and Fitter, A.H. (2009) Relative roles of niche and
623 neutral processes in structuring a soil microbial community. *The Isme Journal* **4**: 337.
- 624
- 625 Fargione, J., Brown, C.S., and Tilman, D. (2003) Community assembly and invasion: An experimental test
626 of neutral versus niche processes. *Proceedings of the National Academy of Sciences* **100**: 8916-8920.
- 627
- 628 Faust, K., and Raes, J. (2012) Microbial interactions: from networks to models. *Nature Reviews*
629 *Microbiology* **10**: 538.
- 630
- 631 Ferrenberg, S., O'Neill, S.P., Knelman, J.E., Todd, B., Duggan, S., Bradley, D. et al. (2013) Changes in
632 assembly processes in soil bacterial communities following a wildfire disturbance. *The Isme Journal* **7**:
633 1102.
- 634
- 635 Flynn, T.M., Sanford, R.A., Ryu, H., Bethke, C.M., Levine, A.D., Ashbolt, N.J., and Santo Domingo, J.W.
636 (2013) Functional microbial diversity explains groundwater chemistry in a pristine aquifer. *BMC*
637 *Microbiology* **13**: 146.
- 638
- 639 Fukami, T. (2015) Historical Contingency in Community Assembly: Integrating Niches, Species Pools, and
640 Priority Effects. *Annual Review of Ecology, Evolution, and Systematics* **46**: 1-23.
- 641
- 642 Fukami, T., Dickie, I.A., Paula Wilkie, J., Paulus, B.C., Park, D., Roberts, A. et al. (2010) Assembly history
643 dictates ecosystem functioning: evidence from wood decomposer communities. *Ecology Letters* **13**: 675-
644 684.
- 645
- 646 Graham, B.E., and Stegen, C.J. (2017) Dispersal-Based Microbial Community Assembly Decreases
647 Biogeochemical Function. *Processes* **5**.
- 648
- 649 Graham, E.B., Crump, A.R., Resch, C.T., Fansler, S., Arntzen, E., Kennedy, D.W. et al. (2016a) Coupling
650 Spatiotemporal Community Assembly Processes to Changes in Microbial Metabolism. *Frontiers in*
651 *Microbiology* **7**: 1949.
- 652

- 653 Graham, E.B., Crump, A.R., Resch, C.T., Fansler, S., Arntzen, E., Kennedy, D.W. et al. (2017) Deterministic
 654 influences exceed dispersal effects on hydrologically-connected microbiomes. *Environmental*
 655 *Microbiology* **19**: 1552-1567.
- 656
- 657 Graham, E.B., Knelman, J.E., Schindlbacher, A., Siciliano, S., Breulmann, M., Yannarell, A. et al. (2016b)
 658 Microbes as Engines of Ecosystem Function: When Does Community Structure Enhance Predictions of
 659 Ecosystem Processes? *Frontiers in Microbiology* **7**: 214.
- 660
- 661 Griebler, C., and Lueders, T. (2009) Microbial biodiversity in groundwater ecosystems. *Freshwater*
 662 *Biology* **54**: 649-677.
- 663
- 664 Griebler, C., and Avramov, M. (2015) Groundwater ecosystem services: a review. *Freshwater Science* **34**:
 665 355-367.
- 666
- 667 Griebler, C., Malard, F., and Lefébure, T. (2014) Current developments in groundwater ecology — from
 668 biodiversity to ecosystem function and services. *Current Opinion in Biotechnology* **27**: 159-167.
- 669
- 670 Griebler, C., Mindl, B., Slezak, D., and Geiger-Kaiser, M. (2002) Distribution patterns of attached and
 671 suspended bacteria in pristine and contaminated shallow aquifers studied with an in situ sediment
 672 exposure microcosm. *Aquatic Microbial Ecology* **28**: 117-129.
- 673
- 674 Grösbacher, M., Spicher, C., Bayer, A., Obst, M., Karwautz, C., Pilloni, G. et al. (2016) Organic
 675 contamination versus mineral properties: competing selective forces shaping bacterial community
 676 assembly in aquifer sediments. *Aquatic Microbial Ecology* **76**: 243-255.
- 677
- 678 Hanemaaijer, M., Röling, W.F.M., Olivier, B.G., Khandelwal, R.A., Teusink, B., and Bruggeman, F.J. (2015)
 679 Systems modeling approaches for microbial community studies: from metagenomics to inference of the
 680 community structure. *Frontiers in Microbiology* **6**: 213.
- 681
- 682 Hug, L.A., Thomas, B.C., Brown, C.T., Frischkorn, K.R., Williams, K.H., Tringe, S.G., and Banfield, J.F. (2015)
 683 Aquifer environment selects for microbial species cohorts in sediment and groundwater. *The ISME*
 684 *Journal* **9**: 1846–1856.
- 685
- 686 Jackson, C.R. (2003) Changes in community properties during microbial succession. *Oikos* **101**: 444-448.
- 687
- 688 Jones, A.A., and Bennett, P.C. (2017) Mineral Ecology: Surface Specific Colonization and Geochemical
 689 Drivers of Biofilm Accumulation, Composition, and Phylogeny. *Frontiers in Microbiology* **8**: 491.
- 690
- 691 Kalmbach, S., Manz, W., Bendinger, B., and Szewzyk, U. (2000) In situ probing reveals *Aquabacterium*
 692 *commune* as a widespread and highly abundant bacterial species in drinking water biofilms. *Water*
 693 *Research* **34**: 575-581.
- 694
- 695 Kembel, S.W., Cowan, P.D., Helmus, M.R., Cornwell, W.K., Morlon, H., Ackerly, D.D. et al. (2010) Picante:
 696 R tools for integrating phylogenies and ecology. *Bioinformatics* **26**: 1463-1464.
- 697
- 698 Konopka, A., Lindemann, S., and Fredrickson, J. (2015) Dynamics in microbial communities: unraveling
 699 mechanisms to identify principles. *The ISME Journal* **9**: 1488-1495.
- 700

