

1 **Title**

2 Mature beech and spruce trees under drought – higher C investment in reproduction at the
3 expense of whole-tree NSC stores

4

5 **Running head**

6 NSC concentrations and pool sizes in mature trees under drought

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35 **Key words**

36 BALANCE, biomass increment, *Fagus sylvatica*, *Picea abies*, non-structural carbohydrates,
37 carbon balance

38 **Highlights**

- 39 • A mixed beech/spruce forest was exposed to two years of recurrent summer drought.
- 40 • Under drought, NSC concentrations remained unaffected among all tree organs
- 41 • Total NSC pools size decreased by about 40% in both species under drought
- 42 • Unlike growth, investment in reproductive tissues remained stable under drought
- 43 • Carbon starvation was not evident in mature trees under two years of drought

44 **Abstract**

45 Under drought, potential plant death from depletion of carbon (C) stores, referred to as carbon
46 starvation, is thought to result from negative carbon balance during (partial) stomatal closure.
47 As evidence for C starvation is rare for mature trees, we investigated the C dynamics in mature
48 beech and spruce under drought, focusing on non-structural carbohydrates (NSC) as an
49 indicator of their C balance. Trees were exposed to complete exclusion of precipitation
50 throughfall over two consecutive growing seasons. We assessed NSC concentrations during the
51 early dormant season in leaves, twigs, stem phloem/xylem, coarse and fine roots. Up-scaling to
52 whole-tree NSC pools was achieved using a process-based single-tree model (“BALANCE”),

53 estimating tree biomass increment. While there were distinct differences in NSC concentration
54 among different tissues in both beech and spruce (root < stem < twig < leaves and xylem <
55 phloem), drought did not affect NSC concentrations. However, compared to controls, the
56 whole-tree NSC pool size significantly decreased under drought in both beech (42 %) and
57 spruce (36 %), in parallel to a significant growth decline of overall 52 % and 57 %, respectively.
58 Nevertheless, drought-stressed beech and spruce invested almost twice as much C in
59 reproductive structures relative to total C investment (i.e. 6.0 ± 3.3 and 52.3 ± 8.71 %) compared
60 to control trees (3.1 ± 1.8 and 29.2 ± 7.8 %). This highlights the high priority of C investment
61 into reproduction relative to growth under drought. Given that NSC concentrations are
62 maintained even under severe drought over two growing seasons, NSC pool sizes appear to be
63 a better proxy to assess whole-tree's carbon status in mature trees. Overall, trees maintained
64 NSC availability, avoiding carbon starvation, by downregulating a major C sink (i.e. growth)
65 while upholding reproduction.

66 **1. Introduction**

67 During drought, plants are facing several difficulties with the lack of water and decreased
68 carbon (C) gain as severe consequences (McDowell 2011). Under harsh drought conditions,
69 these can lead to hydraulic failure and/or carbon starvation (Sevanto et al., 2014) and eventually
70 kill the plant. The water conducting system will be disrupted as conductive elements, i.e. vessels
71 and tracheids, begin to cavitate and water transport is limited, resulting eventually in hydraulic
72 failure (Barigah et al., 2013; Urli et al., 2013). With less water being transported to the leaves,
73 stomatal conductance is often reduced and therefore less C can be assimilated (Brodribb and
74 McAdam, 2011). Plants can follow different strategies to deal with drought. Isohydric plants
75 close their stomata early under drought to minimize water loss and avoid hydraulic failure at
76 the risk of carbon starvation, while anisohydric plants keep their stomata open to maximize C
77 gain, avoiding carbon starvation at a higher risk of hydraulic failure (McDowell et al., 2008).
78 With ongoing climate change, drought periods and heat waves will occur more frequently
79 (Burke et al., 2006; IPCC, 2014), and such stressful conditions will be a major challenge for
80 ecosystems, especially for long-living and slow growing organisms such as trees (Hartmann et
81 al., 2018). Consequences of such extreme events for Central European forests have been
82 obvious upon extreme drought years in 2003, 2015 and 2018/19 (Asner et al., 2016; Büntgen
83 et al., 2021; Schuldt et al., 2020) with severe decreases of photosynthesis and primary
84 production (Ciais et al., 2005), as such the C balance of trees could be critically disturbed

85 (Anderegg et al., 2012). During extended drought stress, C stores may become depleted and the
86 plant could be limited in its anabolism (Palacio et al., 2014).

87 Carbon gained by photosynthesis is needed for many catabolic and anabolic processes. NSCs,
88 consisting mainly of soluble sugars (e.g. sucrose) and starch are invested in five predominant
89 sinks: storage, metabolism (e.g. growth), transport, osmoregulation and export (e.g. exudates,
90 Hartmann and Trumbore 2016). Under drought, these sinks will be affected in different ways.
91 One of the first consequences is a decrease of the photosynthetic rate accompanied by a
92 depletion of C stores (Maguire and Kobe, 2015). The export of metabolites, e.g. root exudates
93 or volatile organic compounds, also changes under drought (Gessler et al., 2017; Rennenberg
94 et al., 2006). In mature trees, transport of photoassimilates and water, which are highly
95 connected, are strongly decreased under drought (Adams et al., 2017; Hesse et al., 2019;
96 Sevanto, 2018) and therefore C sinks at a longer distance from the primary source (i.e. leaves)
97 could be affected stronger under whole-tree C limitation (Landhäusser and Lieffers, 2012).
98 Especially in mature trees with long transport distances, this could potentially lead to a C
99 imbalance along the tree under drought, with fine roots (longest transport distance from the
100 source organs) being affected stronger than e.g. branches (short transport distance, Ryan and
101 Asao, 2014). So far, most experiments on C allocation at the whole plant level under drought
102 were carried out on juvenile trees (e.g. Hagedorn et al. 2016, Chuste et al. 2020) or under short-
103 term drought (Anderegg et al., 2012). In this paper, we focus on long-term drought effects (i.e.
104 two subsequent growing seasons) on the primary metabolism, in particular NSC, and its
105 consequences for growth and investment in reproduction in a mixed forest with mature trees of
106 Norway spruce (*Picea abies* (L.) H Karst) and European beech (*Fagus sylvatica* L.). The two
107 species follow a rather opposing strategy, when it comes to stomatal sensitivity and regulation
108 under drought (Pretzsch et al., 2014). The degree of isohydry can be seen as continuum with
109 two extreme ends (Hartmann et al., 2021; Hochberg et al., 2018). Strictly isohydric species
110 close their stomata very early under drought stress and try to avoid a sudden drop in water
111 potential, at the expense of a limited CO₂ uptake. Conversely, more anisohydric species keep
112 their stomata open even during severe drought conditions to further assimilate CO₂ at the
113 expense of continued water loss (Hochberg et al., 2018). Following the literature and own
114 observations at the experimental site of this study (Pretzsch et al. 2014), mature spruce is
115 generally behaving more isohydric under drought (Lyr et al., 1992; Oberhuber et al., 2015;
116 Pashkovskiy et al., 2019), although some variability in provenances exist in seedlings
117 (Jamnická et al., 2019). Conversely, mature beech follows a more anisohydric drought strategy
118 (Leuschner, 2020; Magh et al., 2019). We assessed NSC concentrations of different tree organs

119 and tissues during winter early dormancy. We took advantage of the KROOF experiment, a
120 long-term drought experiment, with retractable roofs to simulate summer drought (Grams et al.
121 2021). By means of the process-based single-tree model BALANCE (Grote and Pretzsch, 2002;
122 Rötzer et al., 2010), we estimated the growth increments of tree's organs allowing for an
123 upscaling of NSC concentrations to estimate overall NSC pools to test the following hypothesis:

124 H1. NSC concentration is reduced by drought, with concentrations in sink organs being
125 more affected than in source organs and with distance from source organs, i.e. leaves >
126 branches/twigs > stems > roots.

127 H2. Total NSC pools, measured during dormancy, are reduced in drought stressed trees,
128 a. with the more isohydric spruce being stronger affected than the more
129 anisohydric beech and
130 b. with sink organs being stronger affected than source organs, i.e. leaves >
131 branches/twigs > stems > roots

132 due to decreases in NSC tissue concentration and growth.

133 H3. The amount of invested C in reproductive structures decreases under drought.

134 **2. Material and methods**

135 *2.1 Plant material and experimental design*

136 The experimental site, a mature mixed forest composed of European beech (*Fagus sylvatica*)
137 and Norway spruce (*Picea abies*) planted around 1931 AD and 1951 AD, respectively, is part
138 of the Kranzberg Forest Roof (KROOF) experiment in Southern Germany near Freising,
139 Bavaria. In 2010, 12 plots were trenched to one-meter depth, allowing no further lateral water
140 transport between the plot and outside soil (Pretzsch et al., 2016). Each plot contained between
141 three to seven beech and spruce trees each (Grams et al., 2021). The KROOF experiment started
142 in 2014 when six of the 12 plots were covered with a unique rainfall-activated retractable roof,
143 which closed automatically for precipitation events during the growing season, to exclude any
144 throughfall as well as stem run-off and retracted after the rain events passed, to minimize any
145 other climatic effects. For further details on the experimental design, site (e.g. soil
146 characteristics) and treatment description see Grams et al. (2021). Data presented in this
147 publication were collected in 2013, 2014 and 2015 from two beech and spruce trees of eight
148 plots accessible via canopy crane (four assigned to the throughfall exclusion (TE) plots and four
149 control plots (CO)) to a total of eight TE and CO trees for each species.

150 The long-term average in temperature is 13.8 °C during the growing season (7.8 °C for whole
151 year) and 750-800 mm annual precipitation. For 2013 and 2014, the precipitation was 640 mm
152 (growing season: 444 mm) and 651 mm (growing season: 501 mm), a bit lower than the long-
153 term average. At the same time, growing season temperature was slightly higher than average,
154 i.e. 14.8 °C and 15.2 °C in 2013 and 2014, respectively. With only 495 mm (growing season:
155 330 mm) of precipitation and a growing season temperature of 17.2 °C, the year 2015 was
156 extraordinarily warm and dry. For further details see Grams et al. (2021).

