

Analytical aspects of meet-in-metabolite analysis for molecular pathway reconstitution from exposure to adverse outcome

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Abstract: To explore the etiology of diseases is one of the major goals in epidemiological study. Meet-in-metabolite analysis reconstitutes biomonitoring-based adverse outcome (AO) pathways from environmental exposure to a disease, in which the chemical exposome-related metabolism responses are transmitted to incur the AO-related metabolism phenotypes. However, the ongoing data-dependent acquisition of non-targeted biomonitoring by high-resolution mass spectrometry (HRMS) is biased against the low abundance molecules, which forms the major of molecular internal exposome, i.e., the totality of trace levels of environmental pollutants and/or their metabolites in human samples. The recent development of data-independent acquisition protocols for HRMS screening has opened new opportunities to enhance unbiased measurement of the extremely low abundance molecules, which can encompass a wide range of analytes and has been applied in metabolomics, DNA, and protein adductomics. In addition, computational MS for small molecules is urgently required for the top-down exposome databases. Although a holistic analysis of the exposome and endogenous metabolites is plausible, multiple and flexible strategies, instead of “putting one thing above all” are proposed.

Keywords: human biomonitoring; molecular exposome, metabolome, non-targeted analysis; adverse outcome pathway; system epidemiology

Introduction

Although it is considered a “black box” epidemiology, the traditional epidemiology founded the identification of disease etiologies (Helzlsouer, 1993), where an exposure factor is linked to a disease through statistical association. From exposure to prognostic significance, molecular epidemiology has incorporated a series of biomarkers; these biomarkers can be scattered and suggest pathways that may lead to disease initiation or development based on risk factor occurrence. Successful cases are rare if all the biomarkers are available along the exposure-adverse outcome course. These biomarkers have been individually investigated, and their assessments focus on the reliability and validity of statistical models such as receiver operating characteristic- or sensitive curve-based models. Biological relationships between the upstream and

45 downstream biomarkers, also known as biomarker webs, are typically omitted.
46 Therefore, molecular epidemiology is contingent on how well the data reflect the events
47 and involve the risk factor that induced the AO pathways (AOPs). One well-known
48 example in molecular epidemiology is human exposure to benzo[a]pyrene (BP) and the
49 resulting cancer risk. The glucuronide conjugate of BP-7,8-dihydrodiol in urine can be
50 measured as the exposure biomarker. The further metabolism of BP-7,8-dihydrodiol-
51 9,10-epoxide combination with DNA bases, repairing mutation initiation, and
52 accumulation of urinary 8-hydroxy-2-deoxyguanosine can mark an adverse effect and
53 may cause a cancer diagnosis. However, these biomarkers alone cannot profile a
54 pathway to determine the extent to which environmental risk leads to disease
55 development. Systems biology and molecular epidemiology have combined to
56 introduce the systems epidemiology era. The details of molecular events in response to
57 both environmental exposure and disease or adverse health outcomes can be profiled.
58 These molecular events form a biological continuum on the data of genomics,
59 proteomics, and metabolomics. High-throughput data acquisition has been sufficiently
60 applied for omic-wide associated studies, including effect-oriented exposome-wide
61 and/or metabolome-wide studies. Although exposure and outcome molecular
62 biomarkers have been more recently screened using a holistic view instead of a
63 reductive one, the pathway reconstitution on systems biology, i.e., uncovering the black
64 box between the exposure and AO, has not yet been fully considered. In this review, we
65 firstly profiled the ongoing development of meet-in-metabolite analysis (MIMA) for
66 the AOPs' reconstitution on metabolism (Wu et al., 2021), and then discussed the
67 analytical choke-point of low level molecules in exposomics and metabolomics.
68 Because the latter would hinder the quality of these multiple omics database in the
69 MIMA procedure, the advances, techniques, developments and outlooks of the unbiased
70 traceability of total internal exposome (i.e., chemical pollutants and metabolites) were
71 majorly reviewed, which are important and forms the foundation of the suggested
72 systems epidemiological paradigm.

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74 **Metabolism-based pathway reconstitution from exposure to disease**

