

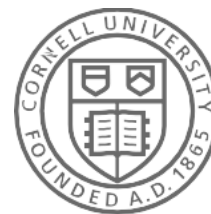
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The effects of mixed-species root zones on the resistance of soil bacteria and fungi to long-term experimental and natural reductions in soil moisture --Manuscript Draft--

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Abstract:	<p>Mixed forest stands tend to be more resistant to drought than species-specific stands partially due to complementarity in root ecology and physiology. We asked whether complementary differences in the drought resistance of soil microbiomes might contribute to this phenomenon. We experimented on the effects of reduced soil moisture on bacterial and fungal community composition in species-specific (single species) and mixed-species root zones of Norway spruce and European beech forests in a 5-year-old throughfall-exclusion experiment and across seasonal (spring-summer-fall) and latitudinal moisture gradients. Bacteria were most responsive to changes in soil moisture, especially members of Rhizobiales, while fungi were largely unaffected, including ectomycorrhizal fungi (EMF). Community resistance was higher in spruce relative to beech root zones, corresponding with the proportions of drought-favored (more in spruce) and drought-sensitive bacterial taxa (more in beech). The spruce soil microbiome also exhibited greater resistance to seasonal changes between spring (wettest) and fall (driest). Mixed-species root zones contained a hybrid of beech- and spruce-associated microbiomes. Several bacterial populations exhibited either enhanced resistance or greater susceptibility to drought in mixed root zones. Overall, patterns in the relative abundances of soil bacteria closely tracked moisture in seasonal and latitudinal precipitation gradients and were more predictive of soil water content than other environmental variables. We conclude that complementary differences in the drought resistance of soil microbiomes can occur and the likeliest form of complementarity in mixed-root zones coincides with the enrichment of drought-tolerant bacteria associated with spruce and the sustenance of EMF by beech.</p>
Response to Reviewers:	

Cornell CALS

College of Agriculture
and Life Sciences



February 11th, 2023

Dear Editors of the *Science of the Total Environment*,

We have made the requested minor revisions to our article of original research entitled: “**The effects of mixed-species root zones on the resistance of soil bacteria and fungi to long-term experimental and natural reductions in soil moisture.**” We thank the reviewers for their thoughtful feedback and hope the manuscript now meets the standards of STOTEN.

Sincerely,

A handwritten signature in blue ink that reads "Taryn Bauerle".

Dr. Taryn L. Bauerle

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Reviewers/Editor comments:

Reviewer #1: Thank you for the good job in revising the article. Another question came to my mind while reading the discussion section.

You attribute the greater community tolerance in spruce stands than in beech stands to factors such as the litter quality, pH, etc (mainly 4.1 part, but also 4.3). These are certainly the most important factors that make the difference between the communities of the two species. However, the physiological nature of these tree species and their rooting depth contribute to the fact that beech stands are much wetter in the upper soil layer during drought periods (which has been confirmed by studies with lysimeters), while spruce stands generally tolerate drought periods less well and the water almost disappears from the upper soil layers. Could not spruce communities be trained to such conditions over time and thus select more drought-resistant species? I would welcome a discussion on this point as well.

This is an important point and was a clear oversight on our part. We are grateful for your suggestion. In section 4.1, we now write: “Physiological differences in rooting depth result in consistently drier conditions in shallow soils in spruce stands relative to beech (Allen et al., 2019; Zwetsloot and Bauerle, 2021). Our findings indicate that this effect is large enough to select for higher proportion of drought stress-tolerant bacteria in spruce soils.” (L384) and, later on, we write: “These adverse conditions, along with consistently lower shallow soil moisture levels in spruce stands, may select for stress-tolerant populations, which may better endure water stress.” (L399).

The effects of mixed-species root zones on the resistance of soil bacteria and fungi to long-term experimental and natural reductions in soil moisture

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Running title: Drought-microbiome effects in mixed rooting zones

Abstract

1 Mixed forest stands tend to be more resistant to drought than species-specific stands partially due
2 to complementarity in root ecology and physiology. We asked whether complementary differences
3 in the drought resistance of soil microbiomes might contribute to this phenomenon. We
4 experimented on the effects of reduced soil moisture on bacterial and fungal community
5 composition in species-specific (single species) and mixed-species root zones of Norway spruce
6 and European beech forests in a 5-year-old throughfall-exclusion experiment and across seasonal
7 (spring-summer-fall) and latitudinal moisture gradients. Bacteria were most responsive to changes
8 in soil moisture, especially members of Rhizobiales, while fungi were largely unaffected, including
9 ectomycorrhizal fungi (EMF). Community resistance was higher in spruce relative to beech root
10 zones, corresponding with the proportions of drought-favored (more in spruce) and drought-
11 sensitive bacterial taxa (more in beech). The spruce soil microbiome also exhibited greater
12 resistance to seasonal changes between spring (wettest) and fall (driest). Mixed-species root zones
13 contained a hybrid of beech- and spruce-associated microbiomes. Several bacterial populations
14 exhibited either enhanced resistance or greater susceptibility to drought in mixed root zones.
15 Overall, patterns in the relative abundances of soil bacteria closely tracked moisture in seasonal
16 and latitudinal precipitation gradients and were more predictive of soil water content than other
17 environmental variables. We conclude that complementary differences in the drought resistance
18 of soil microbiomes can occur and the likeliest form of complementarity in mixed-root zones
19 coincides with the enrichment of drought-tolerant bacteria associated with spruce and the
20 sustenance of EMF by beech.

Key words: plant–soil interactions, forest soil microbiome, drought resistance, precipitation gradient, beech-spruce forest, and root complementarity.

1. Introduction

21 Plant species diversity positively correlates with ecosystem productivity (Hooper and
22 Vitousek, 1997; Liang et al., 2016; Tilman, 2001) and with increased resistance to extremes in
23 water availability, at least in grasslands (Craven et al., 2016; Isbell et al., 2015). Forest ecosystems
24 are vulnerable to the increasing frequency, intensity, and duration of drought caused by changing
25 precipitation patterns (Dai, 2013; IPCC, 2018). However, it remains to be proven whether forests
26 with higher plant diversity or functional richness are more resistant to drought than single species
27 ('species-specific') plantations (García-Valdés et al., 2021). The general relationship between
28 plant species diversity and productivity is, at least, partially due to the effects of biotic feedbacks
29 between plants and soil microorganisms (Hendriks et al., 2013; Schnitzer et al., 2011). Yet, to date,
30 research into diversity-productivity relationships in forests has been primarily focused on
31 aboveground parameters, e.g. annual growth (Paquette and Messier, 2011; Pretzsch et al., 2020,
32 2010). More recently, belowground parameters were found to differ between species-specific and
33 mixed-species forest stands, including tree root lifespan dynamics (Zwetsloot et al., 2019) and root
34 niche partitioning during drought (Altinalmazis-Kondylis et al., 2021; Zwetsloot and Bauerle,
35 2021). These observations raise questions about the potential influence of other belowground
36 biotic interactions on the drought resistance of mixed forest stands.

37 Belowground interactions between root systems of European beech (henceforth 'beech')
38 and Norway spruce (henceforth 'spruce'), a common forest type in Central Europe, can differ
39 between monoculture and mixed-species forests (Schmid and Kazda, 2002). Spruce and beech
40 exhibit anatomical (tracheid vs. xylem vessel elements) and physiological (isohydric vs.
41 anisohydric) differences that effect their hydrological processes and drought response (Lyr et al.,
42 1992; Pretzsch et al., 2013) and their root systems can interact in complementary ways that reduce
43 competition and facilitate access to soil moisture. These effects are evident in the differences

44 between species-specific and mixed root zones, where beech roots exhibit lower mortality than in
45 monospecific root zones, due to decreased intraspecific competition (Zwetsloot et al., 2019), and
46 where spruce roots maintain higher fine root mass during chronic drought (Zwetsloot and Bauerle,
47 2021). This complementarity is due to several aspects of root ecophysiology, which include: (i)
48 differences in rooting depth, with beech colonizing lower soil layers in mixed root zones (Bolte
49 and Villanueva, 2006), and (ii) diverging responses in fine root growth during drought, where
50 spruce become more dormant (cease growth and enhance suberization), while beech maintain a
51 degree of root activity, producing new, albeit shorter lived, fine root mass (Nikolova et al., 2020;
52 Rötzer et al., 2017). Beech and spruce also select for broadly different soil microbiomes (Bárta et
53 al., 2017; Uroz et al., 2016). Complementarity in the function and drought-tolerance between their
54 associated microbiomes might contribute to the altered plant drought responses where roots mix.

55 Root-associated microbes support the nutritional needs of trees (Brzostek et al., 2013;
56 Calvaruso et al., 2014; Gan et al., 2021; Nicolitch et al., 2016), and provide protection from
57 phytopathogens (Lehr et al., 2008; Terhonen et al., 2016) and the effects of mild drought (Pena et
58 al., 2013; Shi et al., 2002). Prior research has shown that the soil microbiome in mixed-species
59 forests can have greater functional and metabolic diversity than species-specific forests (Beugnon
60 et al., 2021; Prada-Salcedo et al., 2021). Thus, complementarity in the function and drought-
61 tolerance of soil microbiomes associated with beech and spruce might confer benefits during
62 drought. Several observations indicate the potential unevenness in the susceptibility or tolerance
63 of members of the beech- and spruce-associated microbiomes. The spruce soil microbiome is
64 characterized by a higher relative abundance of Actinobacteria, Armatimonadetes, and
65 Acidobacteria (and a higher denitrification potential) and greater proportion of ectomycorrhizal
66 fungi (EMF) relative to saprotrophic fungi than the beech soil microbiome, which has higher

67 proportions of Alphaproteobacteria, Planctomycetes, and Verrucomicrobia and different dominant
68 EMF taxa (*Scleroderma*, *Russula* and *Laccaria*) than spruce (*Boletus* and *Thelephora*) (Bárta et
69 al., 2017; Felsmann et al., 2015; Nacke et al., 2016; Uroz et al., 2016). Populations of
70 Alphaproteobacteria and Planctomycetes (beech-associated) are generally more sensitive to
71 drought, while Armatimonadetes and Actinobacteria (spruce-associated) are more tolerant
72 (Bastida et al., 2017; Bouskill et al., 2013; Chodak et al., 2015; Curiel Yuste et al., 2014; Evans et
73 al., 2014; Felsmann et al., 2015; Hartmann et al., 2017; Preece et al., 2019; Zeng et al., 2016; Zhou
74 et al., 2018). EMF also differ in their drought tolerance with more sensitive (ex. within *Laccaria*
75 and *Lactarius*) and more tolerant (ex. within *Russula*, *Cenococcum*, and *Scleroderma*) species
76 among those that associate with beech and spruce (Coleman et al., 1989; Di Pietro et al., 2007;
77 Ortega et al., 2004). Disparities in the drought tolerance or susceptibility of beech- and spruce-
78 associated microbiomes have yet to be experimentally tested.

79 With our study, we investigated whether the effects of experimental and natural reductions
80 in soil moisture on the structure of beech and spruce microbiomes differed in species-specific
81 (single species) versus mixed-species root zones. Drought-affected bacteria and fungi were
82 identified in a 5-year-old throughfall exclusion experiment and changes in these populations were
83 profiled in seasonal and latitudinal gradients in soil moisture and precipitation, respectively (Figure
84 1). The composition of bacterial and fungal communities and root-associated ‘rhizobacteria’
85 (rhizoplane and root endophytes) were determined with 16S rRNA gene and ITS region amplicon
86 sequencing. We expected EMF and rhizobacterial populations to be most sensitive to reductions
87 in soil moisture given the associated reduction in fine root mass (Zwetsloot and Bauerle, 2021).
88 We hypothesized that the impact of reduced soil moisture would affect beech and spruce
89 microbiomes differently, and that these responses would be further altered in soils where roots

90 mixed. Specifically, our experiment tested the null hypothesis that the responses of beech- and
91 spruce-associated bacteria and fungi would not differ in mixed-species root zones relative to
92 species-specific zones. We then tested whether similar differences occurred in response to seasonal
93 variation and latitudinal differences in soil moisture. Our experiment advances an understanding
94 of the effects of drought and natural gradients in soil moisture on beech and spruce soil
95 microbiomes and establishes the potential for complementation to occur where root systems mix.

2. Methods

96 2.1 Study overview and experimental design

97 In 2013, a latitudinal precipitation gradient was established in Bavaria, Germany (Pretzsch
98 et al., 2014). The gradient is composed of five forest sites containing 60 to 90-year-old Norway
99 spruce (*Picea abies* [L.] Karst) and European beech (*Fagus sylvatica* L.) which are (from driest to
100 wettest): Arnstein, Kelheim, Kranzberg, Wasserburg, and Traunstein (Figure 1A). All five sites
101 are similar in mean annual temperature (MAT) and soil type (Cambisol and Luvisol), but differ in
102 their mean annual precipitation (MAP; Table S1). Relative differences in total annual precipitation
103 were consistent among sites during the duration of the experiment (Table S2). However, periods
104 of drought occurred throughout Germany in summer in 2015 and 2018 (Schuldt et al., 2020).
105 Individual sites consist of a species-specific spruce and beech root zone, as well as a mixed spruce
106 and beech root zone with varying degrees of mixing (Table S3). A throughfall-exclusion
107 experiment was established at the middle site of the precipitation gradient, (Kranzberg Forest Roof
108 Experiment or ‘KROOF’; Figure 1B) and is composed of six drought plots, each with a species-
109 specific beech and spruce and mixed-species stand covered by a throughfall exclusion roof, and
110 six corresponding uncovered reference plots, with corresponding stand compositions. At the time
111 of the last sampling in 2018, seasonal throughfall had been excluded for five years, resulting in an
112 overall reduction of ~70% soil moisture across time and significantly greater pre-dawn water stress

113 for beech ($\bar{x} = -0.66 \pm 0.2$ MPa water potential) and spruce ($\bar{x} = -0.98 \pm 0.3$ MPa) compared to
114 reference plots (Grams et al., 2021).