157 *2.2 Drought stress assessment*

158 To assess the effects of the throughfall exclusion on the trees' water status, pre-dawn leaf water
159 potential (Ψ_{PD} in MPa) was measured with a Scholander pressure chamber (mod. 1505D, PMS
160 Instrument CO., Albany, OR, USA) towards the end of the drought period in the middle of
161 October of each year. Volumetric soil water content (SWC in vol.-%), was assessed weekly by
162 Time Domain Reflectometry (TDR, TDR100, Campbell Scientific, Logan, CT, USA) in four
163 different depths (0-7 cm, 10-30 cm, 30-50 cm and 50-70 cm) and at three positions within each
164 plot (total of 144 TDR probes, for details see Goisser et al. 2016). Data were averaged over all
165 four depths and three positions for each plot.

166 *2.3 NSC sampling and concentration assessment*

167 NSC sampling was done in early November 2013 to 2015 at the start of winter dormancy, in
168 order to get the carbon status at the end of the vegetation period, hence one year before the
169 onset and two consecutive year/s of experimentally induced summer drought. For each sample,
170 the sum of soluble carbohydrates (SC, Table S1) and starch (Table S2) was assessed. Via the
171 canopy crane, one-year old leaves of spruce (SC_L & $starch_L$) and branches/twigs of beech and
172 spruce were sampled. For the branches/twigs (2-3 years old), the living bark (inner and outer
173 bark; hereafter referred to as phloem, SC_{TP} & $starch_{TP}$) and xylem (SC_{TX} & $starch_{TX}$) were
174 separated. With an increment borer and a cork borer, the xylem (SC_{SX} & $starch_{SX}$) and phloem
175 (with the dead bark removed, SC_{SP} & $starch_{SP}$) of the stem were collected at 1.3 m above the
176 ground, respectively. For NSC measurements in the xylem, the last 10 years of each increment
177 core were used, combining the conductive and non-conductive portions of the xylem.
178 Additionally, wood chippings of coarse roots (average diameter about 5 cm, SC_{CR} & $starch_{CR}$)
179 were sampled with a driller and fine roots (with a diameter of less than 2 mm, SC_{FR} & $starch_{FR}$)
180 of both species were directly sampled. All samples were stored on ice (0 °C) and microwaved
181 for 30 s within 30 min after sampling, to avoid respiration and degradation losses of NSCs. In
182 the lab, samples were dried for 72 h at 64 °C and ball milled to a fine powder (Retsch® MM200,

183 Haan, Germany). We assessed the concentration of soluble sugars (glucose, fructose and
184 sucrose) and starch, which account for at least 80 % of the mobile carbon pool in both species
185 (Hoch et al., 2003). High performance liquid chromatography with pulsed amperometric
186 detection on a Dionex[®] ICS 3000 ion chromatography system equipped with an autosampler
187 was used to determine sugar/starch concentration (Raessler et al., 2010). Sample preparation
188 for sugars followed the protocol in Hartmann et al. (2013) and for starch in Landh usser et al.
189 (2018).

190 *2.4 Sampling of reproductive structures*

191 In autumn 2015, a masting year for both species in southern Germany, reproductive structures
192 (RS) of beech (complete beechnuts, fruits plus huskers) and spruce (cones) were sampled. For
193 beech, a random column of 1 m² ground area per plot (often combining the canopy of a few
194 beech individuals) was selected and the number of beechnuts of this area were counted (N_{beechnut}
195 in m⁻²). Afterwards, these beechnuts were harvested and oven-dried at 64 °C and the dry mass
196 of each beechnut was assessed (DM_{beechnut} , in g). In combination with the sum of the beech tree
197 crown spread area per plot (average for CO plots: 149.1 ± 25.2 m² and TE plots: 147.9 ± 69.5
198 m²), we then estimated the dry mass of beechnuts per plot ($DM_{\text{beechnut_plot}}$ in kg). For spruce,
199 five random branches from the main stem of 8 CO and 8 TE trees were selected and the number
200 of cones were counted, afterwards the number of branches of the main stem was counted and
201 the total number of cones was calculated (N_{cones} , unitless). Subsequently, the cones of the five
202 branches were harvested, oven-dried at 64 °C and each cone was weighed (DM_{cones} , in g). To
203 get the total dry mass of cones per tree ($DM_{\text{cones_per_tree}}$, in kg), we multiplied N_{cones} with DM_{cones}
204 for each tree individually.

205 *2.5 Modeling biomass increment with process-based model BALANCE*

206 Biomass and growth of organs of individual trees were calculated using the eco-physiological
207 growth model BALANCE (Grote and Pretzsch, 2002), which simulates the three-dimensional
208 development of individual trees based on external environmental factors such as weather (e.g.
209 temperature, precipitation, etc.), CO₂ concentration, soil conditions and water availability.
210 Further on, the model describes the consequences of the individual competition for light, water
211 and nutrients. As individual environmental conditions change with the tree development, the
212 influence of stand structure, species mixture and management options are taken into account,
213 too. Consequently, biomass increment is simulated in dependence of the carbon and nitrogen
214 uptake for each segment on the basis of its energy supply and resource (e.g. water) availability
215 (Fontes et al., 2011). Model outcomes are validated against empirical measurements of the

216 carbon and water balance (including drought scenarios) several times (Rötzer et al., 2017b,
 217 2013, 2012, 2010) and also for the above vs. below ground biomass ratio (Rötzer et al., 2009).
 218 For our study, BALANCE was used to generate the biomass of the fine roots (BM_{FR}), the coarse
 219 roots (BM_{CR}), the stem (BM_S), the branches/twigs (BM_T) and of the leaves (BM_L , for spruce
 220 only, Table S1). The simulation was based on meteorological data from 2013 to 2015 from the
 221 nearby weather stations in Freising with initial tree data and soil characteristics as described in
 222 Rötzer et al. (2017b, 2017a). Precipitation of the drought plots was reduced according to the
 223 throughfall exclusion experiment, i.e., rainfall was set to zero for the period when the roofs
 224 were closed (Grams et al. 2021). To validate the model outcome, we compared the modeled
 225 biomass increment of stems in 2014 and 2015 (BI_{stem} , in kg) with measured data of band
 226 dendrometers (Dendrometer D1, METER, Munich, Germany) at 1.3 m above the ground
 227 (diameter increment D_{stem} , in cm) for both species and treatments (i.e. CO and TE).

228 2.6 NSC pool calculations

229 We scaled the concentration of SC and starch in each tissue with its modeled biomass (BM in
 230 kg, Table S3) to calculate the overall C pools ($SC_{pool_CO/TE}$ and $Starch_{pool_CO/TE}$, both in kg,
 231 except for spruce $Starch_{pool}$ in g). For stem and branches/twigs, we additionally estimated the
 232 proportion of xylem and phloem by measuring the stem circle area at DBH (assessed via
 233 measuring tape) and the annulus area for phloem (phloem thickness was assessed with a caliper
 234 on the samples taken with the cork borer). For beech, the proportion of the stem xylem was CO:
 235 0.989 ± 0.002 and TE: 0.990 ± 0.003 , and for spruce it was CO: 0.988 ± 0.003 and TE: 0.989
 236 ± 0.002 , respectively. We then applied the following formula to calculate the stem and
 237 branch/twig pool sizes:

$$238 \quad SC/Starch_{pool} = concentration_{phloem} * biomass_{phloem} * proportion_{phloem}$$

$$239 \quad + concentration_{xylem} * biomass_{xylem} * proportion_{xylem}$$

240 Overall, the phloem was thicker in spruce compared to beech ($P < 0.001$). In spruce there was
 241 a significant difference between the phloem thickness, with CO trees having a thicker phloem
 242 (0.46 ± 0.05 cm) than TE trees (0.39 ± 0.06 cm) ($P < 0.05$), while in beech we did not find any
 243 statistical difference in phloem thickness ($P = 0.17$). However, in beech there appeared to be a
 244 similar trend with CO trees (0.35 ± 0.05 cm) having somewhat thicker phloem compared to TE
 245 trees (0.31 ± 0.05 cm). For beech, phloem thickness correlated strongly positive with the stem
 246 diameter regardless of treatment ($P < 0.05$, $R^2 = 0.31$), while in spruce, this correlation was only
 247 valid for TE trees ($P < 0.01$, $R^2 = 0.76$) and not for CO trees ($P = 0.91$, $R^2 = 0.00$).

248 *2.7 Statistical analyses*

249 Data were analyzed for statistical differences using R (version 3.6, Development Core Team
250 2008) in RStudio (version 1.2.1335, RStudio Team 2015). Every model was tested for
251 normality of the residuals (Shapiro test) and data was tested for homogeneity of variances
252 (Levene test) beforehand. For differences in Ψ_{PD} , biomass increment, NSC concentrations and
253 pools, a linear mixed effect model ('lme' function) was calculated using the species (beech and
254 spruce), year (2013, 2014 and 2015) and the treatment (CO and TE) as fixed and the tree
255 individual nested in the plot as a random effect (package: nlme, version: 3.1-137). If the mixed
256 effect model showed significant effects, we performed a post-hoc test with the 'emmeans'
257 function with Tukey correction (package: emmeans, version: 1.3.1). For data of cones and
258 beechnuts, the same functions were used, but with the plot as the random effect. To validate the
259 BALANCE-data we used a linear model (package: stats, version: 3.5.0) with the BI_{stem} as the
260 dependent variable and the D_{stem} as the test variable. Data were plotted with the 'boxplot'
261 function (package: graphics, version: 3.5.2, showing the first, second (i.e., median, solid line
262 within the box) and third quartile as a box, the whiskers indicate the 1.5 interquartile range and
263 empty dots showing outliers), the 'plot' function (package: graphics, version: 3.5.2) and the
264 'ggplot' function (package: ggplot2, version: 3.1.0). Data in text and tables are given as mean
265 ± 1 SD.