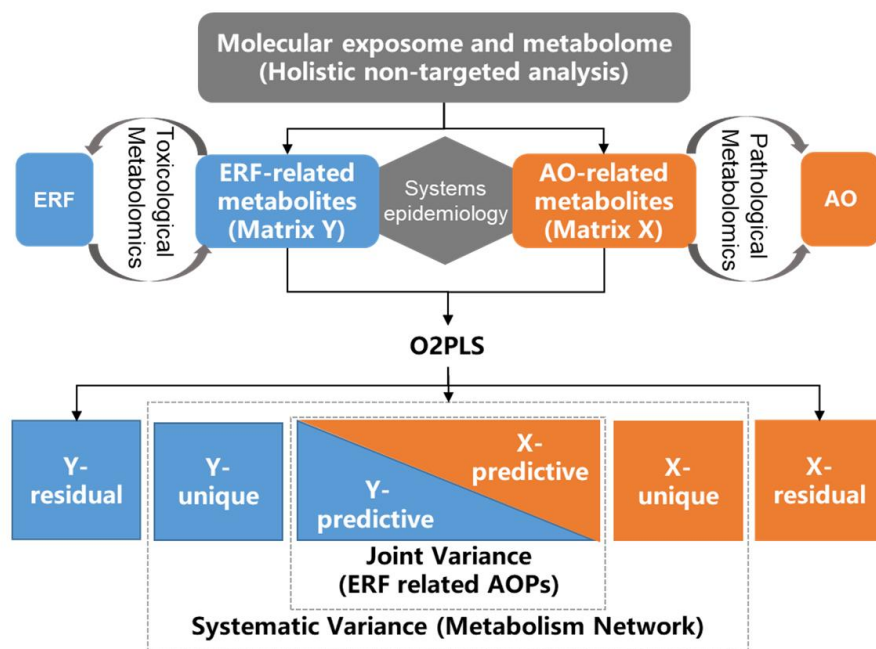
75 Traditional human biomonitoring only focuses on the environmental pollutant
76 measurement in populations, which can include parental chemicals and/or their
77 metabolites. The holistic measurement of endogenous metabolites in humans is
78 typically considered in the context of metabolomics. Non-targeted measurements have
79 involved both endogenous and exogenous small molecules in humans. According to the
80 molecular epidemiology framework, xenobiotic biomonitoring data can relate to
81 metabolic response data. Similarly, a diagnostic outcome may be a reflection of certain
82 metabolic phenotypes. Therefore, it is reasonable to merge the two metabolic modules
83 (endogenous and exogenous) on the metabolome continuum. Pathways linking
84 exposure and AOs can emerge from the complex metabolic web, i.e., the AOP
85 reconstitution. Apart from the omic-wide associated studies, a 'meet-in-the-middle'
86 strategy has been applied in univariate mediation analysis (Chadeau-Hyam et al., 2011;
87 Vineis et al., 2013). MIMA can potentially uncover many hidden molecular events and
88 link exposure to disease or outcome via the metabolic network (Figure 1) (Huang et al.,

89 2018; Huang et al., 2019; Liu et al., 2020b; Wang et al., 2019b; Wu et al., 2018), where
 90 the pathways embedded in the complex metabolomics can be identified. Given arsenic
 91 exposure is the risk of male infertility on the traditional “black box” epidemiology
 92 (Wang et al., 2016), one of the more recent application of the O2PLS model proposed
 93 in Figure 1 has tried to reconstitute the linkages between exposure biology of arsenic
 94 and pathology of male infertility via non-targeted analysis of urinary metabolome (Wu
 95 et al., 2021). In this case-control study, twelve metabolites each have been defined as
 96 the arsenic exposure biomarkers and infertile pathology biomarkers, respectively.
 97 Seven of them can directly bridge the gap between arsenic and infertility. Interestingly,
 98 the core metabolism correlation network further highlighted that testosterone is the vital
 99 hub to transfer the arsenic effects to infertile risk. The network indicated arsenic
 100 methylation that coupled disruption of one-carbon metabolism and oxidation stress and
 101 the adverse effects extended to fatty acid oxidation and steroidogenesis dysfunction.

102 MIMA research can be fundamentally supported by utilizing traditional
 103 epidemiological study designs. However, three aspects are critical to improving these
 104 studies:

- 105 1. Holistic analyses of molecular exposomes and endogenous metabolites.
- 106 2. Reconstructing the global endogenous metabolism network and recognizing the
 107 metabolism modules and hub nodes that respond to environmental factor exposure and
 108 reflect the AO, respectively. Additionally, profiling the pathways and assessing the
 109 biomarkers (with sound sensitivity and specificity).
- 110 3. Conduct causative inferences for biomarkers using mediation or moderation
 111 analysis. The identified biomarkers then return from holism to reductionism because
 112 they are deduced from a global view with fully mechanical information.

113



114

115 **Figure 1: Meet-in-metabolite analysis model: O2PLS application to bridge the**
 116 **gap between environmental risk factors and adverse outcome.**

117 *Note:* Biomonitoring and metabolomics can be integrated into top-down
118 measurements under the MIMA framework and non-targeted analysis in holism is
119 technical foundation of systems epidemiology. Quantitatively and qualitatively
120 annotating endogenous and exogenous analytes and biologically reconstituting system
121 dynamics concerning AOPs or networks are the major goals of AOP-oriented molecular
122 epidemiology in MIMA. ERF = Environmental risk factor; AO = Adverse outcome;
123 AOPs = Adverse Outcome Pathways; O2PLS = Two-way Orthogonal Partial Least-
124 Square; Joint variance are the intersection of metabolite sets that represent both ERF-
125 related and AO-related metabolites, i.e., Y-predictive and X-predictive, respectively.
126 While Y-unique represent the metabolites are independent to AO and X-unique
127 represent the metabolites are independent to ERF, respectively. Y-residual and X-
128 residual are the unexplained parts of the total variations in the model.