115 *2.2 Soil and root sampling*

116 Soils from each root zone were sampled during the fall of 2017 (November), and the spring
117 (April), summer (July), and fall of 2018 (October) at all five sites and from the throughfall
118 exclusion experiment. During each sampling, ten soil cores (1.6 cm in diameter and 30 cm long)
119 were harvested from five plots (approximately 10 m²) randomly located within in each tree root
120 zone and/or throughfall exclusion treatment plot. The O₁ (litter layer) was manually removed, and
121 each soil core was divided into upper (0-8 cm deep) and lower soil layers (9-30 cm) comprised of
122 the O_f + _hA_h and A₁B_v, respectively (KA5 classifications; Nickel et al., 2018). The upper and lower
123 layers of ten cores were separately composited to yield 5 replicates per tree root zone per site
124 and/or per throughfall exclusion treatment. Five grams of the upper and lower composite sample
125 (excluding roots and soil particles > 2 mm) was weighed into Whirl-pak® bags in the field (Nasco
126 Sampling, Madison, WI, USA), and transported on dry ice to the Technische Universität München
127 (TUM) campus in Freising, where they were dried at 60 °C to a constant weight. Oven drying was
128 used to minimize the period in which cells are active and has negligible impact on bacterial and
129 fungal community composition (Castaño et al., 2016; Tzeneva et al., 2009). Dried samples were
130 shipped to Cornell University (Ithaca, NY, USA) and stored at room temperature until DNA was
131 extracted within two months. Over short periods, changes in microbiome composition due to air
132 drying and storage are minor (Clark and Hirsch, 2008; Lauber et al., 2010; Tzeneva et al., 2009).
133 The remaining soil samples were stored at TUM at 4 °C for additional analyses and to determine
134 fine root biomass (mg) (≤ 1 mm diameter), also dried to 60 °C to a constant weight. Prior to the
135 onset of KROOF, in 2013, fine root samples were taken from upper and lower layers in species-
136 specific spruce and beech root zones for the purpose of identifying root-associated bacteria

137 ('rhizobacteria') associated with each tree ($n = 20$), as opposed to 'soil bacteria,' which we use to
138 refer to the amplicon data generated from whole soil cores. Roots were thoroughly cleaned in tap
139 water, identified under a stereomicroscope and frozen until DNA was extracted. The bacterial
140 populations identified via this method correspond to a combination of 'rhizoplane' and root
141 endophyte populations.

142 *2.3 Analyses of soil properties*

143 The remaining soil from each sample was used to measure soil water content (SWC), pH,
144 and soil organic matter (SOM). SWC was averaged from triplicate measurements using 3 g of
145 fresh soil, which was calculated as the percentage mass lost after drying for 24 hours at 105°C.
146 Soil pH was measured using a VWR Scientific Products model 2000 pH meter, in a 1:10 (w/v)
147 dilution with distilled deionized water, after the sample was shaken for one minute and allowed to
148 settle for 10 minutes. SOM was estimated using loss on ignition method, according to (Howard
149 and Howard, 1990), by weighing 2 g of soil onto clay trays that were put in a muffle furnace where
150 temperature was slowly increased by 10 °C increments until a final temperature of 550 °C was
151 reached and maintained for 3 h.

152 *2.4 Analyses of soil bacterial and fungal communities and rhizobacteria*

153 DNA was extracted from 0.25 g of soil sample using the Qiagen PowerSoil kits
154 (Germantown, Maryland, USA) according to the manufacturer's protocol with a bead beating
155 treatment of 3 min at 5.5 m·s⁻¹ (Bio-spec, Bartellsville, OK, USA). DNA quantification and PCR
156 amplification of bacterial (V4 region of 16S rRNA gene; 515f/806r) and fungal phylogenetic gene
157 markers (ITS1; nBITS2f/58A2r) was performed as previously described (Sridhar et al., 2022).
158 Duplicate PCR reactions per sample were pooled prior to Illumina MiSeq (2 x 250 bp; v2)
159 sequencing, which was performed at Cornell Biotechnology Resource Center (Ithaca, NY) using
160 dual-indexed bar-coded primers (Koechli et al., 2019; Kozich et al., 2013). Seventy-five ITS

161 sequencing libraries from the Kranzberg throughfall-exclusion experiment were discarded due to
162 poor sequencing quality. These were discarded at random without impacting the balance of our
163 experimental design (Table S4). To identify rhizobacteria, DNA was extracted from 0.35-0.45 g
164 of root using the PowerSoil kit as previously described (Nickel et al., 2018) and 16S rRNA
165 amplicon libraries were prepared targeting rhizobacteria using the same methods (details in SI)
166 except that the V3-V4 region was targeted using primers optimized for plant DNA-rich samples
167 (335f/769r), as previously described (Dorn-In et al., 2015). All sample metadata is provided in
168 Table S5, and raw sequencing data was archived with the European Nucleotide Archive under the
169 BioProject accession: PRJEB36981 (data reference, see Wilhelm et al., 2022).

170 Sequencing data was processed using QIIME2 (v. 2020.2; Bolyen et al., 2019) with a
171 dependency on DADA2 (v. 1.10; Callahan et al., 2016) to assign sequences to operational
172 taxonomic units (i.e., amplicon sequence variants). Taxonomic classification was performed using
173 the QIIME2 ‘q2-feature-classifier’ trained on the Silva database (v. 132; Quast et al., 2013) and
174 UNITE database (v. 7.2; Nilsson et al., 2019) for bacteria and fungi, respectively. OTUs found in
175 the non-template controls and in low abundance were removed, namely those present in fewer than
176 three samples, or at a total relative abundance $< 0.01\%$. All counts were normalized by proportion
177 of total reads and presented as counts per thousand reads. R package *phyloseq* (v. 1.34; McMurdie
178 and Holmes, 2013) was used to characterize the communities and estimate diversity parameters on
179 rarified libraries ($n_{\text{bact}} = 15,300$ and $n_{\text{fungi}} = 2,544$ reads per sample). Fungal taxonomic
180 classifications were used to identify EMF, endophytic, and saprotrophic fungi using the FUNGuild
181 database (v. 1; Nguyen et al., 2016). Samples from throughfall-exclusion plots were removed
182 during all analyses of seasonal or latitudinal effects.

183 2.5 Bioinformatic and statistical analyses

184 The differential abundance of OTUs between reference and throughfall exclusion plots
185 (i.e., ‘drought-affected’) were identified using the *R* software package *indicspecies* (v. 1.7.9;
186 Cáceres and Legendre, 2009). Indicator species analysis was performed independently for soil
187 layers, season, and forest type, then combined and de-duplicated. Non-significant indicator OTUs
188 ($p_{\text{adj}} < 0.05$) and those with low indicator values (< 0.35) were excluded. Alpha-diversity was
189 measured as species richness, Shannon diversity, and Pielou’s evenness. Beta-diversity was
190 assessed using Bray-Curtis dissimilarity and differences in community composition were
191 visualized using principal coordinates analysis and tested with PERMANOVA ($n_{\text{perm}} = 999$) using
192 ‘adonis’ from the *R* package *vegan* (v. 2.5.7; Oksanen et al., 2015). Resistance (R) to change in
193 microbiome composition due to throughfall exclusion was measured as the Bray-Curtis
194 dissimilarity between reference and exclusion plots (D), such that $R = 1 - D$. A lower R value
195 indicates a greater dissimilarity between reference and throughfall exclusion plots, indicating a
196 greater drought effect (De Vries and Shade, 2013). R was calculated for the average of all
197 permutations of replicates between reference and exclusion plots. The environmental and
198 microbiome features that were most predictive of SWC were identified using random forest-based
199 feature selection implemented in the *R* package *Boruta* (v. 7.0; Kursu and Rudnicki, 2010).
200 Features included in model selection were environmental (soil layer, root zone, site, season, year)
201 and microbiome (aggregated counts at the taxonomic rank of Order scaled with the ‘scale’ function
202 in *R*). Subsequently, Pearson’s correlations between selected bacterial and fungal features and
203 SWC were performed using ‘rcorr’ from the *R* package *Hmisc* (v. 4.5; Harrell and Dupont, 2015).
204 Differences in the relative abundance of OTUs among tree root zones were tested with ANOVA
205 (‘aov’ function in *R*). The main effects and interactions between throughfall exclusion, season, or
206 tree root zone on the relative abundances of OTUs, were determined by fitting to fixed effects

207 linear models ('lm' function). P-values were adjusted for multiple test correction according to the
208 Benjamini and Hochberg false discovery rate. Only the most abundant OTUs (> 0.05% of sample
209 reads) were included in statistical testing. Significant effects are denoted by asterisk: $p < 0.05$ (*),
210 $p < 0.01$ (**), and $p < 0.001$ (***). All analyses can be reproduced with scripts included in the
211 Supplementary Data package available through the Open Science Foundation (doi:
212 10.17605/OSF.IO/DN9CH).

3. Results

213 Our initial analyses were performed to establish the drought susceptible and tolerant
214 (section 3.1) and spruce- and beech-associated soil microbiome and rhizobacteria (section 3.2)
215 prior to testing whether the response of these groups differed in species-specific versus mixed root
216 zones (section 3.3). Additional analyses were performed to characterize the natural variation
217 (seasonal and precipitation gradient) in drought-affected populations among tree root zones and
218 soil layers (section 3.4).

3.1.1 General effects of throughfall exclusion on soil properties and the soil microbiome

220 Across the full 5-year experimental period, throughfall exclusion resulted in a 70%
221 reduction in SWC, on average, during the growing season (Grams et al., 2021) and, at the time of
222 our sampling, the throughfall exclusion plots had significantly lower SWC in upper ($\bar{x} = -27.8\%$)
223 and lower soil layers ($\bar{x} = -22.4\%$) across all seasons (Figure S1). Upper layers had higher average
224 SWC though the extent of moisture reduction was comparable in both layers ($t = -3.9$ vs. -3.2 ,
225 respectively). Throughfall exclusion significantly decreased fine root mass, with a greater effect
226 on upper layers ($\bar{x} = -51.1\%$; $t = -4.2$; $p < 0.001$) compared to lower layer soils ($\bar{x} = -29.8\%$; $t = -$
227 3.2 ; $p < 0.001$). Exclusion had minor effects on soil pH (+1% in upper layer; $p = 0.03$); and had
228 mostly insignificant effects on SOM and DNA yield (Figure S1). Throughfall exclusion accounted
229 for a relatively low proportion of variation in the beta-diversity of soil bacterial and fungal

230 communities (Figure 2AB), which was primarily attributable to differences among soil layer and
231 tree root zone. Communities from mixed root zones exhibited an intermediate similarity between
232 species-specific root zones (Figure 2A).

233 3.1.2 Drought-affected soil bacteria and fungi

234 The differences in the relative abundance of OTUs between reference and throughfall
235 excluded plots were used to identify ‘drought-favored’ or ‘drought-sensitive’ bacterial and fungal
236 populations. The soil microbiome at Kranzberg was comprised of 10,357 bacterial and 3,282
237 fungal OTUs. Of these, a total of 369 bacterial and 23 fungal OTUs were indicative of throughfall
238 exclusion (Table S6). More OTUs were drought-sensitive ($n_{\text{bact}} = 194$ and $n_{\text{fungi}} = 15$) than drought-
239 favored ($n_{\text{bact}} = 175$ and $n_{\text{fungi}} = 8$). More drought-affected OTUs were observed in upper ($n = 215$)
240 versus lower soil layers ($n = 130$), but OTUs did not differ in their likelihood of being drought-
241 affected based on their soil layer association (Fisher’s test; $p = 0.8$).

242 Drought-favored fungi were primarily classified as Ascomycota (80%) and belonged to
243 genera designated as saprotrophic (*Pseudogymnoascus*, *Niesslia*, and *Ciliolarina*). Two of the
244 eight drought sensitive fungal OTUs were classified to genera of EMF (*Inocybe* and *Lactarius*).
245 However, the overall relative abundance of EMF was largely unaffected by throughfall exclusion
246 (Figure S2A). Fungal endophytes (genus *Phialocephala*) were significantly more abundant in
247 throughfall-excluded upper soils (Figure S3A) and were relatively more abundant in spruce root
248 zones (Figure S3B).

249 Throughfall exclusion affected a phylogenetically diverse group of bacteria, but most of
250 the affected OTUs were classified to Planctomycetes, Alphaproteobacteria and Actinobacteria
251 (Figure S4; complete list in Table S6). Actinobacteria were near uniformly drought-favored
252 (primarily classified as Pseudonocardiales, Solirubrobacterales, and Micrococcales), while
253 Planctomycetes were primarily drought-sensitive (Pirellulales, Gemmatales, and

254 Planctomycetales), though certain clades of Planctomycetes also contained a high proportion of
255 drought-favored taxa (Tepidisphaerales and Isosphaerales). Within Alphaproteobacteria, several
256 orders were favored by throughfall exclusion, namely populations of Caulobacterales
257 (*Phenylobacterium*), Elsterales, and Sphingomonadales (*Sphingomonas*) while others were
258 sensitive, namely populations of Rhizobiales (*Roseiarcus*, *Rhodoplanes*, and *Bradyrhizobium*) and
259 Reyranelles (*Reyranelia*). Other major bacterial groups that were drought sensitive included:
260 Deltaproteobacteria (Myxococcales), Verrucomicrobia (Pedosphaeraceae and
261 Xiphinematobacteriaceae), and Dependuntiae (Vermiphilaceae).

262 3.2.1 Beech- and spruce-associated soil bacteria and fungi and rhizobacteria

263 Soil microbiome composition significantly differed between beech and spruce root zones,
264 though the most abundant bacterial and fungal OTUs ($\geq 0.2\%$ of total reads) were present in all
265 root zones (97% of bacteria and 52% of fungi; Figure S5). The relative abundances of beech- and
266 spruce-associated OTUs in mixed zones were characteristically intermediate between the relative
267 abundance in the species-specific zones (Figure S6). The spruce-associated microbiome was
268 dominated by several orders of Actinobacteria (Frankiales and Solirubrobacterales) and
269 Acidobacteria (Acidobacteriales and Solibacterales; Figure 3A) and had a higher
270 Ascomycota:Basidiomycota ratio compared to beech (Figure S7). The beech microbiome had a
271 higher proportion of Proteobacteria (Alpha-, Gamma- and Delta-), Verrucomicrobia and
272 Planctomyces and a higher proportion of EMF (Figure S2B). Beech-associated taxa were
273 significantly more likely to be drought-sensitive than spruce-associated taxa (Fisher's Exact, O.R.
274 = 23.8; $p < 0.001$), with spruce-associated taxa tending to increase in relative abundance in
275 throughfall excluded plots.