266 **3. Results**

267 *3.1 Drought treatment effects on soil and tree water status*

268 In 2013 prior to establishing the drought, leaf Ψ_{PD} was not different between trees in the
269 treatment plots for both species (beech: -0.41 ± 0.09 MPa and spruce: -0.51 ± 0.06 MPa Tab.
270 1). Upon throughfall exclusion, TE trees of both species had significantly lower Ψ_{PD} than CO
271 trees. In beech in 2014 and 2015, Ψ_{PD} was reduced to -0.62 ± 0.08 and -1.14 ± 0.31 MPa (Tab.
272 1) corresponding to reductions by 51 and 178 % compared to 2013, respectively. Ψ_{PD} in Spruce
273 was reduction stronger to -1.33 ± 0.12 and -1.68 ± 0.13 MPa in 2014 and 2015 (Tab. 1),
274 respectively (reductions by 160 and 229 % relative to 2013, respectively). However, during the
275 extraordinary dry year of 2015 the CO trees of both species showed signs of drought stress in
276 their water potential (beech: -0.76 ± 0.24 MPa and spruce: -1.38 ± 0.22 MPa), with values
277 similar to those found in TE trees in 2014, for both species respectively (Tab. 1).

278 Before the onset of the drought treatment in 2013, soil water content (SWC) to a soil depth of
279 70 cm was very similar between CO and TE plots (mean difference between CO and TE
280 averaged over all depths: 0.1 ± 1.8 vol.-%, p-value > 0.05 , Tab. 2). During the growing seasons

281 of 2014 and 2015, SWC was generally lower on TE compared to CO plots. In 2014, the mean
282 difference between CO and TE was 6.0 ± 4.0 vol.-% (p-value = 0.023) and in 2015 9.5 ± 4.5
283 vol.-% (p-value = 0.002, Tab. 2). During winter, when roofs were permanently open, SWC on
284 all plots was partially recharged by precipitation. In August and September 2015, the SWC on
285 CO plots decreased due to very warm and dry conditions and reached values similar to TE plots.

286 *3.2 Modelled biomass increments using BALANCE*

287 We found strong overall positive linear correlations (Pearson) between the measured stem
288 increment at 1.3 m above ground level and modeled stem biomass increment (Fig. 1), with
289 strong positive correlations for each group (Pearson correlation: P-value/Pearson coefficient of
290 CO spruce 2014: 0.090/0.64 & 2015: 0.011/0.83, TE spruce 2014: <0.001/0.97 & 2015:
291 0.027/0.76, CO beech 2014: 0.028/0.76 & 2015: 0.003/0.89 and TE beech 2014: <0.001/0.97
292 & 2015: 0.003/0.92, Fig. 1). Slopes and intercepts were not significantly different between
293 treatments for beech and spruce within the single years, suggesting that our model appears to
294 generate realistic results, with the TE trees showing significantly decreased growth for almost
295 all tissues compared to CO trees. Overall, spruce showed a higher growth than beech, especially
296 of stems and coarse roots (Tab. 3). Modelled biomass of spruce fine root increment (BI_{FR}) was
297 34% lower for both drought years in TE compared to CO trees but not significantly different
298 (p-value 2014: < 0.1 and 2015: > 0.1), and significantly lower by 46 % in TE beech compared
299 to CO trees (p-value 2014: < 0.05 and 2015: <0.1 0.1). Coarse root growth increment (BI_{CR})
300 was significantly different between the two treatments for both species only in 2015 (p-value
301 spruce: <0.05 and beech: <0.05), with 2-times higher growth increment in CO than TE trees
302 (Tab. 3). Stem diameter growth of both species was significantly different between CO and TE
303 in both years, with a growth reduction in TE by more than 50 % (p-value spruce 2014: <0.001
304 & 2015: <0.05 and beech 2014: <0.01 & 2015: <0.01, Tab. 3). For branch/twig tissue, growth
305 reduction was also about 30 to 40 % in TE trees compared to CO trees and significant different
306 between treatments for both species and years (p-value spruce 2014: <0.05 & 2015: <0.05 and
307 beech 2014: <0.01 & 2015: <0.05, Tab. 3). For leaves of spruce, there were no significant
308 differences in both years, nevertheless reductions of 30 and 39 % under TE compared to CO
309 were estimated in 2014 and 2015, respectively (Tab. 3).

310 *3.3 Biomass investment in reproductive structures during a masting year (2015)*

311 In the masting year, TE trees of spruce produced about three times more cones per tree than CO
312 trees (N_{cones} , CO: 235 ± 66 and TE: 775 ± 349 , Fig. 2), while the dry mass per cone (DM_{cone})
313 was on average about three times smaller in TE than in CO trees (TE: 22.0 ± 6.5 and CO: 59.0

314 ± 17.2 g, respectively, Fig. 2). Thus, the overall invested biomass in cones per tree
315 ($DM_{\text{cones_per_tree}}$) was not different between CO (13.4 ± 4.8 kg) and TE (15.7 ± 5.6 kg, Fig. 2).
316 In beech, neither the amount of beechnuts per ground area (N_{beechnut} , CO: 21.0 ± 1.6 m⁻² and
317 TE: 19.3 ± 1.8 m⁻²) nor the mass per beechnut (DM_{beechnut} , CO: 0.73 ± 0.21 g and TE: $0.82 \pm$
318 0.28 g) differed between the treatments. Therefore, also the total dry mass of fruits per area
319 ($DM_{\text{beechnut_plot}}$, CO: 2.27 ± 0.66 kg and TE: 2.31 ± 1.32 kg) was very similar in CO and TE
320 (Fig. 2). Accordingly, the invested biomass of both tree species into reproductive structures was
321 not affected by drought although the number of spruce cones was three time higher in TE
322 compared to controls.

323 *3.4 Biomass investment (BI) in structural sinks vs. reproductive structures in 2015*

324 For the masting year 2015, we compared the investment of biomass (BI) into the sink tissues
325 (sum of fine/coarse root, stem and branch/twig tissues, BI_{sum} in kg) and reproductive structures
326 (cone-mass and whole beechnut-mass ($DM_{\text{beechnut_plot}}$ divided by the number of trees on each
327 plot)) per tree. For beech, the BI_{sum} of TE trees (8.4 ± 3.4 kg) was about half of CO trees (17.5
328 ± 4.8 kg, P-value < 0.01), while the beechnut-drymass did not change (CO: 0.5 ± 0.3 kg and
329 TE: 0.5 ± 0.3 kg, Fig 3a). The relative biomass investment was calculated by dividing the weight
330 of single tissues by the overall sum of all tissues (including reproductive structures). Therefore,
331 in beech the ratio of reproductive structures compared to structural tissues (e.g.
332 reproduction/(roots+stem+branches)) was 3 ± 1 % and 7 ± 3 % (P-value < 0.01), for CO and
333 TE, respectively (Fig. 3b). For spruce, the BI_{sum} was reduced by 57 % (Fig. 3a) in TE ($13.6 \pm$
334 3.1 kg) compared to CO (31.5 ± 5.6 kg, P-value < 0.0001), but the cone-drymass was very
335 similar between CO (13.4 ± 4.8 kg) and TE (15.7 ± 5.7 kg, Fig. 3a). Thus, the relative ratio of
336 the investment in reproductive structures compared to other tree tissues combined (e.g.
337 reproduction/(roots+stem+branches)) in spruce was 44 ± 7 % and 115 ± 42 % (P-value < 0.001),
338 for CO and TE, respectively. Most of the difference in structural tissues between CO and TE
339 was caused by a reduced stem and coarse root growth (Figure 3a). However, the relative
340 biomass investment of structural tissues was only significantly reduced in the stem (p-value =
341 0.0020) and coarse roots (p-value = 0.0475) of spruce and tendentially in stem of beech (p-
342 value = 0.0774, Figure 3b).

343 *3.5 NSC concentrations and NSC pools*

344 For both species, there was no significant influence of drought on the concentration of SCs
345 during the dormant season (P = 0.45 for beech and 0.23 for spruce) detected and thus all data
346 points are close to the 1:1 line in Fig. 4a and 4c. However, concentrations were different among

347 plant tissues. Structural tissues with low metabolic activity, i.e. stem xylem (SX) and coarse
348 roots (CR), showed the lowest concentrations (mean of SX and CR: $22.0 \pm 6.1 \text{ mg g}^{-1}$ for beech
349 and $7.4 \pm 5.0 \text{ mg g}^{-1}$ for spruce). Conversely, recently built, highly active and/or transport
350 tissues, i.e. fine roots (FR), stem phloem (SP), branch/twig phloem/xylem (TP/TX) and spruce
351 leaves (L), showed higher concentrations (mean of FR, SP, TP, TX and L: $40.2 \pm 8.9 \text{ mg g}^{-1}$ for
352 beech and $50.0 \pm 11.0 \text{ mg g}^{-1}$ for spruce, Fig. 4a, c). Thus, differences in NSC concentrations
353 between structural and active tissues (excluding spruce needles) were more pronounced in
354 beech than in spruce.

355 In comparison to the SC, starch concentrations of the tissues were on average 10 and 6 times
356 lower for spruce and beech, respectively. Similar to SC, starch concentrations did not differ
357 between CO and TE for both species ($P = 0.408$, Fig. 4b, d). Starch concentrations of beech
358 were highest in stem and branch/twig phloem (all years and both treatments averaged: $8.58 \pm$
359 3.48 and $11.49 \pm 2.69 \text{ mg g}^{-1}$, respectively), lowest in the stem xylem and coarse roots (all years
360 and both treatments averaged: 1.50 ± 0.68 and $1.62 \pm 0.90 \text{ mg g}^{-1}$, respectively) and in between
361 for the branch/twig xylem and fine roots (all years and both treatments averaged: 4.08 ± 1.39
362 mg g^{-1} and $7.30 \pm 2.56 \text{ mg g}^{-1}$, Fig. 4d). A similar pattern was found for spruce, with high values
363 in leaves (average over all years and both treatments: $7.76 \pm 1.64 \text{ mg g}^{-1}$) and somewhat lower
364 levels in stem and branch/twig phloem as well as branch/twig xylem (averaged over all years
365 and both treatments: 4.08 ± 2.55 , 4.25 ± 1.77 and $5.31 \pm 2.77 \text{ mg g}^{-1}$, respectively, Fig. 4b).
366 Except from the above, fine root tissue showed the highest values in spruce (averaged over all
367 years and both treatments: $12.15 \pm 4.7 \text{ mg g}^{-1}$). Much lower concentrations were found in the
368 coarse roots and stem xylem (averaged over all years and both treatments: 0.59 ± 0.30 and 0.76
369 $\pm 0.71 \text{ mg g}^{-1}$, respectively).