129

130 **Genomic-wide associated disease and human exposomics**

131 Genome-wide association studies have revealed many genetic associations and
132 mapped certain networks to improve our understanding of the nature of disease;
133 however, these mappings only account for a small fraction of the disease risks and most
134 parts can be the environment-wide associated (Rappaport and Smith, 2010). Utilizing
135 exposomics to assess all individual exposures in a lifetime and how they relate to
136 disease idealizes exposure measurements in epidemiology (Wild, 2005), emphasizing
137 both systematic and accurate analysis characteristics. The National Institute of
138 Environmental Health Sciences has defined the set of environmental exposures that
139 shifts the human body condition from healthy by chemical exposures, diet, physical
140 activity, stress, pre-existing disease, and the use of addictive substances. Considering
141 human exposure biomonitoring, ethically available biological samples such as blood
142 and urine from the investigated population contain a wide variety of global metabolome
143 information including that of exposures. Based on metabolism, the related exposure
144 responses and health outcomes or disease in cohorts can also be compared (Rappaport
145 et al., 2014; Shen et al., 2014). Mathematically, human body is a cohesive
146 conglomeration of interdependent components that are delineated via both spatial and
147 temporal boundaries. Therefore, metabolomics that encompasses complete internal
148 exposure molecular information and associates a disease can be modeled via MIMA.

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150 **Profiling exposure-oriented disease or adverse outcome on metabolism**

151 Given the limited number of pathways for responding to exposure to various
152 pollutants, exposure biology can be applied to address toxicologically related
153 metabolites; along with the pathological metabolite analysis (Compton et al., 2019),
154 how environmental factors initiate the toxic hazards and transformed into AO can be
155 highlighted, which is vital for environmental disease prevention and medicine. For
156 example, regarding arsenic effects (Zhang et al., 2014c) and male infertility (Huang et
157 al., 2019; Shen et al., 2013; Wang et al., 2016; Zhang et al., 2014a; Zhang et al., 2014b),
158 the pathways of arsenic methylation coupling one-carbon metabolism disruption
159 together with oxidation stress can propagate to the fatty acid oxidation and
160 steroidogenesis dysfunction indicated by testosterone has been recently profiled (Wu et

161 al., 2021). Although the change of testosterone, a key molecular event for male fertility,
162 is not so sensitive and specific to both arsenic-related and male infertility-related
163 metabolites (Wu et al., 2021; Zhang et al., 2014c), MIMA procedure by O2PLS dose
164 has centralized its role in-between arsenic exposure and male infertility. Due to the
165 nonlinear dynamics of biological systems, it is no surprise that not all molecular events
166 are sensitive or specific enough to respond to exposure and/or AOs, which may
167 typically be omitted in the traditional omic-wide associated biomarker investigations.

168 Apart from arsenic, the exposure-oriented metabolism of phthalates,
169 perfluorinated compounds, air pollution of PM_{2.5} have been profiled on human
170 monitoring data of urine, blood or meconium samples, in which male infertility,
171 gestational diabetes mellitus (GDM), chronic obstructive pulmonary disease (COPD)
172 or alteration of cardiorespiratory function have been linked via AO-oriented
173 metabolism (Huang et al., 2018; Huang et al., 2019; Liu et al., 2020b; Peng et al., 2015;
174 Wang et al., 2017; Wang et al., 2019b; Wu et al., 2018). Interestingly, the
175 cardiorespiratory effects of a very low ozone exposure (8.7 ± 6.6 ppb) that associated
176 with changes in metabolic profiles among the vulnerable children can be observed, in
177 which ozone below the current indoor standards was associated with the deteriorated
178 cardiovascular function by disturbing bile acid and endogenous nitric oxide-related
179 oxidation and inflammation, and associated with the exacerbated airway inflammation
180 by reducing GPx-related anti-oxidation (Liu et al., 2021). The result implied that MIMA
181 is comprehensive and powerful for tracing the molecular changes from exposure to
182 outcome.

183 Although metabolome has encompassed all biological aspects of exposure,
184 toxicology and pathology, the analytical aspects are still challenged by the holistic
185 measurement of exogenous and/or endogenous metabolome, it is believed that totality
186 but solitary of the risk factors are decisive in environment related disease pathology.
187 Metabolomics usually can be conducted through high-resolution mass spectrometry
188 (HRMS) with high throughput by non-targeted strategy. However, molecular events that
189 are important for reconstituting the entire pathway from exposure to health outcome
190 may be missed due to the concentration biased metabolome detection. Therefore a
191 systematic analysis without bias towards any metabolite and/or pollutant is required.
192 Additional flexible strategies must be adopted when all small molecules are intended to
193 be monitored in viewpoint of metabolomics. Especially some analytical aspects in
194 exposomics are challenging (Dennis et al., 2017), such as the missing of low level
195 xenobiotics in non-targeted analysis, the identification of unknown small molecules
196 without database, and the metabolomic heterogeneity in different human samples.

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198 **Systematic analysis of human exposure to environmental chemicals**

199 We routinely face a man-made chemical world. Over 145,000 chemicals are
200 registered by the European Union Registration, Evaluation, Authorization and
201 Restriction of Chemical Substances regulation
202 ([https://echa.europa.eu/web/guest/information-on-chemicals/pre-registered-](https://echa.europa.eu/web/guest/information-on-chemicals/pre-registered-substances)
203 substances). A recently released update of the Toxic Substances Control Act (TSCA)
204 inventory list for chemicals that are actively being manufactured by the U.S.