276 Rhizobacterial populations associated with beech and spruce roots were identified using
277 indicator analysis. Beech-associated rhizobacterial populations were enriched in

278 Alphaproteobacteria from the family Xanthobacteraceae relative to spruce, which were enriched
279 in Acidobacteriales (*Acidipila*, *Granulicella* and *Occallatibacter*) and Frankiales (*Acidothermus*;
280 Figure 3B, complete list in Table S7). Bacterial species richness and evenness were significantly
281 higher in beech than spruce root zones, while no differences were evident for fungi (Figure S8).

282 3.2.2 Corresponding differences in plant and soil properties among root zones

283 Trends in the soil microbiome corresponded primarily with differences in fine root mass
284 and soil organic matter content in root zones. The upper soil layer in spruce root zones had a
285 significantly higher percentages of soil organic matter (2-fold) and SWC than beech, except for at
286 Arnstein, which received the lowest precipitation (Figure S9A). In contrast, the beech root zone
287 had significantly higher fine root mass in both the upper (66%) and lower layers (60%) at all sites,
288 while mixed zones had intermediate root mass. In all sites, fine root mass followed seasonal trends
289 in SWC between spring and fall, increasing in wetter sites (Wasserburg and Traunstein) and
290 declining in drier sites (Arnstein and Kranzberg), with the trend most pronounced in beech stands
291 (Figure S9B). Soil pH was significantly more acidic in spruce ($\bar{x} = 4.04$) than beech ($\bar{x} = 4.30$),
292 with intermediates values in mixed root zones ($\bar{x} = 4.21$).

293 3.3.1 Contrasting the effects of soil moisture on microbiomes among root zones

294 We tested for the generalized effects of soil moisture reduction on soil bacteria and fungi
295 among root zones using the community resistance (R) metric. In our study, lower R values indicate
296 a greater change in community composition resulting from reduced soil moisture (i.e., a lower
297 resistance). On average, R values for bacterial communities differed by tree root zone, but not for
298 fungal communities. Bacterial communities in the spruce root zones had higher R values than in
299 the beech root zones, though this difference was only significant in the lower soil layer (Figure
300 4A). Differences in R among root zones corresponded with the proportion of drought-favored
301 (higher in spruce) and drought-sensitive bacterial OTUs (higher in beech; Figure 4B). The spruce

302 soil microbiome also exhibited greater R values in relation to seasonal changes between spring
303 (wettest) and fall (driest) at Kranzberg and, in this case, the mixed root zone also exhibited
304 significantly higher R than species-specific beech (Figure S10A). These differences also
305 corresponded to significant seasonal increases in drought-favored and drought-sensitive taxa
306 (Figure S10B). R values for fungal communities were variable and did not significantly differ
307 among tree root zones (Figure S11A). Yet, the relative abundance of drought-favored fungal OTUs
308 was also significantly higher in spruce root zones (Figure S11B).

309 *3.3.2 Contrasting responses in the soil microbiome of species-specific and mixed root zones*

310 The effects of throughfall exclusion on several abundant beech- and spruce-associated
311 bacteria were significantly altered in the mixed-species root zone. Several beech- and spruce-
312 associated OTUs that were affected by throughfall exclusion in species-specific root zones
313 exhibited enhanced tolerance in mixed root zones (Figure 5AB; Table S8). Conversely, several
314 OTUs that were unaffected by throughfall exclusion in species-specific root zones exhibited
315 different responses in mixed-species root zones. These could be divided into beech-associated
316 OTUs, that exhibited a greater susceptibility (Figure 5C), and spruce-associated OTUs, that
317 exhibited a greater resistance, in mixed-species root zones (Figure 5D; Table S9). These trends
318 were evident when profiling the same populations across seasonal differences in SWC at
319 Kranzberg (lower panels in Figure 5C and 5D), and at other sites, though not uniformly (Figure
320 S12). There were no significant interactions between root mixing and throughfall exclusion for
321 soil fungi.

322 *3.4.1 Seasonal and latitudinal trends in drought-affected populations*

323 Complementary to the throughfall exclusion experiment, we profiled changes in the soil
324 microbiome of species-specific and mixed-species root zones across seasonal and latitudinal
325 gradients in soil moisture and precipitation, respectively (Figure 1A). A random forest, decision-

326 tree-based feature selection was used to identify the environmental variables and microbial taxa
327 that were most predictive of soil water content in these gradient (Table 1). Several bacterial orders
328 were selected as predictors of SWC, and the abundance patterns of Rhizobiales outranked all other
329 features in importance, including environmental parameters (Table 1). No fungal taxa were
330 selected as predictors of SWC. The abundance patterns of the top microbiome-based predictors of
331 SWC closely tracked seasonal and latitudinal differences in precipitation in upper (Figure 6) and
332 lower soil layers (Figure S13) and were among the same groups impacted by throughfall exclusion.
333 Among the OTUs identified as drought-affected in data from the throughfall exclusion experiment,
334 the number that significantly differed among high and low precipitation sites in the latitudinal
335 gradient was lowest in spring (n = 92), when SWC was highest, and progressively increased in
336 summer (n = 133) and fall (n = 220).

337 EMF did not exhibit a clear relationship to SWC as their relative abundance did not vary
338 significantly by season or site position in the precipitation gradient. The exception was at Arnstein
339 during the summer, when increased relative abundance of EMF coincided with some of the lowest
340 SWC values measured (Figure S2B). The EMF taxa that increased in Arnstein during the summer
341 included taxa from the genera: *Amphinema*, *Clavulina*, *Otidea*, *Sebacina*, *Tricholoma*, *Inocybe*,
342 and *Lactarius*.

343 3.4.2 Moisture related shifts in rhizobacterial abundance between soil layers

344 Reductions in soil moisture can cause stratification of roots in spruce, to upper soil layers,
345 and beech, to lower soil layers, according to observed root behavior (Bolte and Villanueva, 2006).
346 We found that the relative abundance of spruce- and beech-associated rhizobacteria tended to
347 differ between upper and lower soil layers correspondingly, particularly at Kranzberg (Figure 7).
348 The enrichment of spruce-associated rhizobacteria in upper layer soils was even more pronounced
349 during fall, when SWC was lowest, though this interaction was only significant at Kranzberg and

350 Wasserburg (Figure 7). In contrast, beech-associated rhizobacteria were more prevalent in the
351 lower soil layer, and no shift was seen due to throughfall exclusion or season. Additionally, at
352 Kranzberg, the seasonal shift to the upper soil layer was most pronounced in mixed root zones
353 ($t_{\text{interact.}} = 8.1$; $p < 0.001$) compared to species-specific spruce ($t = 3.9$) or beech root zones ($t =$
354 3.6). No shifts in spruce- and beech-associated fungi were evident, though we did not obtain fungal
355 sequence data from rhizoplane / endophyte samples.

4. Discussion

356 Our study reveals major differences in the response of Norway spruce and European beech
357 soil microbiomes to changes in soil moisture and their interplay in mixed root zones. A
358 significantly higher proportion of beech-associated taxa were sensitive to reductions in soil
359 moisture compared to the more drought-tolerant populations associated with spruce. Furthermore,
360 several populations of rhizobacteria and soil bacteria had increased resistance or susceptibility to
361 drought in mixed-species root zones relative to species-specific. These findings lead us to reject
362 the hypothesis that the response of microbial populations to changes in moisture would not differ
363 in mixed-species root zones, allowing for the possibility that complementation can occur. Here,
364 we discuss the possible reasons why drought may impact spruce- and beech-associated soil
365 microbiomes differently, and the potential consequences of the phenomena observed in mixed-
366 species root zones.

367 *4.1 Explaining the contrasting effects of reduced soil moisture on beech and spruce microbiomes*

368 Taxa that were consistently impacted by experimental or natural reductions in soil moisture
369 were broadly characteristic of the spruce and beech microbiomes described in prior research
370 (Asplund et al., 2019; Bárta et al., 2017; Felsmann et al., 2015; Nacke et al., 2016; Uroz et al.,
371 2016). Several of these broad taxonomic groups were previously shown to be impacted by reduced
372 throughfall in beech and spruce forest (Felsmann et al., 2015), and other forest types (Bastida et

373 al., 2017; Bouskill et al., 2013). Our results demonstrate that soil bacterial communities in species-
374 specific spruce root zones have a higher resistance to the impacts of drought and to seasonal
375 reductions in soil moisture than in beech root zones (Figure 4; Figure S10). This difference
376 corresponded with a higher proportion of drought-tolerant bacteria in the spruce microbiome than
377 in beech, and higher proportions of drought-sensitive bacteria in the beech microbiome. While this
378 is the first time such broad differences in moisture sensitivity have been reported, prior research
379 has shown that the mineral weathering activity of rhizobacteria isolated from beech was higher
380 during wetter periods while the reverse was true for those isolate from spruce (Collignon et al.,
381 2011).

382 Prevailing differences in soil conditions between spruce and beech stands likely help explain
383 the disparity in drought tolerance and sensitivity in their respective soil microbiomes.
384 Physiological differences in rooting depth result in consistently drier conditions in shallow soils in
385 spruce stands relative to beach (Allen et al., 2019; Zwetsloot and Bauerle, 2021). Our findings
386 indicate that this effect is large enough to select for higher proportion of drought stress-tolerant
387 bacteria in spruce soils. Additionally, soil acidification was more pronounced in spruce root zones
388 relative to beech (Figure S1), as indicated by the high relative abundance of acidophilic indicator
389 taxa (Acetobacterales, Frankiales and Acidobacteriales), as previously shown (Sridhar et al.,
390 2022a). Our observations were consistent with current understanding about difference in soil
391 development under beech and spruce stands caused by root and litter chemistry, nutrient leaching
392 and uptake, and mineralization and nitrification rates (Cremer and Prietzel, 2017). Spruce roots
393 and litter contain significantly higher concentrations of polyphenols than beech (Kuiters and
394 Denneman, 1987; Zwetsloot and Bauerle, 2018), which are generally toxic to soil heterotrophs
395 (Adamczyk et al., 2013; Chunmei et al., 2010; Inderjit et al., 2009; Metsämuuronen and Sirén,

396 2019). The high acidity and polyphenol content of spruce soils create adverse growing conditions,
397 which retard decomposition (Albers et al., 2004; Berger et al., 2004) and reduce soil respiration
398 and biomass in spruce stands relative to mixed-species and beech stands (Borken et al., 2002;
399 Borken and Beese, 2005; Lu and Scheu, 2021). These adverse conditions, along with consistently
400 lower shallow soil moisture levels in spruce stands, may select for stress-tolerant populations,
401 which may better endure water stress. This theory is supported by the overlapping physiological
402 stress response to acidity, low osmolarity, and desiccation observed in model bacteria (Ait-
403 Ouazzou et al., 2012; Hengge-Aronis, 2002; Spector and Kenyon, 2012). It is also anecdotally
404 supported by the enrichment of the thermophilic, acid-tolerant genus *Acidothermus* by throughfall
405 exclusion in spruce and mixed root zones (Figure S14), which are characteristic of droughted and
406 arid soils (Eppard et al., 1996; Lacerda-Júnior et al., 2019).

407 The prevalence of drought-sensitive taxa in beech-specific soils could also reflect differences
408 in the degree of rhizosphere activity between beech and spruce. Beech produce more fine root
409 mass (Finér et al., 2007), apparent at all our sites (\bar{x} = 1.4 to 2.1-fold higher), and sustain more
410 microbial biomass and higher soil respiration rates than spruce (Borken et al., 2002; Borken and
411 Beese, 2005; Lu and Scheu, 2021). Thus, the apparent sensitivity of beech-associated taxa may
412 reflect the inability of beech roots to sustain basal levels of microbial activity during drought. The
413 potential diminished influence of beech on the soil microbiome during drought was evident in the
414 decline of microbial populations indicative of higher trophic complexity, including members of
415 the Vermiphilaceae, endosymbionts of amoeba (Delafont et al., 2015), and Candidatus
416 *Xiphinematobacter*, symbionts of nematode (Rius et al., 2021; Vandekerckhove et al., 2000).
417 While we cannot disentangle the relative contributions of roots versus litter to differences in the
418 soil microbiomes between spruce and beech, root traits have a far greater importance in explaining

419 the physicochemical properties of species-specific and mixed forest soils than litter traits (Cesarz
420 et al., 2013; Gillespie et al., 2021).

421 4.2 Effects of root mixing on the impacts of reduced soil moisture

422 Most beech- and spruce-associated fungi or bacteria occurred at an intermediate relative
423 abundance in mixed root zones. This result indicates that the influence of each tree follows a
424 gradient and that taxa associated with either tree species can co-occur. These observations are
425 consistent with prior reports of intermediate heterotrophic activity (Borken et al., 2002; Borken
426 and Beese, 2005), microbial biomass (Lu and Scheu, 2021), litter decomposition rates (Albers et
427 al., 2004), and mineral weathering (Cremer and Prietzel, 2017) in mixed root zones. Thus, the
428 hybrid soil microbiome in mixed root soils reflects the combined, but weakened, influences of
429 each tree.