- 701 Korbek, K., Chariton, A., Stephenson, S., Greenfield, P., and Hose, G.C. (2017) Wells provide a distorted
702 view of life in the aquifer: implications for sampling, monitoring and assessment of groundwater
703 ecosystems. *Scientific Reports* **7**: 40702.
704
- 705 Langenheder, S., and Székely, A.J. (2011) Species sorting and neutral processes are both important
706 during the initial assembly of bacterial communities. *The Isme Journal* **5**: 1086–1094.
707
- 708 Langenheder, S., Wang, J., Karjalainen, S.M., Laamanen, T.M., Tolonen, K.T., Vilmi, A., and Heino, J.
709 (2017) Bacterial metacommunity organization in a highly connected aquatic system. *FEMS Microbiology*
710 *Ecology* **93**: fiw225–fiw225.
711
- 712 Lehman, R.M., Colwell, F.S., and Bala, G.A. (2001) Attached and Unattached Microbial Communities in a
713 Simulated Basalt Aquifer under Fracture- and Porous-Flow Conditions. *Applied and Environmental*
714 *Microbiology* **67**: 2799–2809.
715
- 716 Lyautey, E., Jackson, C.R., Cayrou, J., Rols, J.-L., and Garabétian, F. (2005) Bacterial Community
717 Succession in Natural River Biofilm Assemblages. *Microbial Ecology* **50**: 589–601.
718
- 719 Martiny, A.C., Jørgensen, T.M., Albrechtsen, H.-J., Arvin, E., and Molin, S. (2003) Long-Term Succession of
720 Structure and Diversity of a Biofilm Formed in a Model Drinking Water Distribution System. *Applied and*
721 *Environmental Microbiology* **69**: 6899–6907.
722
- 723 Martiny, J.B.H., Jones, S.E., Lennon, J.T., and Martiny, A.C. (2015) Microbiomes in light of traits: A
724 phylogenetic perspective. *Science* **350**: aac9323.
725
- 726 McMahon, S., and Parnell, J. (2013) Weighing the deep continental biosphere. *FEMS Microbiology*
727 *Ecology* **87**: 113–120.
728
- 729 Nemergut, D.R., Schmidt, S.K., Fukami, T., O'Neill, S.P., Bilinski, T.M., Stanish, L.F. et al. (2013) Patterns
730 and Processes of Microbial Community Assembly. *Microbiology and Molecular Biology Reviews* : *MMBR*
731 **77**: 342–356.
732
- 733 Niederdorfer, R., Peter, H., and Battin, T.J. (2016) Attached biofilms and suspended aggregates are
734 distinct microbial lifestyles emanating from differing hydraulics. *Nature Microbiology* **1**: 16178.
735
- 736 Niederdorfer, R., Besemer, K., Battin, T.J., and Peter, H. (2017) Ecological strategies and metabolic trade-
737 offs of complex environmental biofilms. *npj Biofilms and Microbiomes* **3**: 21.
738
- 739 Ofițeru, I.D., Lunn, M., Curtis, T.P., Wells, G.F., Criddle, C.S., Francis, C.A., and Sloan, W.T. (2010)
740 Combined niche and neutral effects in a microbial wastewater treatment community. *Proceedings of the*
741 *National Academy of Sciences* **107**: 15345–15350.
742
- 743 Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlenn, D. et al. (2018) vegan:
744 Community Ecology Package (R package version 2.5-2).
745
- 746 Peay, K.G., Belisle, M., and Fukami, T. (2012) Phylogenetic relatedness predicts priority effects in nectar
747 yeast communities. *Proceedings of the Royal Society B: Biological Sciences* **279**: 749–758.
748