205 Environmental Protection Agency (EPA) as of February 19, 2019, demonstrated that
206 less than half of the total number of chemicals on the current TSCA inventory list, i.e.,
207 40,655 of the 86,228 (47%) chemicals, are currently being sold. Except for a few
208 persistent organic chemicals (POPs), most of the high production volume (Nikfar et al.,
209 2014) chemicals can be considered pseudo-persistent chemicals because of their
210 continuous emission to the environment, even if their half-lives are short (Bergman et
211 al., 2012; Daughton, 2003). These chemicals plus POPs can be a great source of concern
212 if vulnerable subpopulations are exposed to amounts that result in adverse effects
213 (Nicholson et al., 2004). The National Health and Nutrition Examination Survey, a well-
214 known targeted biomonitoring study, includes a few hundred preselected chemicals for
215 the survey. Therefore, it is an enormous gap between the ongoing biomonitoring and
216 the current human exposure to environmental chemicals, which has hindered the
217 exposure risk assessment for evaluating environmental health risks.

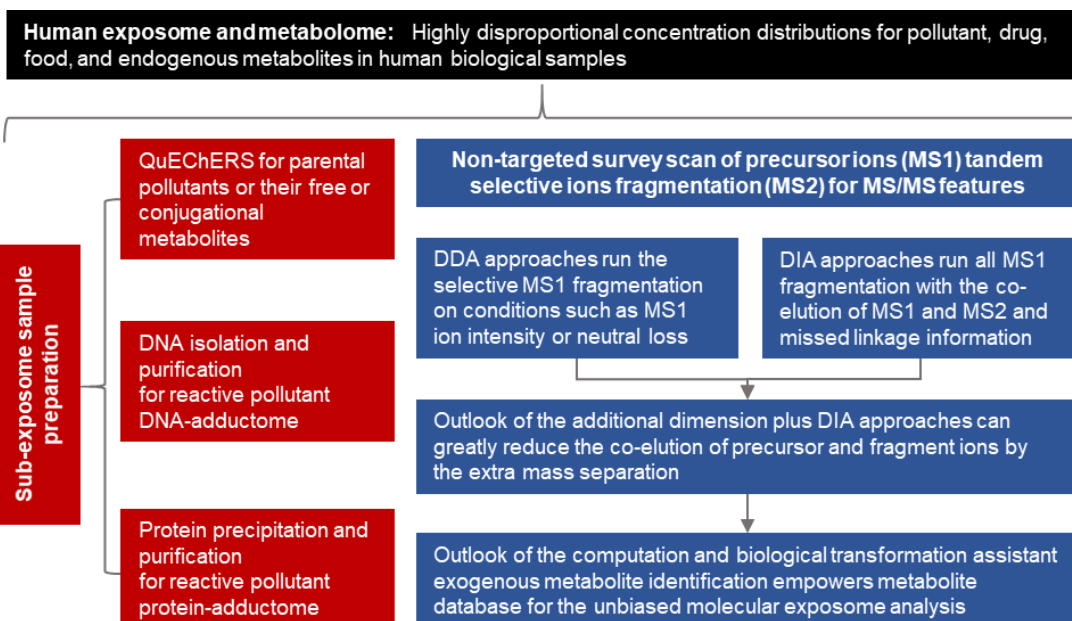
218 The EPA Non-Targeted Analysis Collaborative Trial (ENTACT) has used
219 suspected screening for approximately 1,200 chemical substances from the EPA
220 ToxCast library in house dust, silicone wristbands, and human serum (Ulrich et al., 2019)
221 to systematically assess human exposure to environmental chemical pollutants.
222 Approximately 40,153 to 114,100 small molecules are populated in human body
223 (Wishart et al., 2018), with only 5,835 reconstructed by the endogenous human
224 metabolism (Angione, 2019). It has been estimated that the four major small-molecule
225 categories, namely endogenous chemicals, food chemicals, pollutants, and
226 pharmaceuticals in blood, exhibit concentrations that cover a 10^7 -fold range. One of the
227 greatest challenges to screening and identifying unknown pollutants in human samples
228 is their minuscule concentrations, which are typically thousand times lower (median of
229 2.4×10^{-4} μM) than those of endogenous chemicals (0.94 μM), food chemicals (1.00 μM),
230 and pharmaceuticals (0.30 μM). Given that untargeted HRMS can detect >30,000 small
231 molecule features in human serum (Ivanisevic et al., 2013), a platform bias results in
232 measurement uncertainties for small molecules less than approximately 0.1 μM in 50-
233 L of serum, where approximately 90% of pollutants and 30% of endogenous and food
234 chemicals have been missed regarding the data-dependent acquisition of fragment ions
235 (MS2). This is because only the top 10 most intense precursor ions (MS1) are subjected
236 to MS2 (Wang et al., 2019a; Yan and Yan, 2015). When an expected system
237 biomonitoring model (such as that of a top-down exposome (Rappaport, 2011) is
238 applied, more flexible strategies should be applied to overcome the challenges in the
239 nowadays non-targeted holistic analysis (Figure 2), such as data-dependent acquisition
240 (DDA) approach.