370 In 2013, the SC pools of tissues were similar in CO and TE trees for both species (green symbols
371 in Fig. 5a, c), but in 2014 and 2015 SC pools decreased significantly in TE, except for fine roots
372 and leaves in spruce (Fig. 5a, c, please note logarithmic scaling of axes). Overall, drought
373 significantly reduced the overall SC pools (for beech: $P < 0.001$ and spruce: $P < 0.0001$), with
374 TE trees showing smaller pool sizes in 2014 and 2015 for almost all tissues (for beech, averaged
375 for all tissues: $-49 \pm 8 \%$ and spruce, averaged for all tissues (except leaves): $-45 \pm 23 \%$).
376 Concentration of SCs were not affected by the treatment but varied between tissues, while the
377 SC pools were strongly reduced by in the TE treatment. However, the natural drought year of
378 2015 did not seem to influence the pool size of CO trees (no differences within CO between the
379 years for both species).

380 For the starch pools of beech, the influence of drought was significant ($P < 0.05$, Fig. 5d). For
381 2014 and 2015, a strong decrease in the starch pool size of TE beech compared to CO was found
382 for CR (by 53 %), stem tissue (by 45 %) and branch/twig tissue (by 58 %), except for fine roots
383 which showed an increase (by 47 %, Fig. 5d). For spruce, this signal was not as clear as for
384 beech and no significant reduction was found (Fig. 5b). Stem tissue and needle were not much
385 affected by drought in their starch pool size and the variation in branch/twig tissue and
386 coarse/fine roots over the different years was high (Fig. 5b).

387 Whole-tree pool sizes of NSCs in beech trees were 11.1 ± 4.7 and 7.0 ± 1.9 kg under CO and
388 TE, respectively (values averaged over 2014 and 2015, Fig. 6a). Under drought, NSC pools
389 sizes were significantly reduced by 40 % in 2014 (P -value < 0.05) and 42 % in 2015 (P -value
390 < 0.01 , Fig. 6a) compared to controls. In spruce averaged over 2014 and 2015, whole-tree pool
391 sizes of NSCs were smaller compared to beech with 7.4 ± 1.7 and 4.8 ± 0.7 kg under CO and
392 TE, respectively. Reductions in spruce under drought were less strong compared to beech trees
393 with 23 % in 2014 (P -value < 0.01) and 36 % in 2015 (P -value < 0.001 , Fig. 6a). Irrespective of
394 the drought treatment, the relative proportion of the single tissues in the NSC pool size were
395 remarkably constant in beech with CR = 4.1 ± 0.8 %, FR = 3.1 ± 1.8 %, SX/SP = 90.7 ± 1.4
396 and TX/TP = 2.1 ± 0.5 % (averaged over all years and treatments, Fig. 6b). Conversely, in
397 spruce, the relative proportion of the CR was strongly reduced under drought in both years (by
398 75 %, P -value < 0.0001) and strongly increased in leaves (by 55 %, P -value < 0.0001 , Fig. 6b).
399 Additionally, in 2015 the TE spruce trees showed an increase in the fine root proportion by 33
400 % (P -value < 0.1) compared to previous years and a decrease in the stem proportion by 24 %,
401 P -value < 0.05 , Fig. 6b).

402 **4. Discussion**

403 Comparisons with data from mature trees in other studies for the same NSC components
404 indicate that the trees from this study had rather low NSC levels (specially starch) (e.g. Hoch
405 et al. 2003). This difference, could be related to site conditions and stand structure, such as high
406 stand density with ongoing self-thinning (Gruber et al., 2012; Tsamir et al., 2019). However,
407 the values measured during this experiment were still in the range of other studies (Oberhuber
408 et al., 2011; Schiestl-Aalto et al., 2019) and similar to values from the same stand sampled in
409 2003 (Nikolova et al., 2020). The low values presented here could reflect the seasonal variation
410 of NSC concentrations with lowest levels in early winter (Furze et al., 2019; Martínez-Vilalta
411 et al., 2016) and the parallel hydrolyzation of starch to soluble sugars (Ramirez, 2017).

412 In response to through-fall exclusion during the growing seasons, both species showed signs of
413 moderate to harsh drought stress in 2014 and 2015 with Ψ_{PD} as low as -1.8 MPa, while in 2013,
414 before the start of the experiment, no differences were found between the treatments. This
415 should be accompanied by a reduction in stomatal conductance and assimilation of CO₂ and
416 water potential regulation at the leaf level (Hochberg et al., 2018). This reduction/regulation
417 was stronger and earlier in spruce than in beech, reflecting the more isohydric strategy of spruce
418 under drought compared to beech (Hesse et al., under preparation).

419 *4.1 NSC concentration vs. NSC pools*

420 Two subsequent years of summer drought did not significantly decrease the NSC concentrations
421 of various tissues in both species, indicating the possibility for trees to maintain them at a steady
422 level. The meta-analysis of He et al. (2020) revealed, that only under severe drought saplings
423 would show a decrease in their NSC concentrations. In addition, mature trees are supposedly
424 less strong affected than seedlings (Zhang et al., 2020). As drought did not reduce the NSC
425 concentrations in any tissue in both species, we rejected our first hypothesis that sink organs
426 further away from the source organs will be stronger affected by drought in their NSC
427 concentrations. Nevertheless, we confirm the expected differences in NSC concentrations
428 between tissues, with source and transport (e.g. phloem) structures showing higher values than
429 predominant sink tissues, such as stem growth, which was strongly reduced under drought in
430 the experimental trees (Tab. 3, Pretzsch et al., 2020). Even in C sinks with a transport distance
431 of more than 25 m from the C source, e.g. from leaves to fine roots (Fig. 4) NSC concentrations
432 remained stable, which is well in accordance with studies on juvenile trees (Hagedorn et al.,
433 2016; Hartmann et al., 2013; Muller et al., 2011).

434 NSCs are needed for many other sinks next to growth and maintenance, such as defense
435 compounds, e.g. secondary metabolites, and osmoregulation (Chaves et al., 2003; Kozłowski,
436 1992). NSC investments into osmoregulation, repair processes or defense are unavailable for
437 maintenance processes, but play an important role in acclimation processes of trees to abiotic
438 and biotic stresses. An important sink under drought is the osmoregulation of cells, mostly
439 accompanied by the incorporation of sugars or amino acids as osmotic substances (Jamnická et
440 al., 2019; Sevanto et al., 2014). As reported before (Tomasella et al., 2018), in the second
441 drought summer, i.e. 2015, both beech and spruce showed osmotic adjustments at the leaf level,
442 resulting in a decrease of the turgor loss point by 0.5 MPa, accompanied by a similar decrease
443 of the osmotic potential at full turgor. This change is based on osmoregulation (Bartlett et al.,
444 2012) another C sink the drought-stressed trees had to cope with. However, the increased C

445 demand for osmoregulation under drought and the stable level of NSC concentrations in all tree
446 organs, raises the questions whether all measured NSCs can be seen as “metabolically
447 available” carbohydrates (e.g. for transport or respiration (Prescott et al., 2020). Further studies
448 are needed to separate the “stored or unavailable” amount of NSCs from the “metabolically
449 available” NSCs in different tree organs, as this could prove to be very different between well-
450 watered and drought-stressed plants. Nevertheless, it seems reasonable that trees maintain a
451 certain concentration of NSCs even under drought, yet not much is known about the minimum
452 required to maintain functionality, especially not for mature trees (Adams et al., 2013).

453 Looking at the NSC pool sizes, all growth sinks of the tree (but especially coarse roots) showed
454 a decrease in the total amount of NSCs under drought. Despite the stable concentration of NSCs
455 in almost all organs, the overall carbon pool of tree individuals was reduced by 42 % in beech
456 and 36 % in spruce after two subsequent summer droughts. Winter NSC pools were reduced
457 under drought in both species, generally consistent with our second hypotheses. However,
458 beech and spruce were affected to a similar extent, which is against the second part of that
459 hypotheses assuming more isohydric spruce to be stronger affected by drought than more
460 anisohydric beech.

461 However, starvation thresholds, i.e. lethal NSC concentrations or pools, are widely unknown
462 for mature trees or long lasting drought periods and many experiments showed NSC remnants
463 in dead tissues and for juvenile trees (Weber et al., 2018; Wiley et al., 2019, 2017). Although
464 NSC pools are significantly reduced in drought-stressed trees of both species after two
465 subsequent summer droughts, they can be estimated sufficient for supporting respiratory
466 processes of trees for at least two years. Assuming that whole-tree respiration in $\mu\text{mol CO}_2 \text{ s}^{-1}$
467 is equal to 0.8 times the whole-plant mass in kg and that 1 g of NSC is needed for 6 g of respired
468 CO_2 (Mori et al., 2010), drought stressed spruce and beech could survive for almost 2 and 4
469 years, respectively, if the whole NSC pools could be accessed and used for maintenance
470 respiration to keep trees alive. While this is certainly not possible, as NSCs are needed for many
471 other mechanisms (see above/below), it demonstrates that, neither the more anisohydric beech
472 nor the more isohydric spruce seems therefore to be limited in their overall C availability under
473 drought (Hoch, 2015; Körner, 2003; Leuschner, 2020). Trees keep a rather constant
474 concentration of NSCs as a buffer by reducing selected C sinks under stress. This is also in
475 accordance with Garcia-Forner et al. (2017), who reported that more isohydric plants are not
476 necessarily more carbon limited than anisohydric ones.