- 749 Pilloni, G., von Netzer, F., Engel, M., and Lueders, T. (2011) Electron acceptor-dependent identification of
750 key anaerobic toluene degraders at a tar-oil-contaminated aquifer by Pyro-SIP. *FEMS Microbiology*
751 *Ecology* **78**: 165-175.
- 752
- 753 Price, M.N., Dehal, P.S., and Arkin, A.P. (2009) FastTree: Computing Large Minimum Evolution Trees with
754 Profiles instead of a Distance Matrix. *Molecular Biology and Evolution* **26**: 1641-1650.
- 755
- 756 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P. et al. (2013) The SILVA ribosomal RNA
757 gene database project: improved data processing and web-based tools. *Nucleic Acids Research* **41**: D590-
758 D596.
- 759
- 760 R Core Team (2018) R: A Language and Environment for Statistical Computing. In. Austria: R Foundation
761 for Statistical Computing.
- 762
- 763 Rummens, K., De Meester, L., and Souffreau, C. (2018) Inoculation history affects community
764 composition in experimental freshwater bacterioplankton communities. *Environmental Microbiology* **20**:
765 1120-1133.
- 766
- 767 Schmidt, S.I., Cuthbert, M.O., and Schwientek, M. (2017) Towards an integrated understanding of how
768 micro scale processes shape groundwater ecosystem functions. *Science of The Total Environment* **592**:
769 215-227.
- 770
- 771 Stegen, J.C., Lin, X., Konopka, A.E., and Fredrickson, J.K. (2012) Stochastic and deterministic assembly
772 processes in subsurface microbial communities. *The Isme Journal* **6**: 1653-1664.
- 773
- 774 Stegen, J.C., Lin, X., Fredrickson, J.K., and Konopka, A.E. (2015) Estimating and mapping ecological
775 processes influencing microbial community assembly. *Frontiers in Microbiology* **6**: 370.
- 776
- 777 Stegen, J.C., Lin, X., Fredrickson, J.K., Chen, X., Kennedy, D.W., Murray, C.J. et al. (2013) Quantifying
778 community assembly processes and identifying features that impose them. *The Isme Journal* **7**: 2069-
779 2079.
- 780
- 781 Stegen, J.C., Konopka, A., McKinley, J.P., Murray, C., Lin, X., Miller, M.D. et al. (2016a) Coupling among
782 Microbial Communities, Biogeochemistry, and Mineralogy across Biogeochemical Facies. *Scientific*
783 *Reports* **6**: 30553.
- 784
- 785 Stegen, J.C., Fredrickson, J.K., Wilkins, M.J., Konopka, A.E., Nelson, W.C., Arntzen, E.V. et al. (2016b)
786 Groundwater–surface water mixing shifts ecological assembly processes and stimulates organic carbon
787 turnover. *Nature Communications* **7**: 11237.
- 788
- 789 Stegen, J.C., Johnson, T., Fredrickson, J.K., Wilkins, M.J., Konopka, A.E., Nelson, W.C. et al. (2018)
790 Influences of organic carbon speciation on hyporheic corridor biogeochemistry and microbial ecology.
791 *Nature Communications* **9**: 585.
- 792
- 793 Svoboda, P., Lindström, E.S., Ahmed Osman, O., and Langenheder, S. (2018) Dispersal timing determines
794 the importance of priority effects in bacterial communities. *The Isme Journal* **12**: 644-646.
- 795
- 796 Tan, J., Pu, Z., Ryberg, W.A., and Jiang, L. (2012) Species phylogenetic relatedness, priority effects, and
797 ecosystem functioning. *Ecology* **93**: 1164-1172.