241 Because of there is a huge concentration differences among the much lower
242 pollutants and the relatively high endogenous chemicals, food chemicals, and drugs,
243 respectively (Ivanisevic et al., 2013; Rappaport, 2014), results in many challenges to
244 apply the holistic measurement to molecular exposome. DDA approaches can only
245 select a few fractions of the most-intense (such as the top 10) mass spectra (MS) of
246 precursor ions (MS1) subject to fragmentation (MS2) (Yan and Yan, 2015), which is
247 the substantial bias for searching less abundant but biological important molecules in
248 metabolomics. In addition to the commonly used DDA, the data-independent

249 acquisition (DIA) can theoretically address all molecules with MS/MS features,
250 however, MS1 and MS2 ions might co-elute and difficult to identify the parent-daughter
251 linkages for small molecules when compared to for proteins. As a result DIA of MS2
252 covers a broad range of precursor ion fragments and the link between MS1 and MS2
253 might get often lost. With the recent development of DIA protocols such as scanning
254 sequential window acquisition of all theoretical spectra (SWATH) (Raetz et al., 2020),
255 a novel DIA method SONAR (Juvvadi et al., 2018), and ion mobility MS/MS (IM-
256 MS/MS) (Zheng et al., 2017; Zhou et al., 2018), MS1 selectivity have been greatly
257 increased. The extra dimension for mass separation in the new DIA protocols of
258 scanning SWATH, SONAR and IM-MS/MS can help to reconstitute MS1/MS2 parent-
259 daughter linkages; then the enhanced unbiased measurement for particularly low
260 abundance metabolites in non-targeted analysis can be readily obtained (Figure 2).

261 The challenge still remains to identify non-targeted small molecules. With only
262 MS/MS features, many pollutants cannot be identified due to the database searching
263 limitations. Currently, over 114,100 metabolites have been listed in the Human
264 Metabolome Database (version 4.0) (Angione, 2019), but few are metabolites from
265 xenobiotics. For example, meconium metabolites of methylepicatechin,
266 methylxanthine, dimethyluric acid and vanilloylglycine are exogenous and are
267 commonly present in green teas, red wine, cocoa products, and many fruits, which have
268 been associated with GDM risk (Peng et al., 2015). Computational MS (Feunang et al.,
269 2016) for small molecules should be further developed in combination with their
270 biological transformations, such as the in-molecule glucuronide feature (Tsugawa et al.,
271 2019; Walmsley et al., 2019). These kinds of features are commonly occurring for the
272 metabolites of xenobiotics for assisting their excretion via urine and/or bile pathways
273 but then computational-based exogenous metabolite molecular structure reconstruction
274 (Scheubert et al., 2013) can become rational and will be applied to molecular exposome
275 analysis. Considering sample preparation methods, sub-grouping strategies can be used
276 to improve the non-targeted analysis of human exposome in the model of metabolome
277 because after extraction, enrichment and purification different xenobiotics can be
278 classified and collected by their properties such as lipophilicity, hydrophilicity or
279 formation of adducts to other biomolecules (Figure 2).

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Figure 2. Suggested perspectives for a holistic non-targeted analysis of human molecular exposome.

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Note: Various levels of small molecules of xenobiotics can be hidden in human metabolome and requires an integral strategy from initial sample preparation to final analytical method selection, which improves systems thinking of the aspects of analytical challenges in molecular exposomics.

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Heterogeneity of environmental chemicals and exposure biomarkers in humans

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Monitoring human exposure to suspected environmental chemicals using internal biomarkers has been accepted as the most accurate exposure assessment approach (Barr et al., 2005; Smolders et al., 2009). Internal chemicals and/or their metabolites should constitute the most abundant fractions and predominately represent the external exposure to be considered suitable biomarkers, where the measured analytes exhibit the sound specificity and sensitivity is also required to respond to their exposure. After entering the human body, xenobiotics are metabolized via the pronounced phase-I and phase-II reactions in the liver and kidney (or by gut microbiota when exposure via diet). In addition, certain exogenous chemicals can be transformed into the activated derivatives of protein and DNA adducts, and some adducts may initiate the immune system response, in which the xenobiotic-protein adducts can form antigen-antibody complexes (Pallardy and Bechara, 2017). Chemical pollutants have been categorized as POPs, non-persistent organic chemicals, bioaccumulative metals, non-bioaccumulative metals, and others (Smolders et al., 2009). Analyte can either surrogate the chronic exposure when a chemical is persistent or accumulated, or they can represent the acute exposure when it is reactive or non-bioaccumulative. Reactive chemical metabolic fractions excreted via urine may indicate acute exposure, or they may be classified as chronic exposure biomarkers when combined with proteins. Although few cases have been compared for environmental chemicals, the use of fasting plasma glucose (free blood glucose level) as acute biomarker and the hemoglobin A1c adduct (HbA1c) as chronic one have been well-documented (Association, 2017). Finally, the relative

311 abundance and time-dependent variation for the reactive chemicals as antigen-antibody
312 complexes have not yet been addressed for biomonitoring.