477 The rather small reduction of NSC pool size in spruce fine roots (FR) might be explained by a
478 somewhat overestimated spruce FR biomass in the model compared to the observed loss of vital
479 fine roots under drought (Nickel et al., 2018; Zwetsloot and Bauerle, 2021). However,
480 accumulation of NSCs in roots of drought-stressed trees has been reported previously for
481 saplings (Hagedorn et al. 2016 and citations within) and is supported here for mature beech by
482 the increased starch concentrations (Fig. 4d). The reduction of NSC pools in sink tissues, in
483 particular of stem xylem (SX) and coarse roots (CR) over time, suggests that phloem transport
484 is hindered under drought as previously reported in the same experiment (Hesse et al., 2019;
485 Hikino et al., submitted). The impaired C transport to roots is also supported by the fact that
486 NSC pools in spruce leaves and by that the replenishment of carbon is not affected by drought.
487 Thus, capacity to transport carbon plays an important role in the carbon starvation scenario
488 (Hartmann et al., 2013; Ivanov et al., 2019), as stems and coarse roots might empty their
489 storages to maintain fine roots as the most important tissues for water uptake and exploitation
490 of new water sources (Jackson et al., 2000). As the annual carbon allocation of the control trees
491 was not affected by the natural drought in 2015, but as the phloem transport was reduced for
492 the same experimental trees (measured in 2015 by Hesse et al., 2019 on the same experimental
493 trees), short term droughts seem to be buffered by tree C stores when carbon transport decreases
494 for a short time. However, overall both species still seem to be far from carbon starvation or
495 even carbon limitation, even in the masting year 2015, as the remaining NSC pools can
496 potentially sustain respiration for more than one whole year.

497 *4.2 Carbon investment in reproduction under drought*

498 While it has been shown that growth is one of the first C sinks to decrease under drought (Chuste
499 et al., 2020; Muller et al., 2011), this is not clear for reproduction and might also differ among
500 species. As the total C investment into reproductive structures in beech and spruce was
501 unaffected even under recurrent summer drought with significant growth reductions,
502 reproduction must be of high priority in the hierarchy of carbon sinks in mature trees (Fatichi
503 et al., 2013; Lacoïnte, 2000). Accordingly, beech is known for showing signs of resource
504 switching towards reproduction in dry masting years (Hacket-Pain et al. 2017 and citations
505 within). Other C investments under stress, e.g. investments in defense (Huang et al., 2019a,
506 2019b) and osmoregulation (Morgan, 1984) are – at least partially – quick responses to stress
507 (Hartmann and Trumbore, 2016), and as they are partially reversible, they might not be too
508 costly for the plant after stress release. Conversely, C investments in growth or reproduction is
509 irreversible, but seems to be regulated very different in different mature tree species under
510 drought. C demand for reproduction might be partially met by the green inflorescence tissue

511 itself (Hoch et al., 2013; Landhäusser, 2011; Mund et al., 2020), or largely supported by leaves
512 of the same branchlet as reported for e.g. beech (Hoch, 2005). Knowledge about the carbon
513 autonomy of spruce cones is scarce, but cone biomass increment depends largely on C
514 assimilated in the canopy (Koppel et al., 1987). Under drought, spruce trees invested similar
515 amounts of C in reproductive and structural tissues. This might reflect a strategy to distribute
516 seeds by zoochory (Dobrovolný and Tesař, 2010) or anemochory (Dobbs, 1976) to locations
517 with potentially more favorable microclimatic conditions or available water. In contrast to our
518 third hypothesis, under drought C invested in reproductive structures was not reduced in either
519 species. Apparently, reproduction has a high priority in mature beech and spruce even under
520 severe summer drought.

521 **5. Conclusions**

522 Under two-years of experimentally induced, severe summer droughts with distinct growth
523 decline, mature beech and spruce trees maintained a stable level of NSC concentrations across
524 all studied tissues. This preserves trees' ability to react to their environment, e.g. defense against
525 biotic attacks. Shorter drought events, as the natural summer drought of 2015, hardly affected
526 the C allocation of beech and spruce, as the control trees were not affected in their NSC
527 household compared to previous years. While C pools in all tissues, except leaves and fine roots
528 of spruce, were reduced at least by about half after two consecutive drought years, both species,
529 even the more isohydric spruce, were apparently not carbon limited. Trees were still able to
530 maintain functionality by preserving rather constant NSC concentrations through reducing C
531 demand for growth. With NSC pool size in leaves in spruce being rather unaffected by drought,
532 C transport might be critical for carbon availability in non-green tissues in mature trees, as NSC
533 pool sizes in branch/twig, stem and coarse roots were affected the most. Conversely, C
534 investment in reproduction remained unaffected, partially supported by carbon autonomy of
535 inflorescence structures and the proximity to C source tissue but clearly representing a high
536 priority among C sinks of trees. Finally, a decrease in NSC concentration would only occur
537 once the NSC pools run completely empty. Therefore, we suggest assessing not only NSC
538 concentrations, but also the overall amount, i.e. pool sizes, of NSCs to assess the carbon status
539 of mature trees under drought.

540 **Declaration of Competing Interest**

541 None.

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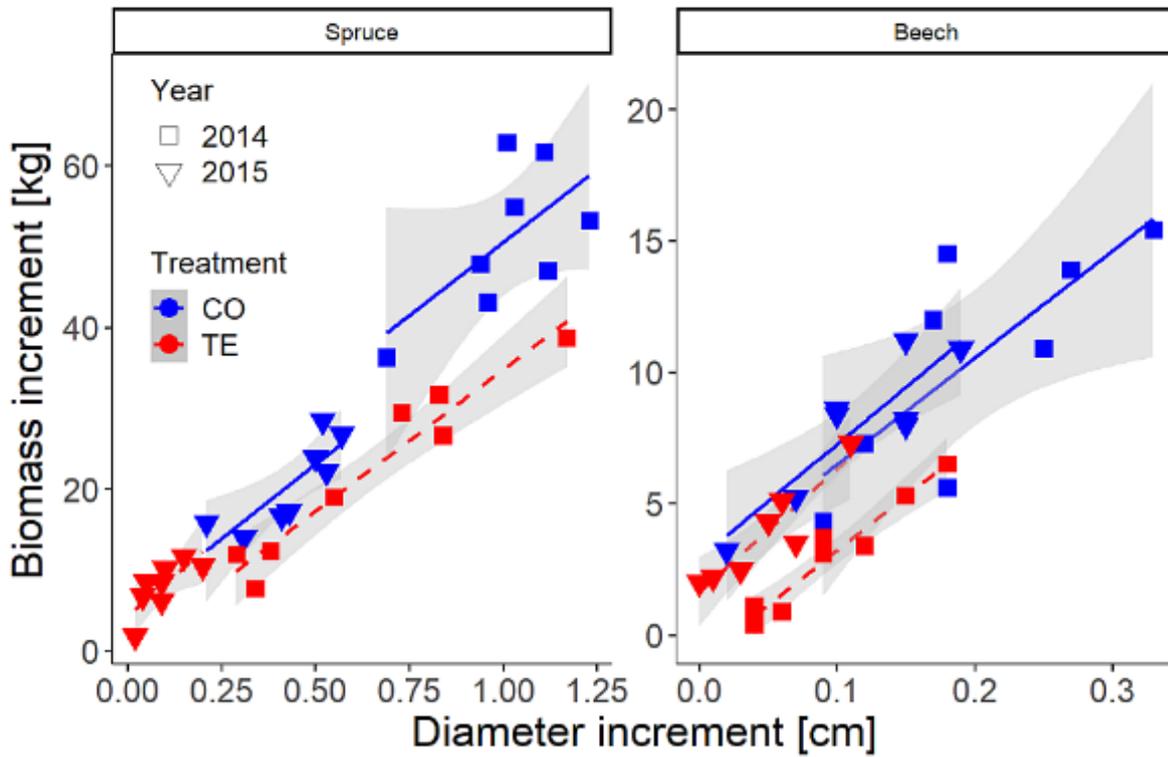
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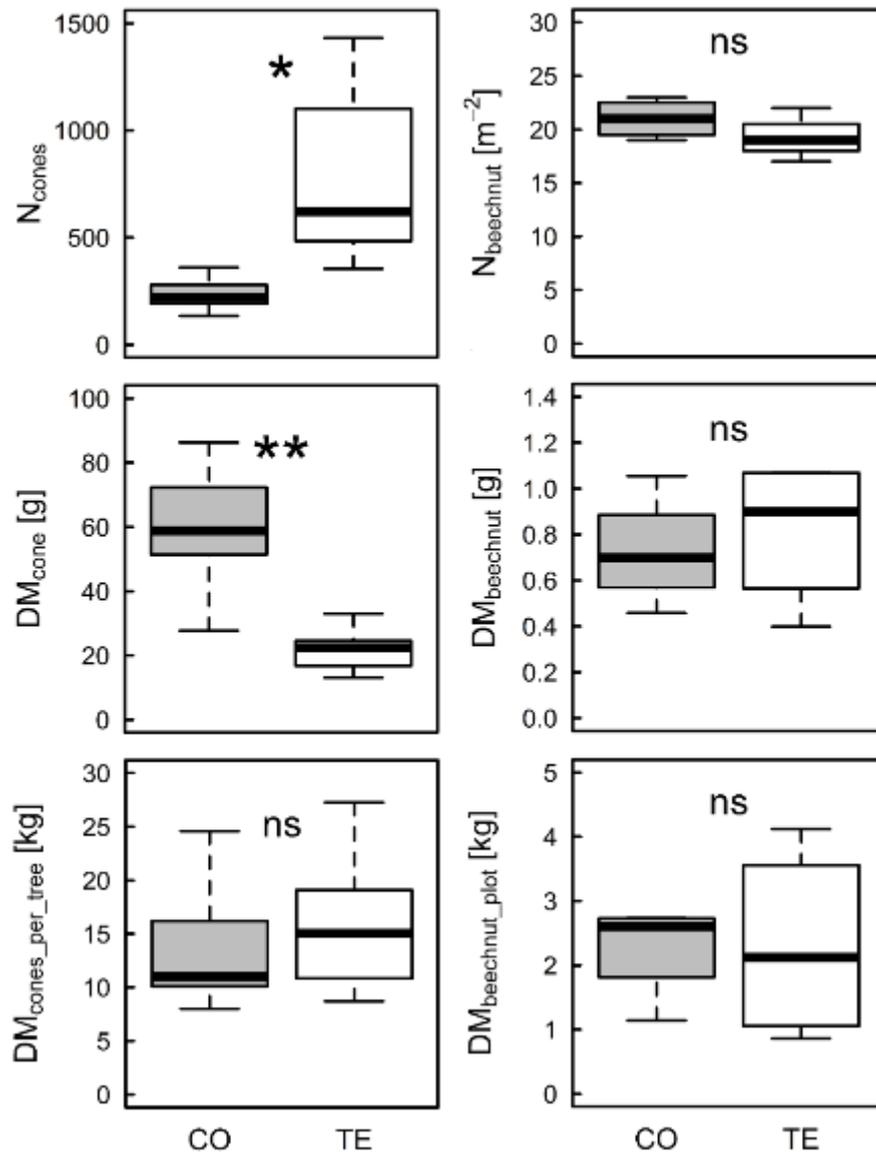
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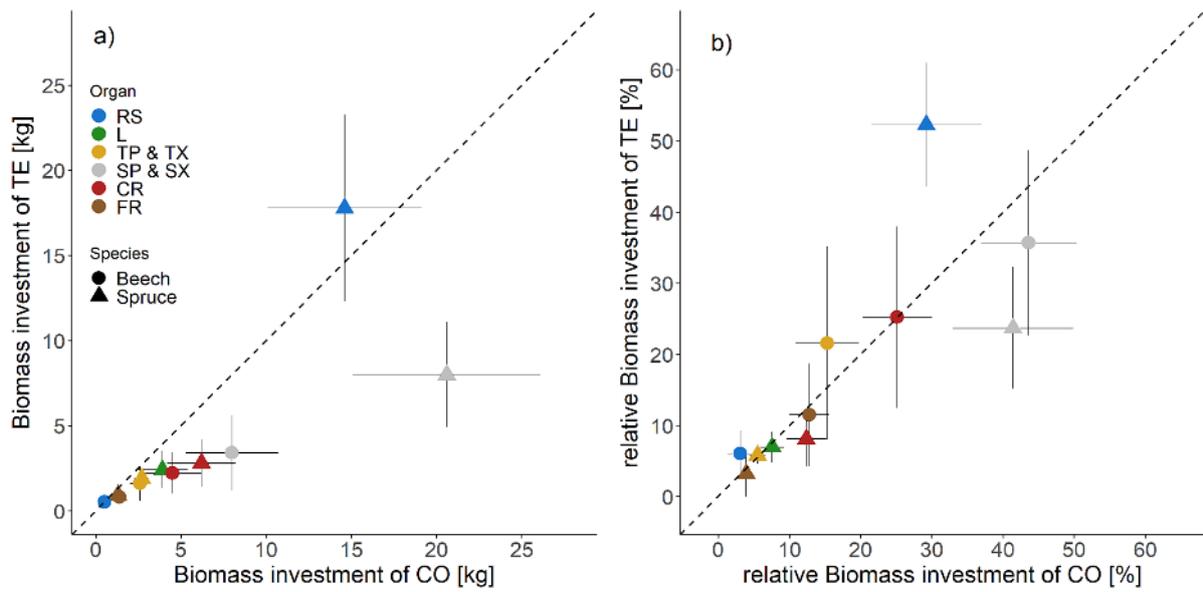
876 *Figure 1: Measured dendrometer increment at 1.3 m height vs modeled stem biomass increment*
 877 *versus; with triangles showing the values for spruce and circles for beech and in red for*
 878 *throughfall exclusion (TE) and blue for control (CO). Red and blue lines show Pearson*
 879 *correlations for each group of beech and spruce for CO (solid lines) and TE (dashed lines).*
 880 *Gray areas show the 95% confidence interval for each group. Please note the different scaling*
 881 *of the beech and spruce graph.*