- 798
799 Tilman, D. (2004) Niche tradeoffs, neutrality, and community structure: A stochastic theory of resource
800 competition, invasion, and community assembly. *Proceedings of the National Academy of Sciences* **101**:
801 10854-10861.
802
- 803 Veach, A.M., Stegen, J.C., Brown, S.P., Dodds, W.K., and Jumpponen, A. (2016) Spatial and successional
804 dynamics of microbial biofilm communities in a grassland stream ecosystem. *Molecular Ecology* **25**:
805 4674-4688.
806
- 807 Wang, J., Shen, J., Wu, Y., Tu, C., Soininen, J., Stegen, J.C. et al. (2013) Phylogenetic beta diversity in
808 bacterial assemblages across ecosystems: deterministic versus stochastic processes. *The Isme Journal* **7**:
809 1310-1321.
810
- 811 Woodcock, S., and Sloan, W.T. (2017) Biofilm community succession: a neutral perspective. *Microbiology*
812 **163**: 664-668.
813
- 814 Woodcock, S., van der Gast, C.J., Bell, T., Lunn, M., Curtis, T.P., Head, I.M., and Sloan, W.T. (2007) Neutral
815 assembly of bacterial communities. *FEMS Microbiology Ecology* **62**: 171-180.
816
- 817 Zhou, J., and Ning, D. (2017) Stochastic Community Assembly: Does It Matter in Microbial Ecology?
818 *Microbiology and Molecular Biology Reviews* **81**: e00002-00017.
819
- 820 Zhou, J., Liu, W., Deng, Y., Jiang, Y.-H., Xue, K., He, Z. et al. (2013) Stochastic Assembly Leads to
821 Alternative Communities with Distinct Functions in a Bioreactor Microbial Community. *mBio* **4**: e00584-
822 00512.
823
- 824 Zhou, J., Deng, Y., Zhang, P., Xue, K., Liang, Y., Van Nostrand, J.D. et al. (2014) Stochasticity, succession,
825 and environmental perturbations in a fluidic ecosystem. *Proceedings of the National Academy of*
826 *Sciences* **111**: E836-E845.
827
- 828 Zhou, Y., Kellermann, C., and Griebler, C. (2012) Spatio-temporal patterns of microbial communities in a
829 hydrologically dynamic pristine aquifer. *FEMS Microbiology Ecology* **81**: 230-242.

830

831 **Figure legends**

832 **Figure 1:** Schematic illustration of the main geographical features of the study site, the locations of the
833 two monitoring wells, and general groundwater flow directions.

834

835 **Figure 2:** Temporal dynamics of biomass and alpha diversity for sediment-attached communities and
836 planktonic communities in the groundwater at MIT052 and MIT039, respectively. Biomass is given as the
837 number of prokaryotic cells per mL groundwater or the equivalent volume (cm³) of sediment. The time in
838 days for each time point is given in parentheses.

839
840 **Figure 3:** (A) Changes over time in the fraction of newly-arriving OTUs (S_n) relative to the total number
841 of OTUs (S) in sediment-attached communities. (B) Changes over time in the cumulative relative
842 abundance of newly-arriving OTUs that first occurred at the time point indicated in the upper right corner
843 of each graph.

844
845 **Figure 4:** (A) Taxonomic microbial community composition of sediment-attached communities and
846 planktonic communities in the groundwater per time point. Taxonomic groups are summarized at family
847 level. Families with an abundance $< 10\%$ in the entire dataset were grouped as ‘Diverse others’ for clarity
848 of display (mean abundance: 0.1%; max: 4.7%). (B) NMDS plot showing differences in microbial
849 community composition based on abundance-weighted β -MNTD (stress: 0.09).

850
851 **Figure 5:** Values for β -NTI from pairwise microbial community comparisons. The range of β -NTI
852 indicating stochastic community turnover is shaded in grey. (A) Spatial community turnover of planktonic
853 communities in the groundwater and sediment-attached communities, respectively, between MIT052 and
854 MIT039 per time point. (B) Community turnover between planktonic and sediment-attached communities
855 within sites per time point (note: the bar corresponding to the comparison of communities at MIT052 in
856 July is not visible; β -NTI = 0.04). (C) Temporal community turnover of planktonic and sediment-attached
857 communities, respectively, between consecutive time points within sites (note: bar for the comparison of
858 July vs. December for planktonic communities at MIT052 is not visible; β -NTI = -0.005). NA: data not
859 available.