313 Exposure biomarker selection is also subject to monitoring and molecular
314 exposure life stages (Dennis et al., 2017). This is because the toxicokinetics (Liu et al.,
315 2020a) and ethical convenience of sampling are development-specific. Various human
316 sample types have been used as biomonitoring matrices (Smolders et al., 2009; Pallardy
317 and Bechara, 2017). Placenta, cord blood, and meconium (Huang et al., 2019) samples
318 are utilized to assess perinatal exposure, which may cover all chemical types. For
319 example, meconium is formed by the fetus as early as the 12th week of gestation and
320 accumulates until birth; it is a repository of endogenous and exogenous agonists and
321 metabolites, is capable of capturing disease-relevant metabolic profile changes and
322 identifying novel biomarkers (Peng et al., 2015). However, the use of meconium must
323 be further explored because the extraction for many organic pollutants in this matrix
324 exhibits more difficulties than in placenta and cord blood. Sampling from newborns and
325 infants is more difficult than sample collection from other developmental stages;
326 therefore, the diaper urine and blood spots must be deeply mined for biomonitoring.
327 For the two most convenient types of samples, blood and urine, persistent or
328 bioaccumulative chemicals are likely found in the blood, while the reactive and non-
329 bioaccumulative ones are likely to be concentrated in urine in their free forms as
330 parental or their metabolites.

331 Subject to the chemical type, POP biomonitoring has been well-documented for
332 targeted analytical strategies; the related biomarkers are typically the parental chemicals
333 in serum samples (Smolders et al., 2009; Pallardy and Bechara, 2017). Certain
334 persistent organohalogen compounds (OHCs) may be natural chemicals (Agarwal et al.,
335 2014; Teuten et al., 2005). However, most of them are man-made, and a systematic
336 analysis of these trace levels of POPs in humans can be conducted using both target and
337 non-target approaches, such as gas chromatography (GC)-HRMS for semi-volatile and
338 lipophilic POPs (Alonso et al., 2017; Goto et al., 2020; Yang et al., 2019) and liquid
339 chromatography-HRMS for hydrophilic POPs such as perfluoroalkyl and
340 polyfluoroalkyl substances (PFASs) (Concha-Grana et al., 2018; Ruan and Jiang, 2017).

341 The exposure biomarker selection for reactive pollutants is more challenging to
342 acquire than that for persistent ones. Reactive chemicals and their metabolites may
343 occur in free forms in metabolites through derivatization via functional groups such as
344 -OH, -SH, -NH₂, and -COOH, and/or by conjugation with glucuronic acid, sulfate,
345 glutathione, or acetyl. Free metabolites and small molecular conjugates are typically
346 concentrated in urine and are ready for excretion. Therefore, they are typically selected
347 as biomarkers to assess acute exposure to reactive pollutants. The electrophilic activated
348 chemicals can also covalently adduct to nucleophilic sites in proteins (such as the
349 sulfhydryl group of cysteine, ε-amino group of lysine, and imidazole group of histidine
350 in proteins (Shibata and Uchida, 2019) in hemoglobin and albumin) and DNA (Cooke
351 et al., 2018; Guo and Turesky, 2019). When measuring these derivatives in adductomics
352 (Rappaport et al., 2012), hemoglobin adducts and human serum albumin (HSA) are
353 preferable to those of DNA and glutathione for characterizing chronic exposure because
354 of their greater abundance and longer residence times in human blood (Needham and

355 Sexton, 2000). Consequently, these circulating protein adducts can be used as chronic
356 exposure biomarkers for reactive chemicals. In addition to free protein adducts,
357 circulating antigen-antibody complexes (Aguirre-Gamboa et al., 2016; Sharma et al.,
358 2017) may pool the pro-/pre- or haptenic xenobiotics for immunity-oriented assessment.

359 Metal biomarkers are comparable to organic chemicals, including elements or their
360 metabolic species in urine and blood samples. The common approach to address metal
361 speciation in biosciences is called metallomics, and various metals and species can be
362 sensitively measured by LC-inductively coupled plasma (ICP)-MS. Urinary and serum
363 samples are typically digested and measured by ICP-MS or ICP-optical emission
364 spectroscopy, where the elements can be applied as biomarkers to indicate their total
365 exposure status (Medda et al., 2016; Troisi et al., 2019). Certain metal or metalloid
366 species, such as methylated arsenic and mercury, can be measured in urine and blood
367 samples by LC-ICP-MS. Many small (oxalate, citrate, tartrate, amino acids, and
368 oligopeptides) and large (nucleic acids, polysaccharides, and proteins) biomolecules
369 can bind to metals, generating various chemical species (Lopez-Barea and Gomez-
370 Ariza, 2006). Similar to organic chemicals, metals can occur in free forms, small
371 metabolic species, or metalloproteomic forms with proteins (Coverdale et al., 2019).
372 For example, serum albumin is a highly abundant plasma protein associated with the
373 transport of metal ions (Curry et al., 1998); therefore, HSA can be used in the exposome
374 proteomic forms to metals (Curry et al., 1998; Varshney et al., 2010). The inducible
375 metallothioneins are Cys-rich metal-binding proteins, and they can be used as both
376 exposure and stress biomarkers for metal exposomics (Coverdale et al., 2019).