430 The resistance of several beech and spruce-associated bacteria to throughfall exclusion was
431 enhanced in mixed root zones (Figure 5AB). These populations included members of
432 Acidobacteria (*Bryobacter* and *Occallatibacter*, both spruce-associated rhizobacteria), which are
433 reputed for their production of extracellular polymeric substances, which confer stress tolerance
434 to cells and may influence soil moisture dynamics (Foesel et al., 2016; Kielak et al., 2016;
435 Kulichevskaya et al., 2010). In contrast, several members of Xanthobacteraceae (*Bradyrhizobium*
436 and *Rhodoplanes*) were more susceptible to throughfall exclusion in mixed root zones (Figure 5C).
437 Xanthobacteraceae are known to increase in abundance in proximity to beech trunks (Nacke et al.,
438 2016), suggesting that their capacity to resist drought may depend on the density of beech roots or
439 litter. Conversely, several spruce-associated taxa exhibited enhanced resistance to drought
440 exclusively in mixed roots zones (Figure 5D), suggesting some populations may benefit from the
441 reduced competition where drought-sensitive, beech-associated taxa are diminished. Our
442 explanations for trends in mixed root zones remain to be tested, but these observations illustrate

443 the range of interactions occurring where the influence of trees and, by extension, their
444 microbiomes mix.

445 4.3 Effects of reduced soil moisture on root-associated taxa

446 The rhizosphere activity of spruce and beech varies by season and moisture availability, with
447 the highest activity in spring (Calvaruso et al., 2014) and diminished photosynthate-derived
448 microbial rhizosphere biomass during drought (Ruehr et al., 2009). We hypothesized that
449 reductions in soil moisture would disproportionately impact root-associated taxa, like EMF, which
450 are important contributors to nutrient acquisition by spruce and beech (Brzostek et al., 2013;
451 Calvaruso et al., 2014; Collignon et al., 2011; Gan et al., 2021; Nicolitch et al., 2016). Contrary to
452 expectations, we did not observe any compositional changes in soil EMF populations from
453 throughfall exclusion, seasonal variation, or across the precipitation gradient. The lack of shift in
454 EMF populations is not without precedent in spruce and beech forests (Gorfer et al., 2021; Nickel
455 et al., 2018; Žifčáková et al., 2015). The apparent resiliency of EMF to changes in community
456 composition may reflect their direct connection to live roots and a privileged access to root
457 exudates relative to bacteria, consistent with the more severe impacts in bacterial populations we
458 observed and reported elsewhere (de Vries et al., 2018; Fuchslueger et al., 2014). Access to
459 exudates and shelter within the root might explain why endophyte populations (*Phialocephala*)
460 were among the few fungi favored by throughfall exclusion (Figure S3).

461 Rhizobacteria were among the most affected by throughfall exclusion and seasonal and
462 latitudinal variation in soil moisture. Members of the Rhizobiales and Burkholderiaceae were
463 greatly diminished by throughfall exclusion in both beech and spruce root zones (Figure 6; Figure
464 S15). These drought-sensitive populations were dominated by members of *Bradyrhizobium* and
465 *Rhodoplanes* (Xanthobacteraceae) and *Caballeronia* and *Paraburkholderia* (Burkholderiaceae)
466 which are consistently associated with the rhizosphere of European beech and Norway spruce

467 (Colin et al., 2017; Uroz et al., 2016) and are common root- and mycorrhizae-associated bacteria
468 isolated from forest soils (Burke et al., 2008; Izumi et al., 2007, 2006; Kataoka et al., 2008; Tanaka
469 and Nara, 2009; Uroz et al., 2012; Wilhelm et al., 2020). The apparent drought sensitivity of these
470 rhizobacteria has implications for soil nutrient cycling and plant nutrition, given their involvement
471 in priming decomposition (Wilhelm et al., 2021; Zwetsloot et al., 2020), phosphate solubilization
472 in the beech and spruce rhizosphere (Lepleux et al., 2012; Nicolitch et al., 2016), and endophytic
473 nitrogen fixation (Puri et al., 2020).

5. Conclusions

474 We conclude that any potential complementation of soil microbiome function during drought
475 is likely to confer greater benefits to beech than spruce, given the higher proportion of drought-
476 favored spruce-associated taxa and their sustained resistance in mixed root zones. The enhanced
477 drought tolerance of beech in mixed stands with spruce has been reported (Schäfer et al., 2017),
478 though the benefits of mixing are more commonly reported to favor Norway spruce (del Río et al.,
479 2014; Ding et al., 2017; Rötzer et al., 2017). Thus, it remains to be seen what functions the
480 complementation of soil microbiomes might have in the drought resistance of mixed beech-spruce
481 forests. The impact of reduced soil moisture was greater on rhizobacteria than EMF populations,
482 with a higher proportion of EMF occurring in the beech soil microbiome. Thus, the likeliest form
483 of complementarity in mixed-root zones might correspond with the enrichment of drought-tolerant
484 populations by spruce and the sustenance of EMF activity by beech.

485 Forests and their soil microbiome are complex adaptive systems in which legacy and context
486 shape biological responses to water limitation (Bouskill et al., 2013). Yet, the moisture sensitivity
487 of major drought-affected populations, such as members of the Rhizobiales, were consistent across
488 sites and season, providing evidence for the potential widespread occurrence of phenomena
489 reported here. Future research is needed to understand the ecological and environmental drivers of

490 differences across sites and to test the functional consequences of complementation between
491 microbiomes of beech and spruce, and other abundant tree species, in mixed root zones.

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Data Availability Statement

943 All sample metadata is provided in Table S5. All analyses can be reproduced with scripts included
944 in the Supplementary Data package available through the Open Science Foundation (doi:
945 10.17605/OSF.IO/DN9CH). Raw sequencing data was archived with the European Nucleotide
946 Archive under the BioProject accession: PRJEB36981.

Author Contributions

947 RCW performed the data analysis, research, and writing. JMU contributed to data analysis
948 and research. Field sampling was performed by JMU, TLB, MG, FW, and KP and sample
949 processing (soil parameters, root biomass) by JMU and FW. Amplicon sequencing libraries were
950 prepared by RCW, JMU, and FW. TLB, KP, and DHB guided all research efforts, including
951 analyses and writing. The authors declare we have no competing or conflicts of interest.

Tables

952 **Table 1.** The top environmental and soil microbiome predictors of soil water content were
953 identified, and ranked, using Boruta random forest-based feature selection. The relative abundance
954 pattern of aggregated counts of Rhizobiales was ranked as the best predictor of SWC, followed by
955 soil layer and precipitation gradient site. Fungal and bacterial count data, aggregated by Order,
956 were included in feature selection, but only bacteria had predictive value. Bacterial orders that had
957 significant Pearson's correlations with SWC content in both layers ($r > |0.2|$ and $p_{\text{adj}} < 0.05$) and
958 were among the top 20 most important features are displayed (full list in Table S10). The family
959 Gimesiaceae is historically known as 'Planctomycetaceae.'

Figures

960 **Figure 1.** A schematic overview of the experimental design used to examine the effects of natural
961 and experimental variation in soil moisture on the soil microbiome of species-specific and mixed-
962 species stands of European beech and Norway spruce. In (A), samples were collected in spring,
963 summer, and fall and at sites spread across a natural precipitation in Bavaria, Germany. In (B), at
964 Kranzberg, the mid-point of the gradient, samples were taken from a five-year old throughfall
965 exclusion experiment where rain-out roofs had been used to reduce soil moisture during the
966 growing season. Each symbol accurately reflects the distribution of trees species (full details in
967 Rötzer *et al.* 2017). In (C), at all sites, and in the throughfall exclusion plots at KROOF, samples
968 were taken from species-specific or mixed-species tree root zones. Ten soil cores (30 cm deep)
969 were randomly sampled from five to six plots for each tree root zone at each site and during every
970 season and subdivided into upper and lower soil layers before being composited.

971 **Figure 2.** The composition of soil bacterial and fungal communities at Kranzberg differed
972 primarily by soil layer and tree root zone, as evident in (A) the grouping of samples by principal
973 coordinates analysis and (B) the proportion of variance explained (R^2) in a PERMANOVA
974 analysis based on the Bray-Curtis dissimilarity in community composition.

975 **Figure 3.** Beech and spruce root zones had marked differences in the taxonomic composition of
976 soil bacteria and fungi and rhizobacteria. In (A), the bar plots provide a summary of the relative
977 proportions of indicator OTUs for beech- and spruce-associated bacteria ($n = 506$) and fungi ($n =$
978 63) according to the ratio of their aggregated relative abundance at rank Order. The subset of tree-
979 associated taxa affected by drought are labeled on the y-axis. In (B), the taxonomic profile or
980 rhizobacteria associated with either beech or spruce according to indicator analysis using 16S
981 rRNA gene amplicon data generated from root material.

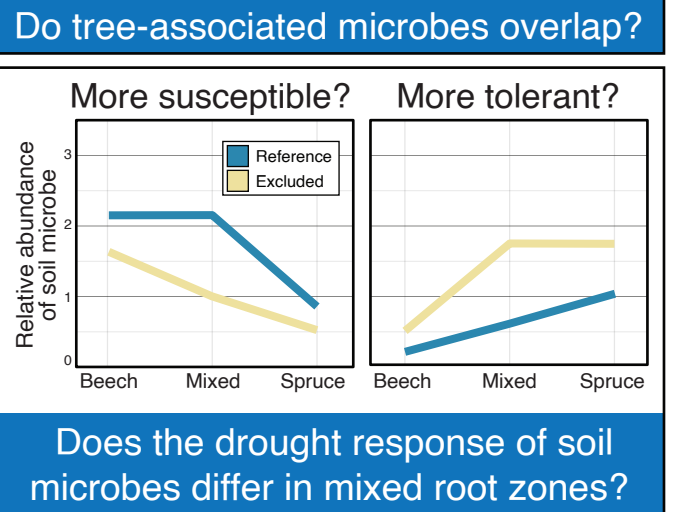
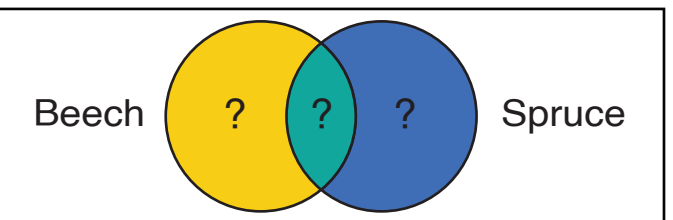
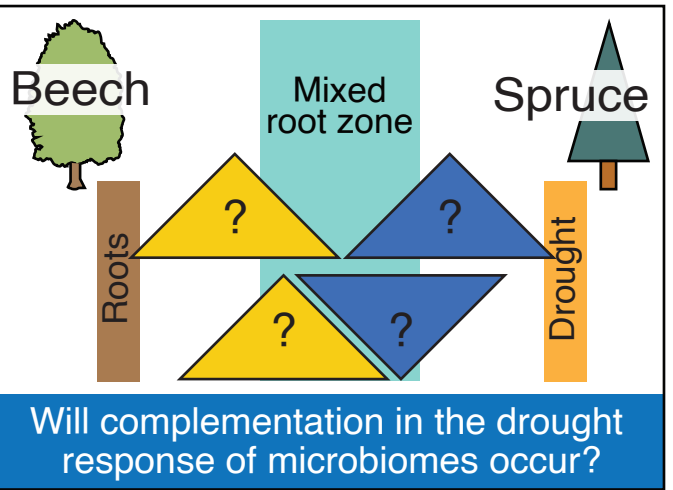
982 **Figure 4.** The resistance (R) of soil bacterial communities to the effects of throughfall exclusion
983 in soil moisture was greatest in spruce root zones at Kranzberg. Differences in R among root zones
984 (B) corresponded with the relative abundance of drought-favored and drought-sensitive
985 populations. Pairwise differences in resistance among tree root zones sites were tested using
986 TukeyHSD ($p_{\text{adj}} < 0.05$). In (B), the effects of throughfall exclusion ($p_{\text{excl.}}$) were significant for all
987 groups. Any significant differences between tree root zone were denoted by bars with asterisk and
988 interactions between throughfall treatment and tree root zone were denoted by asterisk.

989 **Figure 5.** The responses of beech- and spruce-associated OTUs to throughfall exclusion were
990 significantly altered in soil microbiome of mixed root zones at Kranzberg. In (A), the relative
991 abundance of several drought-sensitive beech-associated OTUs ($n = 10$) were largely
992 undiminished in mixed root zones. In (B), several drought-sensitive spruce-associated OTUs ($n =$
993 5) had enhanced resistance to throughfall exclusion in mixed root zones. Several OTUs that were
994 not significantly affected by throughfall exclusion in species-specific root zones showed, in (C),
995 an increased susceptibility ($n = 8$) or, in (D), an increased resilience ($n = 4$) in mixed zones. These
996 trends were consistent with trends in relative abundances between spring (wettest) and fall (driest
997 season) at Kranzberg (lower panel of C and D). Individual OTU exhibited significant interactions
998 between throughfall exclusion and tree root zone (Table S8 and Table S9) but were displayed in
999 aggregate. Any significant differences between tree root zone were denoted by bars with asterisk
1000 and interactions between throughfall treatment (or season) and tree root zone were denoted by
1001 asterisk.