860
861 **Figure S1:** Changes in physicochemical parameters over time measured in the groundwater at the two
862 sites. DOC: dissolved organic carbon; AOC: assimilable organic carbon; DO: dissolved oxygen.

863

864 **Figure S2:** Occurrence frequencies of the most dominant families and most dominant single genera within
865 those families among newly-arriving OTUs (S_n) in the sediment-attached communities that first occurred
866 at the indicated time point and were still detected in the final communities at MIT052 and MIT039,
867 respectively. Families with an individual occurrence frequency $< 3\%$ are grouped as ‘Diverse others’ for
868 clarity of display. Bars representing each family are ordered by occurrence frequency in descending order
869 from top to bottom.

870
871 **Figure S3:** Differential abundances of genera that contributed most to the dissimilarity between sediment-
872 attached communities and planktonic communities in the groundwater identified by SIMPER analysis.
873 Only the genera with the highest significant contribution to the dissimilarity are shown ($> 0.1\%$; $p < 0.05$);
874 the average contribution of each displayed genus is indicated by the color intensity of the bars. (A) Log_{10} -
875 ratios of differential average abundances in planktonic communities over sediment-attached communities
876 for genera found in both community types. (B) Average relative abundances of genera exclusively found
877 in one community type.

878
879 **Figure S4:** Relative contributions of turnover and nestedness to the total Jaccard dissimilarity between
880 sediment-attached and planktonic communities at each site per time point inferred from beta diversity
881 partitioning.

882
883 **Figure S5:** Relative contributions of turnover and nestedness to the total Jaccard dissimilarity between
884 sediment-attached communities within sites across time points inferred from beta diversity partitioning.

885
886 **Figure S6:** Phylogenetic signal inferred from Mantel correlograms showing Pearson correlation between
887 phylogenetic distances and differences in environmental optima between OTUs within phylogenetic
888 distance classes evaluated at distance steps of 0.01 for (A) sediment-attached and (B) planktonic
889 communities. Filled symbols indicate significant correlations ($p < 0.05$).

890

891 **Figure S7:** Comparison of the outcomes of null model simulations to estimate (A) β -NTI and (B) RC_{bray}

892 based on different regional species pools for pairwise community comparisons shown in Figure 5 and

893 discussed in the main text. The horizontal axes represent results based on regional species pools

894 constructed from OTUs found in the full data set; vertical axes show results based on regional species

895 pools constructed from OTUs found in subsets of samples within time points, or in case of comparisons to

896 estimate temporal community turnover, from two consecutive time points (only for β -NTI). Colors

897 represent the different investigated turnover processes shown in Figure 5 (see main text). Dashed lines

898 mark significance thresholds for each index (see main text). Linear regression slopes of the straight line \pm

899 0.95 confidence intervals and Pearson correlation coefficients are indicated in the figures. Flags indicate

900 pairwise comparisons for which the outcomes of the null models did not agree between the two strategies

901 for constructing regional species pools.

902

903 **Tables**

904 **Table 1:** Mean values and standard deviation (SD) of physicochemical parameters averaged over the two investigated sites and all time points (n =
 905 10). Individual values for each site and time point are shown in Figure S1.

	Water table (m bs) ^a	DOC ^b (mg L ⁻¹)	AOC ^c (µg L ⁻¹)	pH	Temperature (°C)	Electrical conductivity (µS cm ⁻¹)	DO ^d (mg L ⁻¹)	PO ₄ ³⁻ (mg L ⁻¹)	SO ₄ ²⁻ (mg L ⁻¹)	NO ₃ ⁻ (mg L ⁻¹)	Cl ⁻ (mg L ⁻¹)	K ⁺ (mg L ⁻¹)	Na ⁺ (mg L ⁻¹)	Ca ²⁺ (mg L ⁻¹)	Mg ²⁺ (mg L ⁻¹)
Mean	8.9	1.10	11.4	7.81	6.88	297	10.21	0.011	14.5	2.49	1.80	0.30	1.03	46.1	12.1
SD	5.0	0.37	14.7	0.10	0.59	41	0.82	0.015	8.8	0.78	1.13	0.07	0.62	5.5	1.0