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378 **Screening strategies for parental chemicals and their free form metabolites**

379 Monitoring pre-selected target chemical pollutants (including their demarcated
380 metabolites) in various human samples and their applications throughout the life stages
381 has been sufficiently performed (Smolders et al., 2009). In clinical and forensic
382 toxicological fields, the parental chemicals or their defined metabolites can be screened
383 in the pre-selected list as suspected unknowns in human samples. The typical strategy
384 is the chemical parental structure-dependent screening. For the suspected unknown
385 OHC screening, sample preparation procedures and GC-MS analysis have determined
386 chemicals of neutral, lipophilic, and semi-volatile properties, which exhibit well-
387 demarcated chemical subclasses and can match a holistic screening. With the assistance
388 of sample clean-up and target concentration, non-target OHC screening was performed
389 using a full-scan screening on the GC×GC-HRToFMS (resolution power >5000 with
390 mass error ± 10 ppm) and qualitative analysis by GC-magnetic-sector HRMS
391 (resolution power 8000-12000 with mass error <10 ppm) (Goto et al., 2020), and over
392 300 OHCs were identified in marine dolphins (Alonso et al., 2017). Over 4,000 PFASs
393 have been suspected to enter the environment (Ritscher et al., 2018). PFASs contain at
394 least one perfluoroalkyl moiety (C_nF_{2n}), this common feature is useful for screening
395 their human exposure. The systematic measurement of PFASs (Shibata and Uchida,
396 2019) is urgent and can be acquired using LC-time-of-flight or Orbitrap HRMS, which
397 is equipped with electrospray ionization for ionic PFASs and atmospheric pressure
398 photoionization for neutral PFASs. The proportions of unidentified organofluorines

399 rose from approximately 20% to 50% in German plasma samples collected during
400 2000-2009, indicating human exposure to various unknown PFASs (Yeung and Mabury,
401 2016). Holistic screening can be performed on the perfluoroalkyl moiety $-C_nF_{2n}-$
402 through a combination of criteria such as mass balance (Shibata and Uchida, 2019).
403 Regarding the in-molecule diagnostic features of $[C_2F_5]-$ (m/z 118.992) and $[C_3F_7]-$
404 (m/z 168.988), low levels of unknown C5-C17 poly- and perfluoroalkyl substances
405 have been identified in water (Liu et al., 2015). The data requirements include both
406 DDA and DIA modes in PFAS HRMS target and non-target analyses. These strategies
407 can also be used for other chemicals.

408 The pre-selected suspected unknown screening for free reactive chemicals and
409 their metabolites can be made available when the molecular structures of the biomarkers
410 in blood or urine have been documented, in which the unknown screening procedures
411 (Plassmann et al., 2015) can be referenced with additional sample preparation
412 procedures. For example, the quick, easy, cheap, rugged, and safe preparation method
413 can be used to extract compounds covering a broad domain for GC or LC-MS analysis
414 (Perestrelo et al., 2019). To extend the suspected unknown list, the ENTACT has
415 attempted to advance xenobiotic chemical analysis in environmental and biological
416 media by using the ToxCast library of chemical substances, DSSTox database, and
417 CompTox Chemicals Dashboard (Ulrich et al., 2019). However, free reactive chemicals
418 and their metabolites may only account for the small fractions of these pollutants, and
419 most of them should have been biologically transformed. Thus, the pollutant dependent
420 screening may only be suitable for the small subclasses of POPs.

421

422 **Screening strategies for biologically transformed environmental chemicals**

423 For reactive chemical pollutants, the metabolism biological basis involves
424 converting lipid-soluble, non-polar, and non-excretive forms to water-soluble and polar
425 forms that are excretive in bile and urine. Their metabolites may exist in free forms (by
426 exposing or adding functional groups) or in conjugated forms (to large water-soluble
427 and charged endogenous molecules), all of which are ready for excretion via urine, and
428 their urinary concentration can indicate acute exposure. Certain fractions may also form
429 adducts with DNA and proteins and exist in the blood circulation; their half-lives
430 communicate the related macromolecule degradation and can therefore indicate certain
431 chronic exposure types. The identification of a reactive chemical that
432 adducts/conjugates to an endogenous biomolecule can be assisted by using in-molecule
433 diagnostic features (i.e., the endogenous sub-structures) (Plassmann et al., 2015).

434 *Urinary sub-exposome of conjugates:* Xenobiotic reactions primarily occur in the
435 liver (hepatocytes) and occur less frequently in the kidney (proximal tubular cells),
436 lungs (Clara cells and alveolar cells), intestines (mucosa lining cells), skin (epithelial
437 cells), and testes (seminiferous tubules and Sertolis cells). Molecular sub-exposome of
438 urinary metabolites of the four major phase II metabolism conjugations (glucuronic acid,
439 sulfate, glutathione, and acetyl) can be separately profiled, where the sulfates,
440 glucuronide, acetyl, and mercapturic acids (Frigerio et al., 2020; Yao et al., 2016) may
441 be used as in-molecule flagging features for the systemic unknown screening,
442 respectively. In addition, the human metabolome (Walker et al., 2019) and related

443 databases can be used to aid the screening process. Because of the mass spectral
444 acquisition in parallel modes cycles back and forth, where the MS1 of the full scan with
445 ultrahigh resolving power (such as RP = 120,000, mass accuracy ≤ 3 ppm) and the MS2
446 with in-source fragmentation flagging scans to flag the suspected target retention times
447 using diagnostic fragments, the lower xenobiotic concentrations can be theoretically
448 addressed. However, challenges remain for computation-based chemical identification
449 (Bocker and Duhrkop, 2016; Dührkop et al., 2013; Hufsky et al., 2014; Ruttkies et al.,
450 2016; Scheubert et al., 2013).