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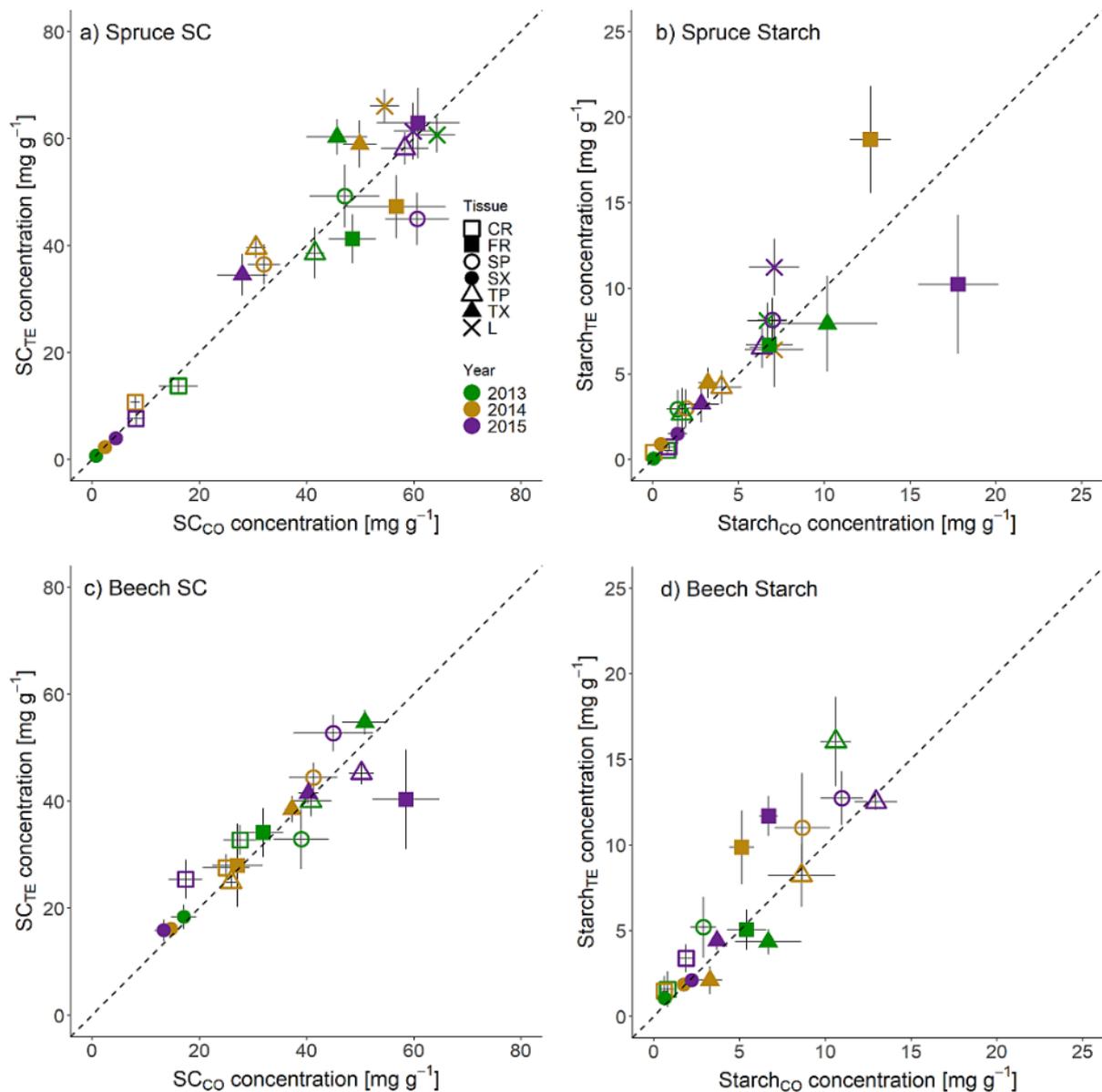
884 *Figure 2: Number (top), dry mass of single (middle) and accumulated mass (bottom) of*
 885 *reproductive structures in spruce per tree (left) and beech per plot (right; p-values: ns > 0.05,*
 886 ** < 0.05 and ** < 0.01).*