906 ^a meter below surface907 ^b dissolved organic carbon908 ^c assimilable organic carbon909 ^d dissolved oxygen

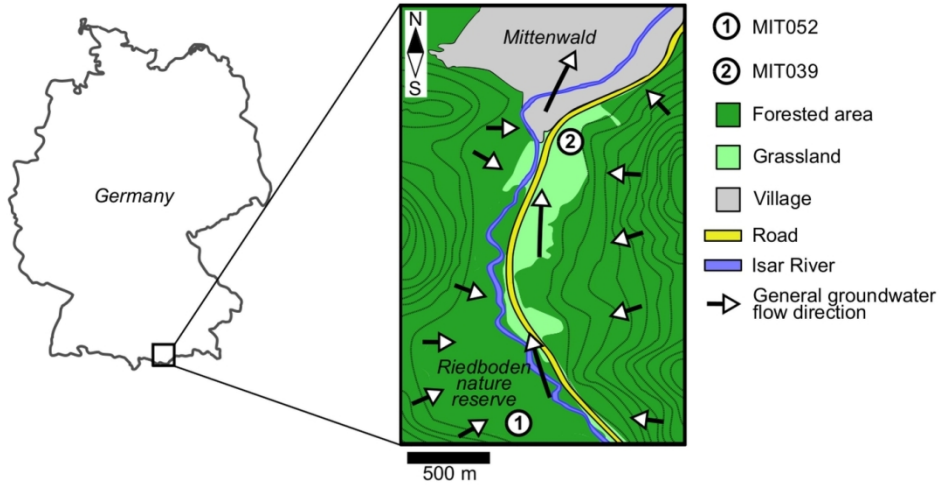


Figure 1: Schematic illustration of the main geographical features of the study site, the locations of the two monitoring wells, and general groundwater flow directions.

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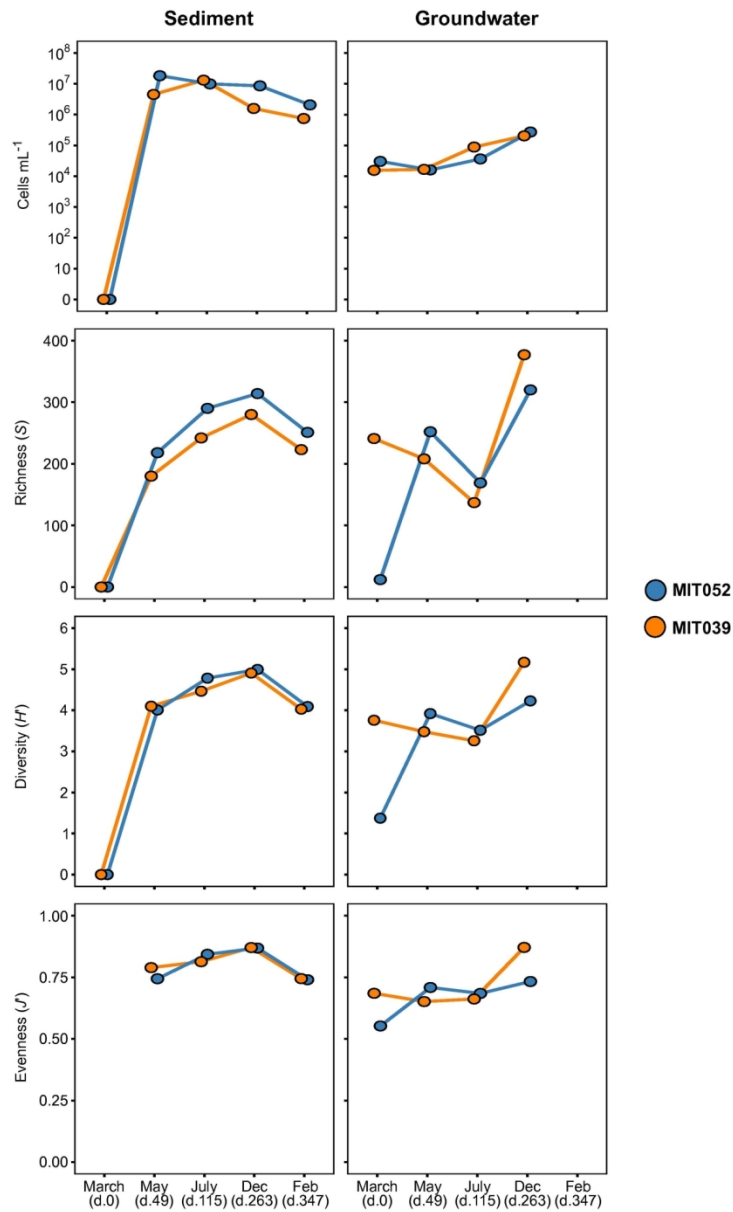