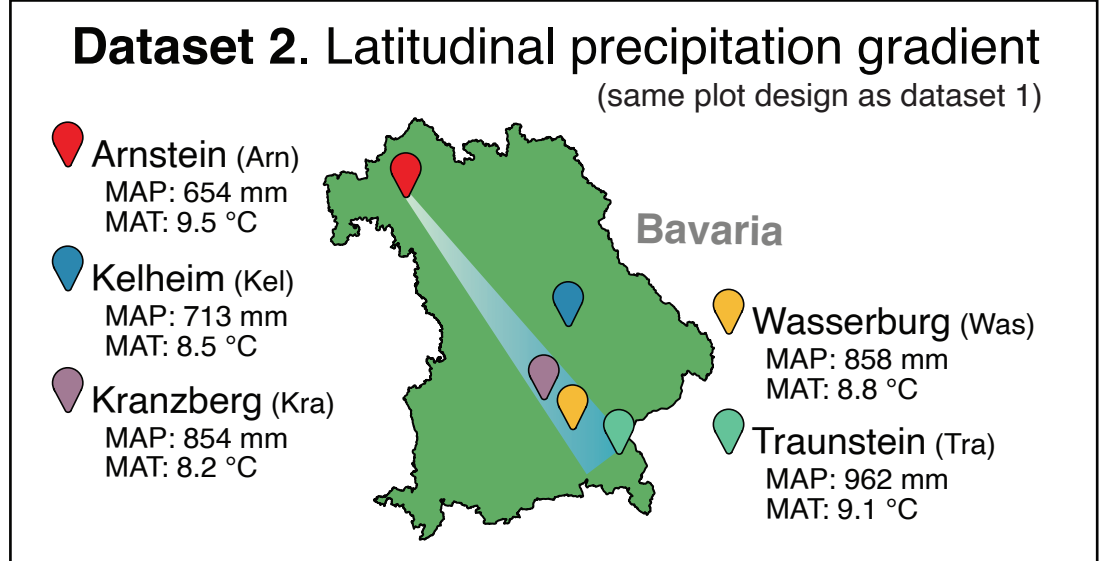
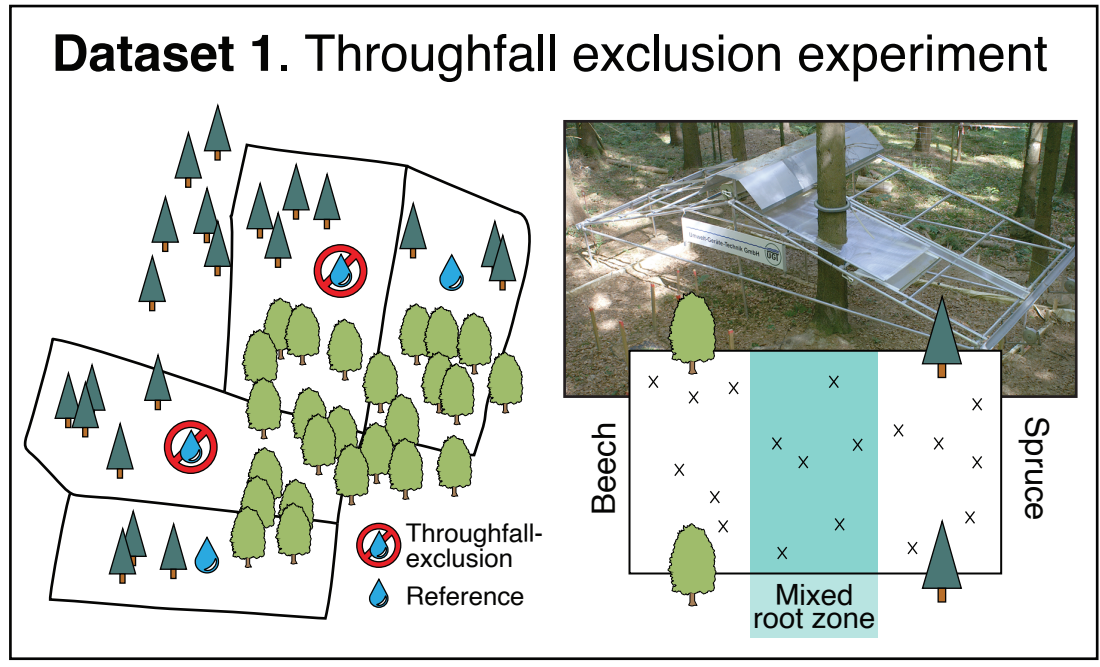
1002 **Figure 6.** The relative abundance of four major bacterial orders followed changes in soil water
1003 content caused by throughfall exclusion plots, seasonal variation, and across the latitudinal
1004 precipitation gradient. In (A), the soil water content in the upper soil layer was reduced by
1005 throughfall exclusion (box plot), and across seasons (x-axis) and gradient sites (lines). In (B), the
1006 relative abundances of the four bacterial Orders identified in feature selection as predictive of SWC
1007 (Table 1). The same trends were evident in lower layer soils, though less pronounced (Figure S13).
1008 Significant differences between mean SWC or relative abundance between seasons were denoted
1009 by bars with asterisks. Significant interactions between season and site were denoted by colored
1010 asterisks. Lettering denotes significant differences among sites according to TukeyHSD ($p_{adj} <$
1011 0.05).

1012 **Figure 7.** Shifts in the relative abundance of rhizobacteria in response to (A) throughfall exclusion
1013 or (B) seasonal differences between spring and fall revealed the putative partitioning of roots
1014 between soil layers. Spruce-associated rhizobacteria (upper panel) tended to occur at higher
1015 proportions in the upper soil layer, while beech-associated rhizobacteria (lower panel) tended to
1016 be more abundant in the lower layer. Significant differences between soil layer means were
1017 denoted by bars with asterisk and interactions between throughfall treatment (or season) and soil
1018 layer were denoted by asterisk.

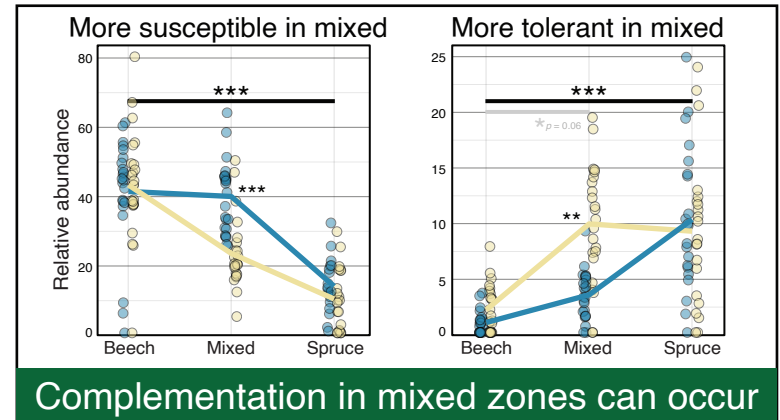
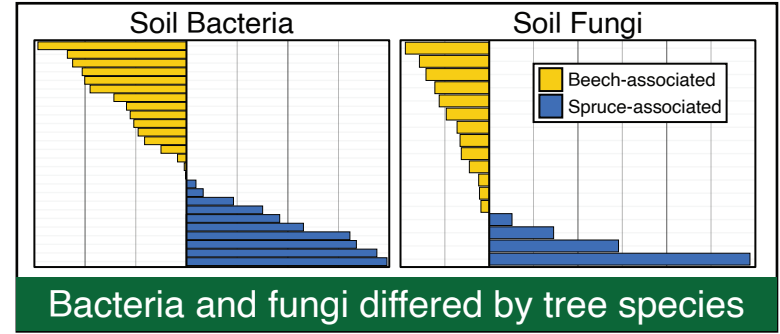
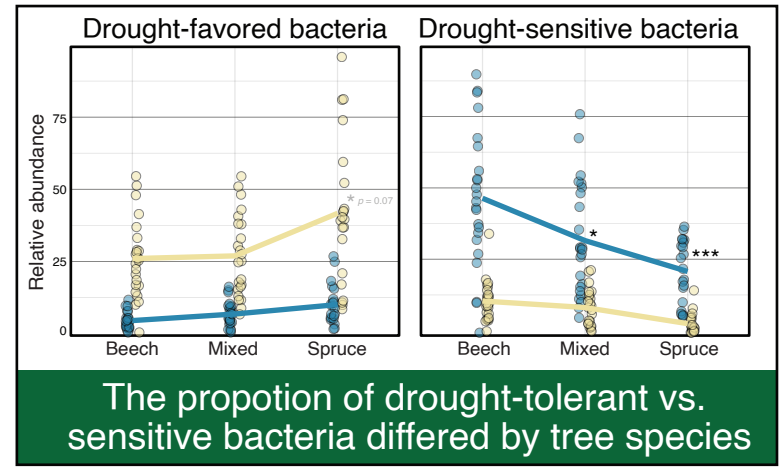
Research questions



Experimental design



Results



Highlights

- Long-term reduction in moisture had little impact on mycorrhizal population structure
- More drought-sensitive bacteria were present in beech relative to spruce root zones
- Mixed-species root zones were a hybrid of beech- and spruce-associated microbiomes
- Several bacterial populations were more resistant to drought in mixed root zones
- Complementation in the drought resistance of tree-associated microbiomes can occur

The effects of mixed-species root zones on the resistance of soil bacteria and fungi to long-term experimental and natural reductions in soil moisture

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Running title: Drought-microbiome effects in mixed rooting zones

Abstract

1 Mixed forest stands tend to be more resistant to drought than species-specific stands partially due
2 to complementarity in root ecology and physiology. We asked whether complementary differences
3 in the drought resistance of soil microbiomes might contribute to this phenomenon. We
4 experimented on the effects of reduced soil moisture on bacterial and fungal community
5 composition in species-specific (single species) and mixed-species root zones of Norway spruce
6 and European beech forests in a 5-year-old throughfall-exclusion experiment and across seasonal
7 (spring-summer-fall) and latitudinal moisture gradients. Bacteria were most responsive to changes
8 in soil moisture, especially members of Rhizobiales, while fungi were largely unaffected, including
9 ectomycorrhizal fungi (EMF). Community resistance was higher in spruce relative to beech root
10 zones, corresponding with the proportions of drought-favored (more in spruce) and drought-
11 sensitive bacterial taxa (more in beech). The spruce soil microbiome also exhibited greater
12 resistance to seasonal changes between spring (wettest) and fall (driest). Mixed-species root zones
13 contained a hybrid of beech- and spruce-associated microbiomes. Several bacterial populations
14 exhibited either enhanced resistance or greater susceptibility to drought in mixed root zones.
15 Overall, patterns in the relative abundances of soil bacteria closely tracked moisture in seasonal
16 and latitudinal precipitation gradients and were more predictive of soil water content than other
17 environmental variables. We conclude that complementary differences in the drought resistance
18 of soil microbiomes can occur and the likeliest form of complementarity in mixed-root zones
19 coincides with the enrichment of drought-tolerant bacteria associated with spruce and the
20 sustenance of EMF by beech.

Key words: plant–soil interactions, forest soil microbiome, drought resistance, precipitation gradient, beech-spruce forest, and root complementarity.

1. Introduction

21 Plant species diversity positively correlates with ecosystem productivity (Hooper and
22 Vitousek, 1997; Liang et al., 2016; Tilman, 2001) and with increased resistance to extremes in
23 water availability, at least in grasslands (Craven et al., 2016; Isbell et al., 2015). Forest ecosystems
24 are vulnerable to the increasing frequency, intensity, and duration of drought caused by changing
25 precipitation patterns (Dai, 2013; IPCC, 2018). However, it remains to be proven whether forests
26 with higher plant diversity or functional richness are more resistant to drought than single species
27 ('species-specific') plantations (García-Valdés et al., 2021). The general relationship between
28 plant species diversity and productivity is, at least, partially due to the effects of biotic feedbacks
29 between plants and soil microorganisms (Hendriks et al., 2013; Schnitzer et al., 2011). Yet, to date,
30 research into diversity-productivity relationships in forests has been primarily focused on
31 aboveground parameters, e.g. annual growth (Paquette and Messier, 2011; Pretzsch et al., 2020,
32 2010). More recently, belowground parameters were found to differ between species-specific and
33 mixed-species forest stands, including tree root lifespan dynamics (Zwetsloot et al., 2019) and root
34 niche partitioning during drought (Altinalmazis-Kondylis et al., 2021; Zwetsloot and Bauerle,
35 2021). These observations raise questions about the potential influence of other belowground
36 biotic interactions on the drought resistance of mixed forest stands.

37 Belowground interactions between root systems of European beech (henceforth 'beech')
38 and Norway spruce (henceforth 'spruce'), a common forest type in Central Europe, can differ
39 between monoculture and mixed-species forests (Schmid and Kazda, 2002). Spruce and beech
40 exhibit anatomical (tracheid vs. xylem vessel elements) and physiological (isohydric vs.
41 anisohydric) differences that effect their hydrological processes and drought response (Lyr et al.,
42 1992; Pretzsch et al., 2013) and their root systems can interact in complementary ways that reduce
43 competition and facilitate access to soil moisture. These effects are evident in the differences

44 between species-specific and mixed root zones, where beech roots exhibit lower mortality than in
45 monospecific root zones, due to decreased intraspecific competition (Zwetsloot et al., 2019), and
46 where spruce roots maintain higher fine root mass during chronic drought (Zwetsloot and Bauerle,
47 2021). This complementarity is due to several aspects of root ecophysiology, which include: (i)
48 differences in rooting depth, with beech colonizing lower soil layers in mixed root zones (Bolte
49 and Villanueva, 2006), and (ii) diverging responses in fine root growth during drought, where
50 spruce become more dormant (cease growth and enhance suberization), while beech maintain a
51 degree of root activity, producing new, albeit shorter lived, fine root mass (Nikolova et al., 2020;
52 Rötzer et al., 2017). Beech and spruce also select for broadly different soil microbiomes (Bárta et
53 al., 2017; Uroz et al., 2016). Complementarity in the function and drought-tolerance between their
54 associated microbiomes might contribute to the altered plant drought responses where roots mix.