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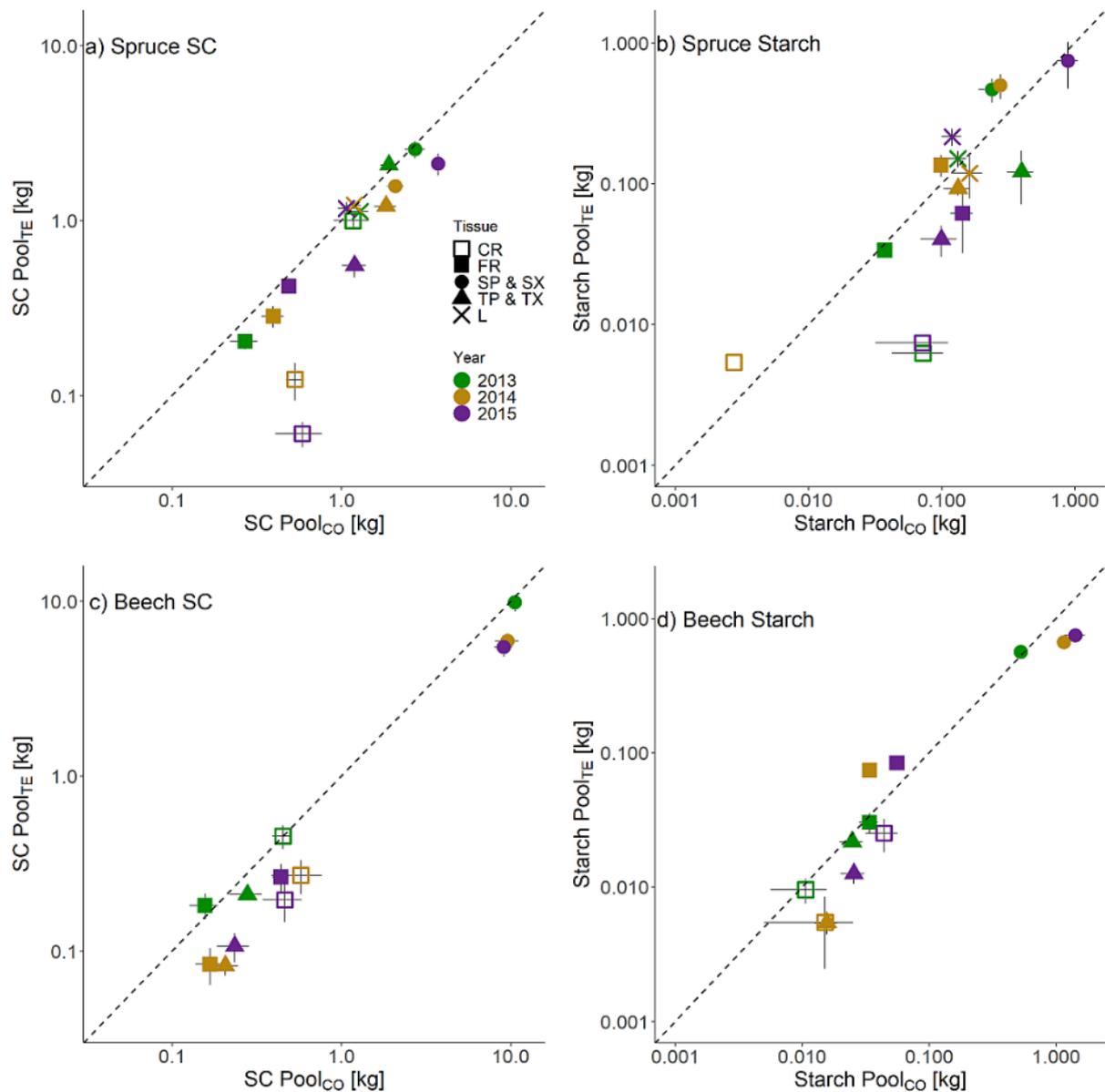
888 *Figure 3: Total (a, data from table 3) and relative modeled biomass investment (b) in the*
 889 *masting year 2015. (RS = beechnut or cones (blue), leaves = L (green), branches/twigs = TP*
 890 *& TX (beige), stem = SP & SX (gray), coarse roots = CR (red) and fine roots = FR (brown).*
 891 *Data are given as the mean \pm 1SD.*

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894 *Figure 4: Mean soluble carbohydrates (SC, left) and starch (right) concentrations of spruce*
 895 *(top) and beech (bottom) in different tissues. X-axis is giving the values of control (CO) trees*
 896 *and the Y-axis of throughfall-exclusion (TE = drought stressed) trees. Coarse root = CR (open*
 897 *square), fine root = FR (closed square), stem phloem = SP (open circle), stem xylem = SX*
 898 *(closed circle), branch/twig phloem = TP (open triangle), branch/twig xylem = TX (closed*
 899 *triangle) and leaves = L (cross). Years are given by colors: 2013 (before drought in green),*
 900 *2014 in beige and 2015 in purple. Data are given as mean \pm 1 SE, dashed line = 1 to 1 line,*
 901 *with drought treatment: $P > 0.05$, tissue: $P < 0.0001$ and year: $P < 0.0001$ for both species).*



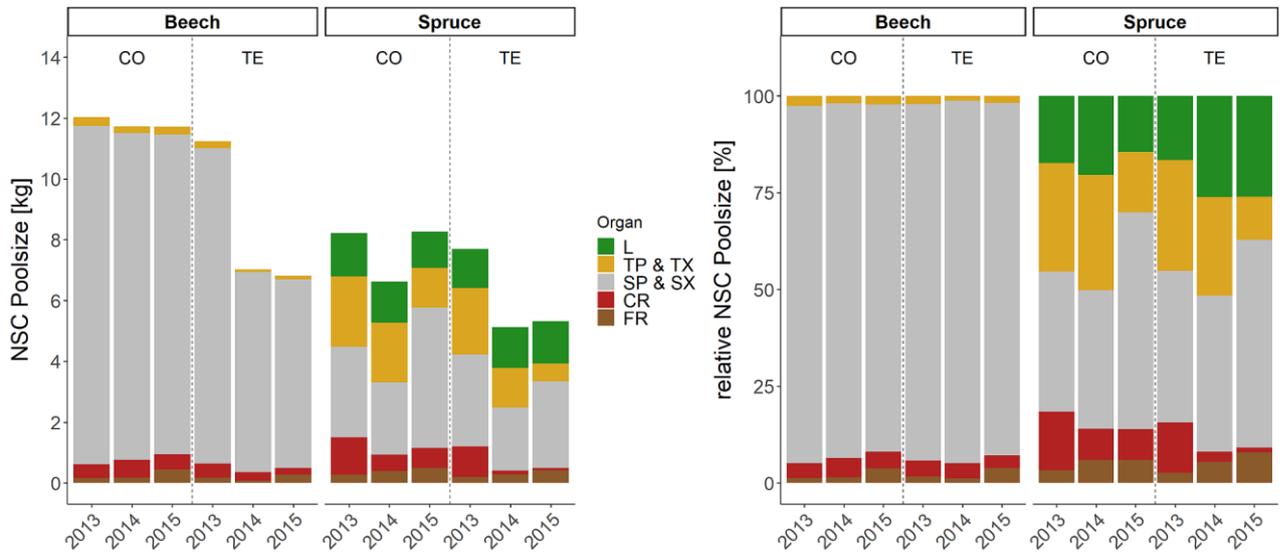
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903 *Figure 5: Mean soluble carbohydrates pool (SC, left) and starch pool (right) sizes of spruce*
 904 *(top) and beech (bottom) in different tissues. Coarse root = CR (open square), fine root = FR*
 905 *(closed square), stem phloem and xylem = SP & SX (closed circle), branch/twig phloem and*
 906 *xylem = TP & TX (closed triangle) and leaves = L (cross) and years (2013 (green) = before*
 907 *drought, 2014 (beige) = 1st year drought and 2015 (purple) = 2nd year drought) with the x-axis*
 908 *(in log-scale) showing the values of control (CO) trees and the y-axis (in log scale) of*
 909 *throughfall-exclusion (TE = drought stressed) trees. (Data are given as mean ± 1 SE, dashed*
 910 *line = 1 to 1 line, with treatment: P < 0.01, tissue: P < 0.0001 and year: P < 0.0001 for both*
 911 *species).*

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916 *Figure 6: Total (a) and relative (b) NSC pool size of beech and spruce for different tissues*
917 *(leaves = L (green), sum of branch/twig phloem and xylem = TP & TX (beige), sum of stem*
918 *phloem and xylem = SP & SX (gray), coarse roots = CR (red) and fine roots = FR (brown)*
919 *over three years, i.e. 2013 (before onset of drought treatment) and in 2014 and 2015 in control*
920 *(CO) and drought stressed (TE) trees.*

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932 **Tables**

933 *Table 1: Pre-dawn water potential of beech and spruce at the end of the vegetation period 2013,*
 934 *2014 and 2015 (Data are given as mean \pm 1 SD, different letters indicate significant differences*
 935 *between groups (treatment*year)).*

Ψ_{PD} [MPa]	Spruce	
	CO	TE
2013	-0.52 \pm 0.06 a	-0.49 \pm 0.05 a
2014	-0.78 \pm 0.31 a	-1.33 \pm 0.12 b
2015	-1.38 \pm 0.22 b	-1.68 \pm 0.13 c
Ψ_{PD} [MPa]	Beech	
	CO	TE
2013	-0.41 \pm 0.10 ab	-0.40 \pm 0.08 ab
2014	-0.30 \pm 0.04 a	-0.62 \pm 0.08 bc
2015	-0.76 \pm 0.24 c	-1.14 \pm 0.31 d

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938 *Table 2: Mean monthly volumetric soil water content (SWC) of control (CO) and throughfall-exclusion (TE) plots to a soil depth of 70 cm (from*
 939 *January 2013 to December 2015). Timespan of roof closure: from 6th of March to 12th of September in 2014 and 10th of March to 21st of November*
 940 *in 2015. Data are given as the monthly mean ± 1 SE per treatment and year. Statistically significant differences between CO and TE are given between*
 941 *each month for each year separately with p-value $> 0.05 = ns$, $< 0.05 = *$, $< 0.01 = **$ and $< 0.001 = ***$). For further details see Grams et al. 2021.*

SWC [vol.-%]		Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
2013	CO	34.1 \pm 1.2	38.4 \pm 1.0	31.9 \pm 0.7	34.2 \pm 0.9	30.8 \pm 1.1	32.5 \pm 0.6	24.9 \pm 0.6	18.6 \pm 0.6	20.4 \pm 0.7	23.5 \pm 0.6	24.5 \pm 1.2	28.7 \pm 1.1
		ns											
	TE	32.7 \pm 1.3	34.3 \pm 1.4	32.7 \pm 0.8	34.5 \pm 1.0	31.9 \pm 1.1	32.7 \pm 0.7	25.3 \pm 0.6	20.0 \pm 0.7	21.0 \pm 0.7	22.9 \pm 0.6	23.1 \pm 1.3	26.0 \pm 1.2
2014	CO	30.6 \pm 0.8	33.2 \pm 1.0	27.8 \pm 0.6	29.2 \pm 0.5	29.0 \pm 0.7	24.7 \pm 0.6	22.1 \pm 0.5	24.1 \pm 0.6	24.7 \pm 0.6	26.3 \pm 0.7	30.6 \pm 0.7	32.7 \pm 0.6
		ns	ns	ns	ns	*	*	*	**	**	**	***	***
	TE	27.7 \pm 1.0	29.1 \pm 1.1	25.4 \pm 0.7	27.8 \pm 0.6	23.4 \pm 0.7	19.5 \pm 0.6	17.5 \pm 0.5	16.5 \pm 0.6	16.7 \pm 0.5	17.0 \pm 0.6	17.4 \pm 0.6	17.6 \pm 0.6
2015	CO	35.6 \pm 0.5	35.0 \pm 0.5	33.8 \pm 0.5	34.2 \pm 0.4	35.5 \pm 0.5	34.6 \pm 0.4	24.9 \pm 0.5	20.4 \pm 0.4	20.1 \pm 0.5	22.5 \pm 0.5	24.3 \pm 0.5	27.7 \pm 0.5
		***	***	**	***	***	***	**	ns	ns	*	**	**
	TE	20.1 \pm 0.6	21.3 \pm 0.6	21.3 \pm 0.5	22.6 \pm 0.5	20.6 \pm 0.5	19.0 \pm 0.4	17.5 \pm 0.5	16.8 \pm 0.4	16.7 \pm 0.4	16.7 \pm 0.4	16.7 \pm 0.4	17.8 \pm 0.4