Figure 2: Temporal dynamics of biomass and alpha diversity for sediment-attached communities and planktonic communities in the groundwater at MIT052 and MIT039, respectively. Biomass is given as the number of prokaryotic cells per mL groundwater or the equivalent volume (cm³) of sediment. The time in days for each time point is given in parentheses.

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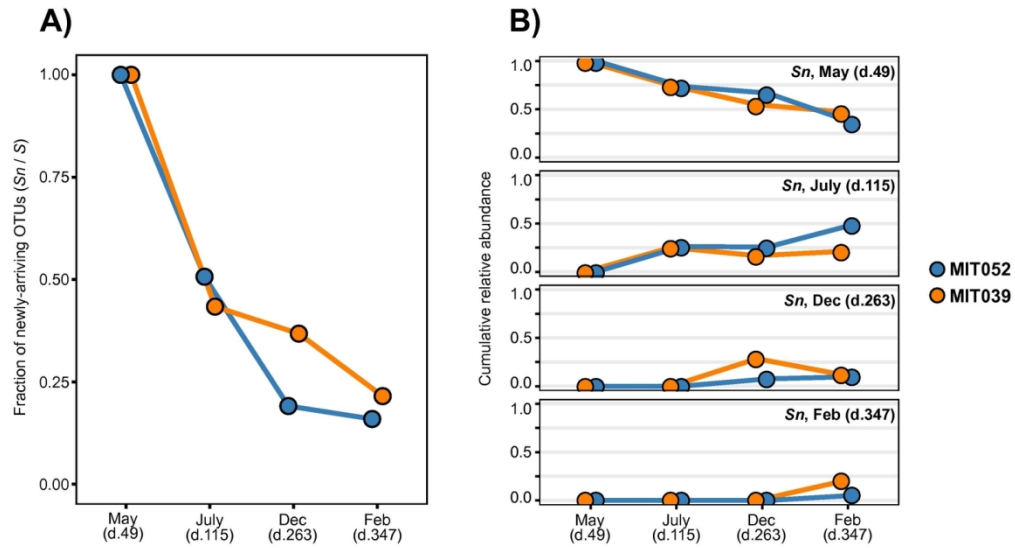


Figure 3: (A) Changes over time in the fraction of newly-arriving OTUs (S_n) relative to the total number of OTUs (S) in sediment-attached communities. (B) Changes over time in the cumulative relative abundance of newly-arriving OTUs that first occurred at the time point indicated in the upper right corner of each graph.

135x72mm (300 x 300 DPI)