55 Root-associated microbes support the nutritional needs of trees (Brzostek et al., 2013;
56 Calvaruso et al., 2014; Gan et al., 2021; Nicolitch et al., 2016), and provide protection from
57 phytopathogens (Lehr et al., 2008; Terhonen et al., 2016) and the effects of mild drought (Pena et
58 al., 2013; Shi et al., 2002). Prior research has shown that the soil microbiome in mixed-species
59 forests can have greater functional and metabolic diversity than species-specific forests (Beugnon
60 et al., 2021; Prada-Salcedo et al., 2021). Thus, complementarity in the function and drought-
61 tolerance of soil microbiomes associated with beech and spruce might confer benefits during
62 drought. Several observations indicate the potential unevenness in the susceptibility or tolerance
63 of members of the beech- and spruce-associated microbiomes. The spruce soil microbiome is
64 characterized by a higher relative abundance of Actinobacteria, Armatimonadetes, and
65 Acidobacteria (and a higher denitrification potential) and greater proportion of ectomycorrhizal
66 fungi (EMF) relative to saprotrophic fungi than the beech soil microbiome, which has higher

67 proportions of Alphaproteobacteria, Planctomycetes, and Verrucomicrobia and different dominant
68 EMF taxa (*Scleroderma*, *Russula* and *Laccaria*) than spruce (*Boletus* and *Thelephora*) (Bárta et
69 al., 2017; Felsmann et al., 2015; Nacke et al., 2016; Uroz et al., 2016). Populations of
70 Alphaproteobacteria and Planctomycetes (beech-associated) are generally more sensitive to
71 drought, while Armatimonadetes and Actinobacteria (spruce-associated) are more tolerant
72 (Bastida et al., 2017; Bouskill et al., 2013; Chodak et al., 2015; Curiel Yuste et al., 2014; Evans et
73 al., 2014; Felsmann et al., 2015; Hartmann et al., 2017; Preece et al., 2019; Zeng et al., 2016; Zhou
74 et al., 2018). EMF also differ in their drought tolerance with more sensitive (ex. within *Laccaria*
75 and *Lactarius*) and more tolerant (ex. within *Russula*, *Cenococcum*, and *Scleroderma*) species
76 among those that associate with beech and spruce (Coleman et al., 1989; Di Pietro et al., 2007;
77 Ortega et al., 2004). Disparities in the drought tolerance or susceptibility of beech- and spruce-
78 associated microbiomes have yet to be experimentally tested.

79 With our study, we investigated whether the effects of experimental and natural reductions
80 in soil moisture on the structure of beech and spruce microbiomes differed in species-specific
81 (single species) versus mixed-species root zones. Drought-affected bacteria and fungi were
82 identified in a 5-year-old throughfall exclusion experiment and changes in these populations were
83 profiled in seasonal and latitudinal gradients in soil moisture and precipitation, respectively (Figure
84 1). The composition of bacterial and fungal communities and root-associated ‘rhizobacteria’
85 (rhizoplane and root endophytes) were determined with 16S rRNA gene and ITS region amplicon
86 sequencing. We expected EMF and rhizobacterial populations to be most sensitive to reductions
87 in soil moisture given the associated reduction in fine root mass (Zwetsloot and Bauerle, 2021).
88 We hypothesized that the impact of reduced soil moisture would affect beech and spruce
89 microbiomes differently, and that these responses would be further altered in soils where roots

90 mixed. Specifically, our experiment tested the null hypothesis that the responses of beech- and
91 spruce-associated bacteria and fungi would not differ in mixed-species root zones relative to
92 species-specific zones. We then tested whether similar differences occurred in response to seasonal
93 variation and latitudinal differences in soil moisture. Our experiment advances an understanding
94 of the effects of drought and natural gradients in soil moisture on beech and spruce soil
95 microbiomes and establishes the potential for complementation to occur where root systems mix.

2. Methods

96 2.1 Study overview and experimental design

97 In 2013, a latitudinal precipitation gradient was established in Bavaria, Germany (Pretzsch
98 et al., 2014). The gradient is composed of five forest sites containing 60 to 90-year-old Norway
99 spruce (*Picea abies* [L.] Karst) and European beech (*Fagus sylvatica* L.) which are (from driest to
100 wettest): Arnstein, Kelheim, Kranzberg, Wasserburg, and Traunstein (Figure 1A). All five sites
101 are similar in mean annual temperature (MAT) and soil type (Cambisol and Luvisol), but differ in
102 their mean annual precipitation (MAP; Table S1). Relative differences in total annual precipitation
103 were consistent among sites during the duration of the experiment (Table S2). However, periods
104 of drought occurred throughout Germany in summer in 2015 and 2018 (Schuldt et al., 2020).
105 Individual sites consist of a species-specific spruce and beech root zone, as well as a mixed spruce
106 and beech root zone with varying degrees of mixing (Table S3). A throughfall-exclusion
107 experiment was established at the middle site of the precipitation gradient, (Kranzberg Forest Roof
108 Experiment or ‘KROOF’; Figure 1B) and is composed of six drought plots, each with a species-
109 specific beech and spruce and mixed-species stand covered by a throughfall exclusion roof, and
110 six corresponding uncovered reference plots, with corresponding stand compositions. At the time
111 of the last sampling in 2018, seasonal throughfall had been excluded for five years, resulting in an
112 overall reduction of ~70% soil moisture across time and significantly greater pre-dawn water stress

113 for beech ($\bar{x} = -0.66 \pm 0.2$ MPa water potential) and spruce ($\bar{x} = -0.98 \pm 0.3$ MPa) compared to
114 reference plots (Grams et al., 2021).

115 *2.2 Soil and root sampling*

116 Soils from each root zone were sampled during the fall of 2017 (November), and the spring
117 (April), summer (July), and fall of 2018 (October) at all five sites and from the throughfall
118 exclusion experiment. During each sampling, ten soil cores (1.6 cm in diameter and 30 cm long)
119 were harvested from five plots (approximately 10 m²) randomly located within in each tree root
120 zone and/or throughfall exclusion treatment plot. The O₁ (litter layer) was manually removed, and
121 each soil core was divided into upper (0-8 cm deep) and lower soil layers (9-30 cm) comprised of
122 the O_f + _hA_h and A₁B_v, respectively (KA5 classifications; Nickel et al., 2018). The upper and lower
123 layers of ten cores were separately composited to yield 5 replicates per tree root zone per site
124 and/or per throughfall exclusion treatment. Five grams of the upper and lower composite sample
125 (excluding roots and soil particles > 2 mm) was weighed into Whirl-pak® bags in the field (Nasco
126 Sampling, Madison, WI, USA), and transported on dry ice to the Technische Universität München
127 (TUM) campus in Freising, where they were dried at 60 °C to a constant weight. Oven drying was
128 used to minimize the period in which cells are active and has negligible impact on bacterial and
129 fungal community composition (Castaño et al., 2016; Tzeneva et al., 2009). Dried samples were
130 shipped to Cornell University (Ithaca, NY, USA) and stored at room temperature until DNA was
131 extracted within two months. Over short periods, changes in microbiome composition due to air
132 drying and storage are minor (Clark and Hirsch, 2008; Lauber et al., 2010; Tzeneva et al., 2009).
133 The remaining soil samples were stored at TUM at 4 °C for additional analyses and to determine
134 fine root biomass (mg) (≤ 1 mm diameter), also dried to 60 °C to a constant weight. Prior to the
135 onset of KROOF, in 2013, fine root samples were taken from upper and lower layers in species-
136 specific spruce and beech root zones for the purpose of identifying root-associated bacteria

137 ('rhizobacteria') associated with each tree ($n = 20$), as opposed to 'soil bacteria,' which we use to
138 refer to the amplicon data generated from whole soil cores. Roots were thoroughly cleaned in tap
139 water, identified under a stereomicroscope and frozen until DNA was extracted. The bacterial
140 populations identified via this method correspond to a combination of 'rhizoplane' and root
141 endophyte populations.

142 *2.3 Analyses of soil properties*

143 The remaining soil from each sample was used to measure soil water content (SWC), pH,
144 and soil organic matter (SOM). SWC was averaged from triplicate measurements using 3 g of
145 fresh soil, which was calculated as the percentage mass lost after drying for 24 hours at 105°C. Soil
146 pH was measured using a VWR Scientific Products model 2000 pH meter, in a 1:10 (w/v) dilution
147 with distilled deionized water, after the sample was shaken for one minute and allowed to settle
148 for 10 minutes. SOM was estimated using loss on ignition method, according to (Howard and
149 Howard, 1990), by weighing 2 g of soil onto clay trays that were put in a muffle furnace where
150 temperature was slowly increased by 10 °C increments until a final temperature of 550 °C was
151 reached and maintained for 3 h.

152 *2.4 Analyses of soil bacterial and fungal communities and rhizobacteria*

153 DNA was extracted from 0.25 g of soil sample using the Qiagen PowerSoil kits
154 (Germantown, Maryland, USA) according to the manufacturer's protocol with a bead beating
155 treatment of 3 min at 5.5 m·s⁻¹ (Bio-spec, Bartellsville, OK, USA). DNA quantification and PCR
156 amplification of bacterial (V4 region of 16S rRNA gene; 515f/806r) and fungal phylogenetic gene
157 markers (ITS1; nBITS2f/58A2r) was performed as previously described (Sridhar et al., 2022).
158 Duplicate PCR reactions per sample were pooled prior to Illumina MiSeq (2 x 250 bp; v2)
159 sequencing, which was performed at Cornell Biotechnology Resource Center (Ithaca, NY) using
160 dual-indexed bar-coded primers (Koechli et al., 2019; Kozich et al., 2013). Seventy-five ITS

161 sequencing libraries from the Kranzberg throughfall-exclusion experiment were discarded due to
162 poor sequencing quality. These were discarded at random without impacting the balance of our
163 experimental design (Table S4). To identify rhizobacteria, DNA was extracted from 0.35-0.45 g
164 of root using the PowerSoil kit as previously described (Nickel et al., 2018) and 16S rRNA
165 amplicon libraries were prepared targeting rhizobacteria using the same methods (details in SI)
166 except that the V3-V4 region was targeted using primers optimized for plant DNA-rich samples
167 (335f/769r), as previously described (Dorn-In et al., 2015). All sample metadata is provided in
168 Table S5, and raw sequencing data was archived with the European Nucleotide Archive under the
169 BioProject accession: PRJEB36981 (data reference, see Wilhelm et al., 2022).

170 Sequencing data was processed using QIIME2 (v. 2020.2; Bolyen et al., 2019) with a
171 dependency on DADA2 (v. 1.10; Callahan et al., 2016) to assign sequences to operational
172 taxonomic units (i.e., amplicon sequence variants). Taxonomic classification was performed using
173 the QIIME2 ‘q2-feature-classifier’ trained on the Silva database (v. 132; Quast et al., 2013) and
174 UNITE database (v. 7.2; Nilsson et al., 2019) for bacteria and fungi, respectively. OTUs found in
175 the non-template controls and in low abundance were removed, namely those present in fewer than
176 three samples, or at a total relative abundance < 0.01%. All counts were normalized by proportion
177 of total reads and presented as counts per thousand reads. R package *phyloseq* (v. 1.34; McMurdie
178 and Holmes, 2013) was used to characterize the communities and estimate diversity parameters on
179 rarified libraries ($n_{\text{bact}} = 15,300$ and $n_{\text{fungi}} = 2,544$ reads per sample). Fungal taxonomic
180 classifications were used to identify EMF, endophytic, and saprotrophic fungi using the FUNGuild
181 database (v. 1; Nguyen et al., 2016). Samples from throughfall-exclusion plots were removed
182 during all analyses of seasonal or latitudinal effects.

183 2.5 Bioinformatic and statistical analyses

184 The differential abundance of OTUs between reference and throughfall exclusion plots
185 (i.e., ‘drought-affected’) were identified using the *R* software package *indicspecies* (v. 1.7.9;
186 Cáceres and Legendre, 2009). Indicator species analysis was performed independently for soil
187 layers, season, and forest type, then combined and de-duplicated. Non-significant indicator OTUs
188 ($p_{\text{adj}} < 0.05$) and those with low indicator values (< 0.35) were excluded. Alpha-diversity was
189 measured as species richness, Shannon diversity, and Pielou’s evenness. Beta-diversity was
190 assessed using Bray-Curtis dissimilarity and differences in community composition were
191 visualized using principal coordinates analysis and tested with PERMANOVA ($n_{\text{perm}} = 999$) using
192 ‘adonis’ from the *R* package *vegan* (v. 2.5.7; Oksanen et al., 2015). Resistance (R) to change in
193 microbiome composition due to throughfall exclusion was measured as the Bray-Curtis
194 dissimilarity between reference and exclusion plots (D), such that $R = 1 - D$. A lower R value
195 indicates a greater dissimilarity between reference and throughfall exclusion plots, indicating a
196 greater drought effect (De Vries and Shade, 2013). R was calculated for the average of all
197 permutations of replicates between reference and exclusion plots. The environmental and
198 microbiome features that were most predictive of SWC were identified using random forest-based
199 feature selection implemented in the *R* package *Boruta* (v. 7.0; Kursu and Rudnicki, 2010).
200 Features included in model selection were environmental (soil layer, root zone, site, season, year)
201 and microbiome (aggregated counts at the taxonomic rank of Order scaled with the ‘scale’ function
202 in *R*). Subsequently, Pearson’s correlations between selected bacterial and fungal features and
203 SWC were performed using ‘rcorr’ from the *R* package *Hmisc* (v. 4.5; Harrell and Dupont, 2015).
204 Differences in the relative abundance of OTUs among tree root zones were tested with ANOVA
205 (‘aov’ function in *R*). The main effects and interactions between throughfall exclusion, season, or
206 tree root zone on the relative abundances of OTUs, were determined by fitting to fixed effects

207 linear models ('lm' function). P-values were adjusted for multiple test correction according to the
208 Benjamini and Hochberg false discovery rate. Only the most abundant OTUs (> 0.05% of sample
209 reads) were included in statistical testing. Significant effects are denoted by asterisk: $p < 0.05$ (*),
210 $p < 0.01$ (**), and $p < 0.001$ (***). All analyses can be reproduced with scripts included in the
211 Supplementary Data package available through the Open Science Foundation (doi:
212 10.17605/OSF.IO/DN9CH).

3. Results

213 Our initial analyses were performed to establish the drought susceptible and tolerant
214 (section 3.1) and spruce- and beech-associated soil microbiome and rhizobacteria (section 3.2)
215 prior to testing whether the response of these groups differed in species-specific versus mixed root
216 zones (section 3.3). Additional analyses were performed to characterize the natural variation
217 (seasonal and precipitation gradient) in drought-affected populations among tree root zones and
218 soil layers (section 3.4).