451 *Sub-exposome of adducts:* When the reacted xenobiotic species are trapped by
452 macromolecules in the targeting tissue or circulation system, reactions can occur at the
453 genome or proteome scale. To improve the analytical efficiency, protein and DNA
454 sample preparation and purification can be applied before their digestion. Similar
455 platforms in metabolome analysis can then be used for adducts. LC-MSⁿ-based DNA
456 adductomic investigations can utilize a common structural feature of
457 deoxyribonucleosides, in which a deoxyribose moiety bound to the nucleobase through
458 a glycosidic bond (Balbo et al., 2014). The DIA wide selective ion monitoring/MS2
459 methodology (Guo et al., 2017) with HRMS can detect many DNA adducts through
460 non-targeted screening and computational data analysis. Reactive organic pollutants
461 most often bind covalently through their electrophilic properties to react with proteins
462 when they are trapped by circulating proteins (Li et al., 2011b), in which the
463 nucleophilic protein sites added to electrophilic toxicants. Many reactions can be
464 observed in hemoglobin and HSA or toxic-targeting DNA (Kanaly et al., 2006). An
465 adductomic strategy can then be directly applied for the suspected xenobiotic
466 biomonitoring. For proteins and nucleic acids that are biologically degraded, the
467 embedding xenobiotics may be measured as metabolites along with metabolome scale
468 reactive chemicals (Li et al., 2011a; Ma and Subramanian, 2006).

469 Circulating antigene-antibody complexes may serve as another sub-exposome
470 protein pool for reactive chemicals. The innate and adaptive immune systems have been
471 developed by living organisms to protect them from “outside” viruses, bacteria, and
472 parasites. Thus, “outside” chemicals can be translated into innate immune system
473 activation, which may be the consequence of different key steps that allow dendritic
474 cells (DCs) to initiate immune system adaptation (Association, 2017). The human
475 immune system is highly reactive to the environment, and 80% of the measured
476 immunological parameters are affected by the environment ($\geq 50\%$ variance) (Brodin et
477 al., 2015). Many environmental chemicals, acting as haptens, can bind to HSA and
478 cause the immune system to misidentify self-tissue as an invader and launch an immune
479 response against it (autoimmunity) (Vojdani et al., 2015). Apart from the liver and other
480 metabolic organs, innate immune cells such as monocytes, macrophages, dendritic cells,
481 and polymorphonuclear cells play a fundamental role in xenobiotic metabolism.
482 Xenobiotic metabolic conversion was observed in dermal Langerhans cells (immature
483 DCs), containing cytochrome P4501A (CYP1A) enzymes. After conversion by the
484 CYP1A enzyme, xenobiotics form a complex with self-proteins, which are processed
485 and presented as major histocompatibility complex class I and class II molecules. The
486 xenobiotic-protein complex presented by antigen-presenting cells is subsequently

487 recognized by T cells, which help B cells with antibody production. The conjugation of
488 xenobiotics to self-proteins makes them highly immunogenic and therefore elicits the
489 production of anti-xenobiotic antibodies, which play a physiological role in clearing
490 xenobiotics from the body (Association, 2017; Sajid and Agrewala, 2019).

491 Many heavy metal pollutants such as mercury, nickel, and cobalt can react with
492 oxidized proteins to form protein metal chelate complexes (Sajid and Agrewala, 2019).
493 These xenobiotics may be indirectly measured via antigen-antibody complexes, which
494 are multi-molecular complexes that are typically stabilized via the reversible interaction
495 of static electricity, hydrogen bonds, or the van der Waals force. These small pollutants
496 directly bind to self-proteins or bind indirectly after hepatic or extrahepatic conversion
497 from prohaptens to haptens, generating hapten-protein adducts (Sajid and Agrewala,
498 2019). When comparing free HSA, circulating antigen-antibody complexes (Sharma et
499 al., 2017; Medda et al., 2016), such as the aflatoxin-HSA adduct, which can elevate IgG
500 and/or IgM (Vojdani et al., 2003), may be concentrated, and xenobiotic pooling should
501 be investigated through immunoprecipitation using HRMS. For example, MS methods
502 revealed that an extreme sensitizer 2,4-dinitro-1-chlorobenzene, and
503 methylchloroisothiazolinone modified a greater number of nucleophilic HSA sites than
504 the moderate sensitizer cinnamaldehyde. However, the weak/non-sensitizer 6-methyl
505 coumarin was restricted to a single nucleophilic residue within HAS (Esser and Martin,
506 2017).

507

508 In summary, the metabolome encompasses all exposure burden information and
509 exhibits greater potential to generate data to enhance exposure assessment regarding
510 exposomics than that of any other omics research. In addition, lifetime dimensional
511 information can be obtained by repeatedly measuring ethically available life-staged
512 blood and urine. From the human bio-monitoring perspective, these biological samples
513 have convened all possible global postnatal information on the investigated population,
514 including chemical exposure, biological response, and the potential linkage to a defined
515 health outcome. Regarding maternal blood and urine during pregnancy, partly attributed
516 'prenatal' exposure cord blood, placenta, and meconium may ethically be available to
517 characterize fetal prenatal conditions. Finally, the metabolome is useful for conducting
518 biology-based estimates of individual and public health risks.

519

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