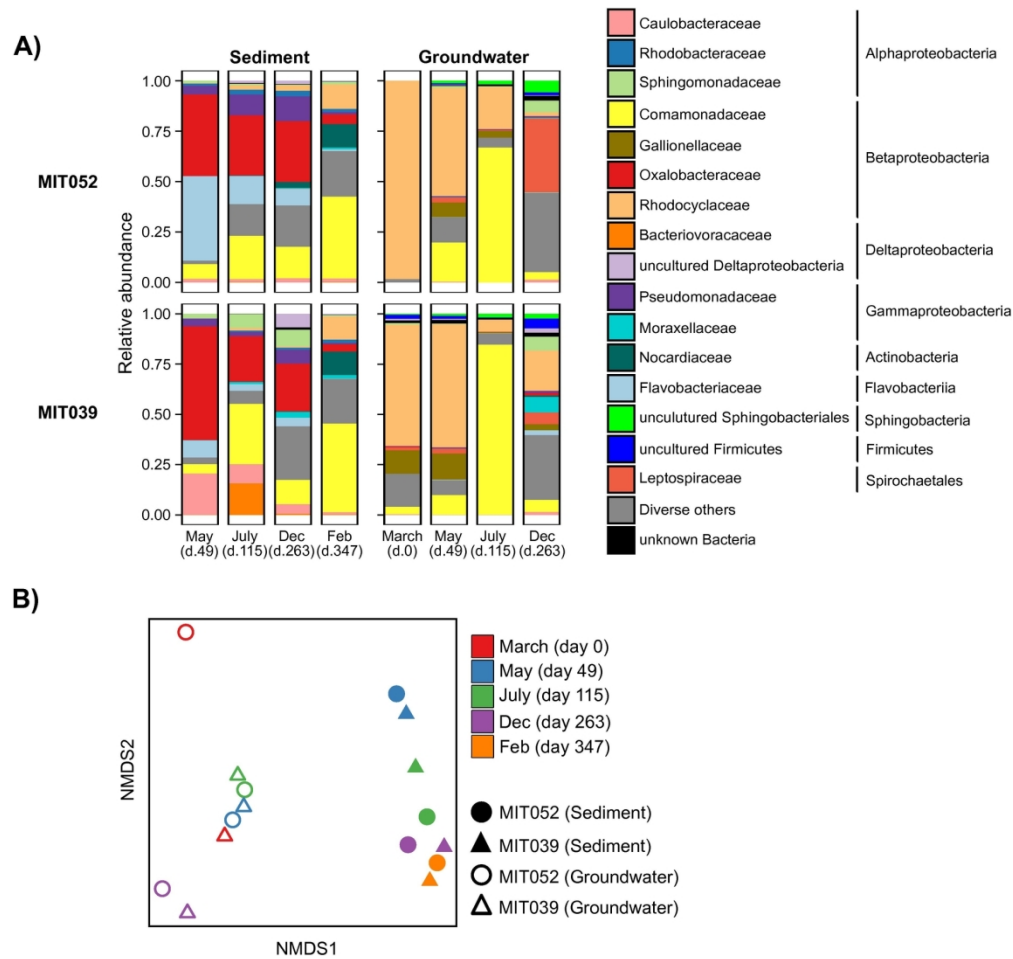


Figure 4: (A) Taxonomic microbial community composition of sediment-attached communities and planktonic communities in the groundwater per time point. Taxonomic groups are summarized at family level. Families with an abundance < 10% in the entire dataset were grouped as 'Diverse others' for clarity of display (mean abundance: 0.1%; max: 4.7%). (B) NMDS plot showing differences in microbial community composition based on abundance-weighted β -MNTD (stress: 0.09).

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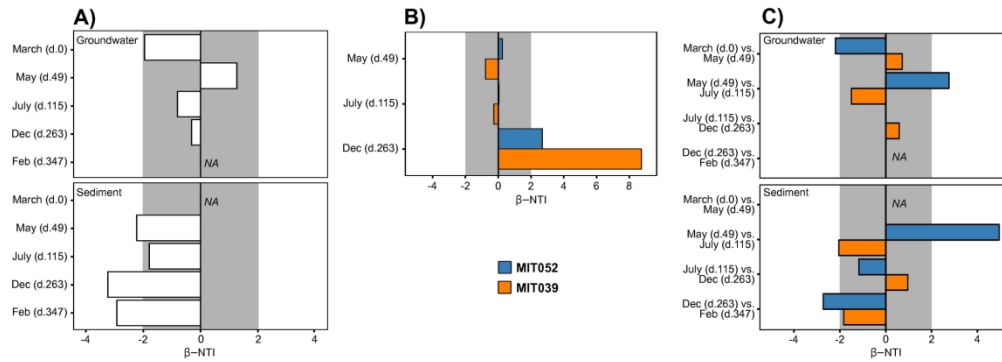


Figure 5: Values for β -NTI from pairwise microbial community comparisons. The range of β -NTI indicating stochastic community turnover is shaded in grey. (A) Spatial community turnover of planktonic communities in the groundwater and sediment-attached communities, respectively, between MIT052 and MIT039 per time point. (B) Community turnover between planktonic and sediment-attached communities within sites per time point (note: the bar corresponding to the comparison of communities at MIT052 in July is not visible; β -NTI = 0.04). (C) Temporal community turnover of planktonic and sediment-attached communities, respectively, between consecutive time points within sites (note: bar for the comparison of July vs. December for planktonic communities at MIT052 is not visible; β -NTI = -0.005). NA: data not available.

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