219 *3.1.1 General effects of throughfall exclusion on soil properties and the soil microbiome*

220 Across the full 5-year experimental period, throughfall exclusion resulted in a 70%
221 reduction in SWC, on average, during the growing season (Grams et al., 2021) and, at the time of
222 our sampling, the throughfall exclusion plots had significantly lower SWC in upper ($\bar{x} = -27.8\%$)
223 and lower soil layers ($\bar{x} = -22.4\%$) across all seasons (Figure S1). Upper layers had higher average
224 SWC though the extent of moisture reduction was comparable in both layers ($t = -3.9$ vs. -3.2 ,
225 respectively). Throughfall exclusion significantly decreased fine root mass, with a greater effect
226 on upper layers ($\bar{x} = -51.1\%$; $t = -4.2$; $p < 0.001$) compared to lower layer soils ($\bar{x} = -29.8\%$; $t = -$
227 3.2 ; $p < 0.001$). Exclusion had minor effects on soil pH (+1% in upper layer; $p = 0.03$); and had
228 mostly insignificant effects on SOM and DNA yield (Figure S1). Throughfall exclusion accounted
229 for a relatively low proportion of variation in the beta-diversity of soil bacterial and fungal

230 communities (Figure 2AB), which was primarily attributable to differences among soil layer and
231 tree root zone. Communities from mixed root zones exhibited an intermediate similarity between
232 species-specific root zones (Figure 2A).

233 3.1.2 Drought-affected soil bacteria and fungi

234 The differences in the relative abundance of OTUs between reference and throughfall
235 excluded plots were used to identify ‘drought-favored’ or ‘drought-sensitive’ bacterial and fungal
236 populations. The soil microbiome at Kranzberg was comprised of 10,357 bacterial and 3,282
237 fungal OTUs. Of these, a total of 369 bacterial and 23 fungal OTUs were indicative of throughfall
238 exclusion (Table S6). More OTUs were drought-sensitive ($n_{\text{bact}} = 194$ and $n_{\text{fungi}} = 15$) than drought-
239 favored ($n_{\text{bact}} = 175$ and $n_{\text{fungi}} = 8$). More drought-affected OTUs were observed in upper ($n = 215$)
240 versus lower soil layers ($n = 130$), but OTUs did not differ in their likelihood of being drought-
241 affected based on their soil layer association (Fisher’s test; $p = 0.8$).

242 Drought-favored fungi were primarily classified as Ascomycota (80%) and belonged to
243 genera designated as saprotrophic (*Pseudogymnoascus*, *Niesslia*, and *Ciliolarina*). Two of the
244 eight drought sensitive fungal OTUs were classified to genera of EMF (*Inocybe* and *Lactarius*).
245 However, the overall relative abundance of EMF was largely unaffected by throughfall exclusion
246 (Figure S2A). Fungal endophytes (genus *Phialocephala*) were significantly more abundant in
247 throughfall-excluded upper soils (Figure S3A) and were relatively more abundant in spruce root
248 zones (Figure S3B).

249 Throughfall exclusion affected a phylogenetically diverse group of bacteria, but most of
250 the affected OTUs were classified to Planctomycetes, Alphaproteobacteria and Actinobacteria
251 (Figure S4; complete list in Table S6). Actinobacteria were near uniformly drought-favored
252 (primarily classified as Pseudonocardiales, Solirubrobacterales, and Micrococcales), while
253 Planctomycetes were primarily drought-sensitive (Pirellulales, Gemmatales, and

254 Planctomycetales), though certain clades of Planctomycetes also contained a high proportion of
255 drought-favored taxa (Tepidisphaerales and Isosphaerales). Within Alphaproteobacteria, several
256 orders were favored by throughfall exclusion, namely populations of Caulobacterales
257 (*Phenylobacterium*), Elsterales, and Sphingomonadales (*Sphingomonas*) while others were
258 sensitive, namely populations of Rhizobiales (*Roseiarcus*, *Rhodoplanes*, and *Bradyrhizobium*) and
259 Reyranelles (*Reyranelia*). Other major bacterial groups that were drought sensitive included:
260 Deltaproteobacteria (Myxococcales), Verrucomicrobia (Pedosphaeraceae and
261 Xiphinematobacteriaceae), and Dependuntiae (Vermiphilaceae).

262 3.2.1 Beech- and spruce-associated soil bacteria and fungi and rhizobacteria

263 Soil microbiome composition significantly differed between beech and spruce root zones,
264 though the most abundant bacterial and fungal OTUs ($\geq 0.2\%$ of total reads) were present in all
265 root zones (97% of bacteria and 52% of fungi; Figure S5). The relative abundances of beech- and
266 spruce-associated OTUs in mixed zones were characteristically intermediate between the relative
267 abundance in the species-specific zones (Figure S6). The spruce-associated microbiome was
268 dominated by several orders of Actinobacteria (Frankiales and Solirubrobacterales) and
269 Acidobacteria (Acidobacteriales and Solibacterales; Figure 3A) and had a higher
270 Ascomycota:Basidiomycota ratio compared to beech (Figure S7). The beech microbiome had a
271 higher proportion of Proteobacteria (Alpha-, Gamma- and Delta-), Verrucomicrobia and
272 Planctomyces and a higher proportion of EMF (Figure S2B). Beech-associated taxa were
273 significantly more likely to be drought-sensitive than spruce-associated taxa (Fisher's Exact, O.R.
274 = 23.8; $p < 0.001$), with spruce-associated taxa tending to increase in relative abundance in
275 throughfall excluded plots.

276 Rhizobacterial populations associated with beech and spruce roots were identified using
277 indicator analysis. Beech-associated rhizobacterial populations were enriched in

278 Alphaproteobacteria from the family Xanthobacteraceae relative to spruce, which were enriched
279 in Acidobacteriales (*Acidipila*, *Granulicella* and *Occallatibacter*) and Frankiales (*Acidothermus*;
280 Figure 3B, complete list in Table S7). Bacterial species richness and evenness were significantly
281 higher in beech than spruce root zones, while no differences were evident for fungi (Figure S8).

282 3.2.2 Corresponding differences in plant and soil properties among root zones

283 Trends in the soil microbiome corresponded primarily with differences in fine root mass
284 and soil organic matter content in root zones. The upper soil layer in spruce root zones had a
285 significantly higher percentages of soil organic matter (2-fold) and SWC than beech, except for at
286 Arnstein, which received the lowest precipitation (Figure S9A). In contrast, the beech root zone
287 had significantly higher fine root mass in both the upper (66%) and lower layers (60%) at all sites,
288 while mixed zones had intermediate root mass. In all sites, fine root mass followed seasonal trends
289 in SWC between spring and fall, increasing in wetter sites (Wasserburg and Traunstein) and
290 declining in drier sites (Arnstein and Kranzberg), with the trend most pronounced in beech stands
291 (Figure S9B). Soil pH was significantly more acidic in spruce ($\bar{x} = 4.04$) than beech ($\bar{x} = 4.30$),
292 with intermediates values in mixed root zones ($\bar{x} = 4.21$).

293 3.3.1 Contrasting the effects of soil moisture on microbiomes among root zones

294 We tested for the generalized effects of soil moisture reduction on soil bacteria and fungi
295 among root zones using the community resistance (R) metric. In our study, lower R values indicate
296 a greater change in community composition resulting from reduced soil moisture (i.e., a lower
297 resistance). On average, R values for bacterial communities differed by tree root zone, but not for
298 fungal communities. Bacterial communities in the spruce root zones had higher R values than in
299 the beech root zones, though this difference was only significant in the lower soil layer (Figure
300 4A). Differences in R among root zones corresponded with the proportion of drought-favored
301 (higher in spruce) and drought-sensitive bacterial OTUs (higher in beech; Figure 4B). The spruce

302 soil microbiome also exhibited greater R values in relation to seasonal changes between spring
303 (wettest) and fall (driest) at Kranzberg and, in this case, the mixed root zone also exhibited
304 significantly higher R than species-specific beech (Figure S10A). These differences also
305 corresponded to significant seasonal increases in drought-favored and drought-sensitive taxa
306 (Figure S10B). R values for fungal communities were variable and did not significantly differ
307 among tree root zones (Figure S11A). Yet, the relative abundance of drought-favored fungal OTUs
308 was also significantly higher in spruce root zones (Figure S11B).

309 *3.3.2 Contrasting responses in the soil microbiome of species-specific and mixed root zones*

310 The effects of throughfall exclusion on several abundant beech- and spruce-associated
311 bacteria were significantly altered in the mixed-species root zone. Several beech- and spruce-
312 associated OTUs that were affected by throughfall exclusion in species-specific root zones
313 exhibited enhanced tolerance in mixed root zones (Figure 5AB; Table S8). Conversely, several
314 OTUs that were unaffected by throughfall exclusion in species-specific root zones exhibited
315 different responses in mixed-species root zones. These could be divided into beech-associated
316 OTUs, that exhibited a greater susceptibility (Figure 5C), and spruce-associated OTUs, that
317 exhibited a greater resistance, in mixed-species root zones (Figure 5D; Table S9). These trends
318 were evident when profiling the same populations across seasonal differences in SWC at
319 Kranzberg (lower panels in Figure 5C and 5D), and at other sites, though not uniformly (Figure
320 S12). There were no significant interactions between root mixing and throughfall exclusion for
321 soil fungi.

322 *3.4.1 Seasonal and latitudinal trends in drought-affected populations*

323 Complementary to the throughfall exclusion experiment, we profiled changes in the soil
324 microbiome of species-specific and mixed-species root zones across seasonal and latitudinal
325 gradients in soil moisture and precipitation, respectively (Figure 1A). A random forest, decision-

326 tree-based feature selection was used to identify the environmental variables and microbial taxa
327 that were most predictive of soil water content in these gradient (Table 1). Several bacterial orders
328 were selected as predictors of SWC, and the abundance patterns of Rhizobiales outranked all other
329 features in importance, including environmental parameters (Table 1). No fungal taxa were
330 selected as predictors of SWC. The abundance patterns of the top microbiome-based predictors of
331 SWC closely tracked seasonal and latitudinal differences in precipitation in upper (Figure 6) and
332 lower soil layers (Figure S13) and were among the same groups impacted by throughfall exclusion.
333 Among the OTUs identified as drought-affected in data from the throughfall exclusion experiment,
334 the number that significantly differed among high and low precipitation sites in the latitudinal
335 gradient was lowest in spring (n = 92), when SWC was highest, and progressively increased in
336 summer (n = 133) and fall (n = 220).

337 EMF did not exhibit a clear relationship to SWC as their relative abundance did not vary
338 significantly by season or site position in the precipitation gradient. The exception was at Arnstein
339 during the summer, when increased relative abundance of EMF coincided with some of the lowest
340 SWC values measured (Figure S2B). The EMF taxa that increased in Arnstein during the summer
341 included taxa from the genera: *Amphinema*, *Clavulina*, *Otidea*, *Sebacina*, *Tricholoma*, *Inocybe*,
342 and *Lactarius*.

343 *3.4.2 Moisture related shifts in rhizobacterial abundance between soil layers*

344 Reductions in soil moisture can cause stratification of roots in spruce, to upper soil layers,
345 and beech, to lower soil layers, according to observed root behavior (Bolte and Villanueva, 2006).
346 We found that the relative abundance of spruce- and beech-associated rhizobacteria tended to
347 differ between upper and lower soil layers correspondingly, particularly at Kranzberg (Figure 7).
348 The enrichment of spruce-associated rhizobacteria in upper layer soils was even more pronounced
349 during fall, when SWC was lowest, though this interaction was only significant at Kranzberg and

350 Wasserburg (Figure 7). In contrast, beech-associated rhizobacteria were more prevalent in the
351 lower soil layer, and no shift was seen due to throughfall exclusion or season. Additionally, at
352 Kranzberg, the seasonal shift to the upper soil layer was most pronounced in mixed root zones
353 ($t_{\text{interact.}} = 8.1$; $p < 0.001$) compared to species-specific spruce ($t = 3.9$) or beech root zones ($t =$
354 3.6). No shifts in spruce- and beech-associated fungi were evident, though we did not obtain fungal
355 sequence data from rhizoplane / endophyte samples.

4. Discussion

356 Our study reveals major differences in the response of Norway spruce and European beech
357 soil microbiomes to changes in soil moisture and their interplay in mixed root zones. A
358 significantly higher proportion of beech-associated taxa were sensitive to reductions in soil
359 moisture compared to the more drought-tolerant populations associated with spruce. Furthermore,
360 several populations of rhizobacteria and soil bacteria had increased resistance or susceptibility to
361 drought in mixed-species root zones relative to species-specific. These findings lead us to reject
362 the hypothesis that the response of microbial populations to changes in moisture would not differ
363 in mixed-species root zones, allowing for the possibility that complementation can occur. Here,
364 we discuss the possible reasons why drought may impact spruce- and beech-associated soil
365 microbiomes differently, and the potential consequences of the phenomena observed in mixed-
366 species root zones.

367 *4.1 Explaining the contrasting effects of reduced soil moisture on beech and spruce microbiomes*

368 Taxa that were consistently impacted by experimental or natural reductions in soil moisture
369 were broadly characteristic of the spruce and beech microbiomes described in prior research
370 (Asplund et al., 2019; Bárta et al., 2017; Felsmann et al., 2015; Nacke et al., 2016; Uroz et al.,
371 2016). Several of these broad taxonomic groups were previously shown to be impacted by reduced
372 throughfall in beech and spruce forest (Felsmann et al., 2015), and other forest types (Bastida et

373 al., 2017; Bouskill et al., 2013). Our results demonstrate that soil bacterial communities in species-
374 specific spruce root zones have a higher resistance to the impacts of drought and to seasonal
375 reductions in soil moisture than in beech root zones (Figure 4; Figure S10). This difference
376 corresponded with a higher proportion of drought-tolerant bacteria in the spruce microbiome than
377 in beech, and higher proportions of drought-sensitive bacteria in the beech microbiome. While this
378 is the first time such broad differences in moisture sensitivity have been reported, prior research
379 has shown that the mineral weathering activity of rhizobacteria isolated from beech was higher
380 during wetter periods while the reverse was true for those isolate from spruce (Collignon et al.,
381 2011).