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943 *Table 3: Modeled biomass increment (BI) by the model BALANCE of spruce (top) and beech (bottom) for 2014 and 2015. Data are given as mean ±*
 944 *1 SD, p-values: ns > 0.1, • < 0.1, * < 0.05, ** < 0.01 and *** < 0.001.*

Spruce		2014		2015		
BI [kg]	CO		TE	CO		TE
Fine root	1.4 ± 0.4	•	0.9 ± 0.2	1.3 ± 0.2	ns	0.9 ± 0.2
Coarse root	7.0 ± 3.1	ns	7.5 ± 3.4	6.2 ± 2.0	*	2.8 ± 1.4
Stem	50.9 ± 9.1	***	22.2 ± 11.1	20.6 ± 5.5	*	8.0 ± 3.1
Branch/Twig	2.6 ± 0.7	*	1.9 ± 0.7	2.7 ± 0.4	*	1.9 ± 0.4
Leaf	4.7 ± 1.9	ns	3.3 ± 1.3	3.9 ± 1.5	ns	2.4 ± 1.1

Beech		2014		2015		
BI [kg]	CO		TE	CO		TE
Fine root	1.4 ± 0.2	*	0.7 ± 0.1	1.4 ± 0.4	•	0.8 ± 0.2
Coarse root	3.3 ± 1.8	ns	2.1 ± 1.2	4.5 ± 1.7	*	2.2 ± 1.2
Stem	10.5 ± 4.3	**	3.1 ± 2.2	8.0 ± 2.7	**	3.4 ± 2.2
Branch/Twig	2.6 ± 0.2	**	1.2 ± 0.7	2.6 ± 0.6	*	1.6 ± 1.0

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948 **Supplement**949 *Table S1: Measured mean soluble carbohydrate (SC in mg/ mg drymass) concentrations of spruce and beech in different tissues for 2013 to 2015.*950 *Data are given as mean \pm 1 SE.*

Spruce		2013		2014		2015	
SC [mg/mg]	CO	TE	CO	TE	CO	TE	
Fine root	48.59 \pm 4.40	41.24 \pm 4.55	56.80 \pm 9.20	47.25 \pm 5.91	60.85 \pm 7.76	62.88 \pm 6.58	
Coarse root	16.13 \pm 3.63	13.74 \pm 1.17	8.04 \pm 0.88	10.67 \pm 1.39	8.20 \pm 1.67	7.66 \pm 1.54	
Stem xylem	0.80 \pm 0.04	0.70 \pm 0.04	2.48 \pm 0.19	2.31 \pm 0.14	4.49 \pm 0.41	3.95 \pm 0.55	
Stem phloem	47.11 \pm 6.53	49.23 \pm 5.86	32.06 \pm 3.01	36.48 \pm 3.70	60.65 \pm 5.91	44.97 \pm 4.88	
Twig xylem	45.69 \pm 5.68	60.26 \pm 3.29	49.93 \pm 3.14	58.94 \pm 4.39	28.02 \pm 4.66	34.52 \pm 3.88	
Twig phloem	41.58 \pm 1.73	38.56 \pm 4.71	30.58 \pm 1.78	39.58 \pm 1.87	58.33 \pm 4.41	58.16 \pm 3.04	
Leaf	64.38 \pm 3.38	60.65 \pm 3.28	54.60 \pm 2.73	66.04 \pm 3.19	59.91 \pm 3.59	61.37 \pm 5.26	
Beech		2013		2014		2015	
SC [mg/mg]	CO	TE	CO	TE	CO	TE	
Fine root	31.97 \pm 3.69	34.11 \pm 4.56	27.16 \pm 4.70	27.98 \pm 7.77	58.58 \pm 6.23	40.32 \pm 18.60	
Coarse root	27.61 \pm 3.05	32.73 \pm 2.80	25.01 \pm 4.37	27.53 \pm 2.51	17.46 \pm 3.15	25.39 \pm 3.62	
Stem xylem	17.15 \pm 2.36	18.33 \pm 2.25	14.78 \pm 1.25	16.08 \pm 0.96	13.38 \pm 1.75	15.83 \pm 2.03	
Stem phloem	39.02 \pm 14.53	32.89 \pm 5.60	41.29 \pm 4.49	44.41 \pm 2.72	45.03 \pm 7.39	52.67 \pm 3.40	

Twig xylem	50.90 ± 4.23	54.74 ± 2.29	37.30 ± 1.32	38.47 ± 2.45	40.35 ± 1.84	41.47 ± 1.96
Twig phloem	40.83 ± 3.82	40.02 ± 2.91	25.94 ± 1.45	24.83 ± 0.99	50.26 ± 2.34	45.19 ± 5.45

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952 *Table S2: Measured starch (in mg/mg drymass) concentrations of spruce and beech in different tissues for 2013 to 2015. Data are given as mean ±*
953 *1 SE.*

Spruce		2013		2014		2015	
Starch [mg/mg]	CO	TE	CO	TE	CO	TE	
Fine root	6.80 ± 1.35	6.70 ± 0.21	12.69 ± 1.19	18.68 ± 3.12	17.81 ± 2.34	10.22 ± 4.05	
Coarse root	0.86 ± 0.34	0.54 ± 0.21	0.05 ± 0.04	0.42 ± 0.10	0.94 ± 0.35	0.74 ± 0.22	
Stem xylem	0.06 ± 0.06	0.05 ± 0.04	0.49 ± 0.03	0.89 ± 0.12	1.45 ± 0.56	1.49 ± 0.58	
Stem phloem	1.45 ± 0.63	2.95 ± 1.10	1.95 ± 0.54	2.98 ± 1.11	6.98 ± 0.85	8.15 ± 1.29	
Twig xylem	10.18 ± 2.92	7.93 ± 2.80	3.21 ± 0.55	4.47 ± 0.88	2.84 ± 1.02	3.24 ± 1.09	
Twig phloem	1.71 ± 0.59	4.71 ± 1.49	4.01 ± 1.16	4.22 ± 0.97	6.37 ± 0.71	6.53 ± 1.18	
Leaf	6.67 ± 1.14	8.11 ± 1.05	7.08 ± 1.70	6.41 ± 2.18	7.08 ± 1.45	11.23 ± 1.65	
Beech		2013		2014		2015	
Starch [mg/mg]	CO	TE	CO	TE	CO	TE	
Fine root	5.41 ± 1.14	5.04 ± 1.18	5.13 ± 0.72	9.58 ± 2.15	6.71 ± 0.54	11.68 ± 1.17	
Coarse root	0.81 ± 0.37	1.57 ± 1.03	0.62 ± 0.26	1.46 ± 0.87	1.89 ± 0.52	3.37 ± 0.81	

Stem xylem	0.64 ± 0.10	1.03 ± 0.16	1.76 ± 0.09	1.84 ± 0.23	2.23 ± 0.36	2.09 ± 0.30
Stem phloem	2.89 ± 0.74	5.19 ± 1.78	8.67 ± 1.60	11.00 ± 3.19	10.97 ± 1.23	12.73 ± 1.57
Twig xylem	6.67 ± 1.93	4.34 ± 0.73	3.27 ± 0.71	2.10 ± 0.80	3.69 ± 0.24	4.40 ± 0.53
Twig phloem	10.61 ± 0.89	16.04 ± 2.61	8.62 ± 1.97	8.22 ± 1.84	12.95 ± 1.23	12.51 ± 0.47

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955 *Table S3: Modeled biomass (BM in kg) of spruce (top) and beech (bottom) for 2013 to 2015. Data are given as mean ± 1 SD.*

Spruce	2013		2014		2015	
BM [kg]	CO	TE	CO	TE	CO	TE
Fine root	5.5 ± 0.4	5.0 ± 0.9	6.9 ± 0.6	5.9 ± 0.9	8.2 ± 0.7	6.8 ± 0.7
Coarse root	63.1 ± 22.6	65.2 ± 18.3	70.1 ± 24.7	72.7 ± 19.3	76.3 ± 26.2	75.5 ± 20.0
Stem	567.7 ± 141.6	563.2 ± 123.5	618.6 ± 143.6	585.4 ± 116.3	639.2 ± 144.5	593.4 ± 117.8
Branch/Twig	17.8 ± 3.3	17.3 ± 4.3	20.4 ± 3.7	19.2 ± 4.7	23.1 ± 4.0	21.1 ± 4.6
Leaf	17.5 ± 5.2	16.8 ± 5.2	22.2 ± 6.8	20.1 ± 4.9	26.0 ± 8.1	22.5 ± 5.5
Beech	2013		2014		2015	
BM [kg]	CO	TE	CO	TE	CO	TE
Fine root	4.8 ± 0.6	5.4 ± 0.8	6.2 ± 0.7	6.1 ± 0.9	7.6 ± 0.9	6.9 ± 0.9
Coarse root	55.3 ± 9.5	53.5 ± 10.0	58.6 ± 11.4	55.6 ± 11.1	63.1 ± 13.0	57.8 ± 12.3
Stem	463.3 ± 191.6	453.2 ± 236.4	473.8 ± 193.7	456.3 ± 237.3	481.8 ± 195.1	459.7 ± 238.6

Branch/Twig	15.5 ± 0.9	15.7 ± 3.3	18.1 ± 0.9	16.9 ± 4.0	20.7 ± 1.0	18.6 ± 5.0
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