382 Prevailing differences in soil conditions between spruce and beech stands likely help explain
383 the disparity in drought tolerance and sensitivity in their respective soil microbiomes.
384 Physiological differences in rooting depth result in consistently drier conditions in shallow soils in
385 spruce stands relative to beach (Allen et al., 2019; Zwetsloot and Bauerle, 2021). Our findings
386 indicate that this effect is large enough to select for higher proportion of drought stress-tolerant
387 bacteria in spruce soils. Additionally, soil acidification was more pronounced in spruce root zones
388 relative to beech (Figure S1), as indicated by the high relative abundance of acidophilic indicator
389 taxa (Acetobacterales, Frankiales and Acidobacteriales), as previously shown (Sridhar et al.,
390 2022a). Our observations were consistent with current understanding about difference in soil
391 development under beech and spruce stands caused by root and litter chemistry, nutrient leaching
392 and uptake, and mineralization and nitrification rates (Cremer and Prietzel, 2017). Spruce roots
393 and litter contain significantly higher concentrations of polyphenols than beech (Kuiters and
394 Denneman, 1987; Zwetsloot and Bauerle, 2018), which are generally toxic to soil heterotrophs
395 (Adamczyk et al., 2013; Chunmei et al., 2010; Inderjit et al., 2009; Metsämuuronen and Sirén,

396 2019). The high acidity and polyphenol content of spruce soils create adverse growing conditions,
397 which retard decomposition (Albers et al., 2004; Berger et al., 2004) and reduce soil respiration
398 and biomass in spruce stands relative to mixed-species and beech stands (Borken et al., 2002;
399 Borken and Beese, 2005; Lu and Scheu, 2021). These adverse conditions, along with consistently
400 lower shallow soil moisture levels in spruce stands, may select for stress-tolerant populations,
401 which may better endure water stress. This theory is supported by the overlapping physiological
402 stress response to acidity, low osmolarity, and desiccation observed in model bacteria (Ait-
403 Ouazzou et al., 2012; Hengge-Aronis, 2002; Spector and Kenyon, 2012). It is also anecdotally
404 supported by the enrichment of the thermophilic, acid-tolerant genus *Acidothermus* by throughfall
405 exclusion in spruce and mixed root zones (Figure S14), which are characteristic of droughted and
406 arid soils (Eppard et al., 1996; Lacerda-Júnior et al., 2019).

407 The prevalence of drought-sensitive taxa in beech-specific soils could also reflect differences
408 in the degree of rhizosphere activity between beech and spruce. Beech produce more fine root
409 mass (Finér et al., 2007), apparent at all our sites (\bar{x} = 1.4 to 2.1-fold higher), and sustain more
410 microbial biomass and higher soil respiration rates than spruce (Borken et al., 2002; Borken and
411 Beese, 2005; Lu and Scheu, 2021). Thus, the apparent sensitivity of beech-associated taxa may
412 reflect the inability of beech roots to sustain basal levels of microbial activity during drought. The
413 potential diminished influence of beech on the soil microbiome during drought was evident in the
414 decline of microbial populations indicative of higher trophic complexity, including members of
415 the Vermiphilaceae, endosymbionts of amoeba (Delafont et al., 2015), and Candidatus
416 *Xiphinematobacter*, symbionts of nematode (Rius et al., 2021; Vandekerckhove et al., 2000).
417 While we cannot disentangle the relative contributions of roots versus litter to differences in the
418 soil microbiomes between spruce and beech, root traits have a far greater importance in explaining

419 the physicochemical properties of species-specific and mixed forest soils than litter traits (Cesarz
420 et al., 2013; Gillespie et al., 2021).

421 4.2 Effects of root mixing on the impacts of reduced soil moisture

422 Most beech- and spruce-associated fungi or bacteria occurred at an intermediate relative
423 abundance in mixed root zones. This result indicates that the influence of each tree follows a
424 gradient and that taxa associated with either tree species can co-occur. These observations are
425 consistent with prior reports of intermediate heterotrophic activity (Borken et al., 2002; Borken
426 and Beese, 2005), microbial biomass (Lu and Scheu, 2021), litter decomposition rates (Albers et
427 al., 2004), and mineral weathering (Cremer and Prietzel, 2017) in mixed root zones. Thus, the
428 hybrid soil microbiome in mixed root soils reflects the combined, but weakened, influences of
429 each tree.

430 The resistance of several beech and spruce-associated bacteria to throughfall exclusion was
431 enhanced in mixed root zones (Figure 5AB). These populations included members of
432 Acidobacteria (*Bryobacter* and *Occallatibacter*, both spruce-associated rhizobacteria), which are
433 reputed for their production of extracellular polymeric substances, which confer stress tolerance
434 to cells and may influence soil moisture dynamics (Foesel et al., 2016; Kielak et al., 2016;
435 Kulichevskaya et al., 2010). In contrast, several members of Xanthobacteraceae (*Bradyrhizobium*
436 and *Rhodoplanes*) were more susceptible to throughfall exclusion in mixed root zones (Figure 5C).
437 Xanthobacteraceae are known to increase in abundance in proximity to beech trunks (Nacke et al.,
438 2016), suggesting that their capacity to resist drought may depend on the density of beech roots or
439 litter. Conversely, several spruce-associated taxa exhibited enhanced resistance to drought
440 exclusively in mixed roots zones (Figure 5D), suggesting some populations may benefit from the
441 reduced competition where drought-sensitive, beech-associated taxa are diminished. Our
442 explanations for trends in mixed root zones remain to be tested, but these observations illustrate

443 the range of interactions occurring where the influence of trees and, by extension, their
444 microbiomes mix.

445 *4.3 Effects of reduced soil moisture on root-associated taxa*

446 The rhizosphere activity of spruce and beech varies by season and moisture availability, with
447 the highest activity in spring (Calvaruso et al., 2014) and diminished photosynthate-derived
448 microbial rhizosphere biomass during drought (Ruehr et al., 2009). We hypothesized that
449 reductions in soil moisture would disproportionately impact root-associated taxa, like EMF, which
450 are important contributors to nutrient acquisition by spruce and beech (Brzostek et al., 2013;
451 Calvaruso et al., 2014; Collignon et al., 2011; Gan et al., 2021; Nicolitch et al., 2016). Contrary to
452 expectations, we did not observe any compositional changes in soil EMF populations from
453 throughfall exclusion, seasonal variation, or across the precipitation gradient. The lack of shift in
454 EMF populations is not without precedent in spruce and beech forests (Gorfer et al., 2021; Nickel
455 et al., 2018; Žifčáková et al., 2015). The apparent resiliency of EMF to changes in community
456 composition may reflect their direct connection to live roots and a privileged access to root
457 exudates relative to bacteria, consistent with the more severe impacts in bacterial populations we
458 observed and reported elsewhere (de Vries et al., 2018; Fuchslueger et al., 2014). Access to
459 exudates and shelter within the root might explain why endophyte populations (*Phialocephala*)
460 were among the few fungi favored by throughfall exclusion (Figure S3).

461 Rhizobacteria were among the most affected by throughfall exclusion and seasonal and
462 latitudinal variation in soil moisture. Members of the Rhizobiales and Burkholderiaceae were
463 greatly diminished by throughfall exclusion in both beech and spruce root zones (Figure 6; Figure
464 S15). These drought-sensitive populations were dominated by members of *Bradyrhizobium* and
465 *Rhodoplanes* (Xanthobacteraceae) and *Caballeronia* and *Paraburkholderia* (Burkholderiaceae)
466 which are consistently associated with the rhizosphere of European beech and Norway spruce

467 (Colin et al., 2017; Uroz et al., 2016) and are common root- and mycorrhizae-associated bacteria
468 isolated from forest soils (Burke et al., 2008; Izumi et al., 2007, 2006; Kataoka et al., 2008; Tanaka
469 and Nara, 2009; Uroz et al., 2012; Wilhelm et al., 2020). The apparent drought sensitivity of these
470 rhizobacteria has implications for soil nutrient cycling and plant nutrition, given their involvement
471 in priming decomposition (Wilhelm et al., 2021; Zwetsloot et al., 2020), phosphate solubilization
472 in the beech and spruce rhizosphere (Lepleux et al., 2012; Nicolitch et al., 2016), and endophytic
473 nitrogen fixation (Puri et al., 2020).

5. Conclusions

474 We conclude that any potential complementation of soil microbiome function during drought
475 is likely to confer greater benefits to beech than spruce, given the higher proportion of drought-
476 favored spruce-associated taxa and their sustained resistance in mixed root zones. The enhanced
477 drought tolerance of beech in mixed stands with spruce has been reported (Schäfer et al., 2017),
478 though the benefits of mixing are more commonly reported to favor Norway spruce (del Río et al.,
479 2014; Ding et al., 2017; Rötzer et al., 2017). Thus, it remains to be seen what functions the
480 complementation of soil microbiomes might have in the drought resistance of mixed beech-spruce
481 forests. The impact of reduced soil moisture was greater on rhizobacteria than EMF populations,
482 with a higher proportion of EMF occurring in the beech soil microbiome. Thus, the likeliest form
483 of complementarity in mixed-root zones might correspond with the enrichment of drought-tolerant
484 populations by spruce and the sustenance of EMF activity by beech.

485 Forests and their soil microbiome are complex adaptive systems in which legacy and context
486 shape biological responses to water limitation (Bouskill et al., 2013). Yet, the moisture sensitivity
487 of major drought-affected populations, such as members of the Rhizobiales, were consistent across
488 sites and season, providing evidence for the potential widespread occurrence of phenomena
489 reported here. Future research is needed to understand the ecological and environmental drivers of

490 differences across sites and to test the functional consequences of complementation between
491 microbiomes of beech and spruce, and other abundant tree species, in mixed root zones.

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Data Availability Statement

944 All sample metadata is provided in Table S5. All analyses can be reproduced with scripts included
945 in the Supplementary Data package available through the Open Science Foundation (doi:
946 10.17605/OSF.IO/DN9CH). Raw sequencing data was archived with the European Nucleotide
947 Archive under the BioProject accession: PRJEB36981.

Author Contributions

948 RCW performed the data analysis, research, and writing. JMU contributed to data analysis
949 and research. Field sampling was performed by JMU, TLB, MG, FW, and KP and sample
950 processing (soil parameters, root biomass) by JMU and FW. Amplicon sequencing libraries were
951 prepared by RCW, JMU, and FW. TLB, KP, and DHB guided all research efforts, including
952 analyses and writing. The authors declare we have no competing or conflicts of interest.

Tables

953 **Table 1.** The top environmental and soil microbiome predictors of soil water content were
954 identified, and ranked, using Boruta random forest-based feature selection. The relative abundance
955 pattern of aggregated counts of Rhizobiales was ranked as the best predictor of SWC, followed by
956 soil layer and precipitation gradient site. Fungal and bacterial count data, aggregated by Order,
957 were included in feature selection, but only bacteria had predictive value. Bacterial orders that had
958 significant Pearson's correlations with SWC content in both layers ($r > |0.2|$ and $p_{adj} < 0.05$) and
959 were among the top 20 most important features are displayed (full list in Table S10). The family
960 Gimesiaceae is historically known as 'Planctomycetaceae.'

Figures

961 **Figure 1.** A schematic overview of the experimental design used to examine the effects of natural
962 and experimental variation in soil moisture on the soil microbiome of species-specific and mixed-
963 species stands of European beech and Norway spruce. In (A), samples were collected in spring,
964 summer, and fall and at sites spread across a natural precipitation in Bavaria, Germany. In (B), at
965 Kranzberg, the mid-point of the gradient, samples were taken from a five-year old throughfall
966 exclusion experiment where rain-out roofs had been used to reduce soil moisture during the
967 growing season. Each symbol accurately reflects the distribution of trees species (full details in
968 Rötzer *et al.* 2017). In (C), at all sites, and in the throughfall exclusion plots at KROOF, samples
969 were taken from species-specific or mixed-species tree root zones. Ten soil cores (30 cm deep)
970 were randomly sampled from five to six plots for each tree root zone at each site and during every
971 season and subdivided into upper and lower soil layers before being composited.

972 **Figure 2.** The composition of soil bacterial and fungal communities at Kranzberg differed
973 primarily by soil layer and tree root zone, as evident in (A) the grouping of samples by principal
974 coordinates analysis and (B) the proportion of variance explained (R^2) in a PERMANOVA
975 analysis based on the Bray-Curtis dissimilarity in community composition.

976 **Figure 3.** Beech and spruce root zones had marked differences in the taxonomic composition of
977 soil bacteria and fungi and rhizobacteria. In (A), the bar plots provide a summary of the relative
978 proportions of indicator OTUs for beech- and spruce-associated bacteria ($n = 506$) and fungi ($n =$
979 63) according to the ratio of their aggregated relative abundance at rank Order. The subset of tree-
980 associated taxa affected by drought are labeled on the y-axis. In (B), the taxonomic profile or
981 rhizobacteria associated with either beech or spruce according to indicator analysis using 16S
982 rRNA gene amplicon data generated from root material.

983 **Figure 4.** The resistance (R) of soil bacterial communities to the effects of throughfall exclusion
984 in soil moisture was greatest in spruce root zones at Kranzberg. Differences in R among root zones
985 (B) corresponded with the relative abundance of drought-favored and drought-sensitive
986 populations. Pairwise differences in resistance among tree root zones sites were tested using
987 TukeyHSD ($p_{adj} < 0.05$). In (B), the effects of throughfall exclusion ($p_{excl.}$) were significant for all
988 groups. Any significant differences between tree root zone were denoted by bars with asterisk and
989 interactions between throughfall treatment and tree root zone were denoted by asterisk.

990 **Figure 5.** The responses of beech- and spruce-associated OTUs to throughfall exclusion were
991 significantly altered in soil microbiome of mixed root zones at Kranzberg. In (A), the relative
992 abundance of several drought-sensitive beech-associated OTUs ($n = 10$) were largely
993 undiminished in mixed root zones. In (B), several drought-sensitive spruce-associated OTUs ($n =$
994 5) had enhanced resistance to throughfall exclusion in mixed root zones. Several OTUs that were
995 not significantly affected by throughfall exclusion in species-specific root zones showed, in (C),
996 an increased susceptibility ($n = 8$) or, in (D), an increased resilience ($n = 4$) in mixed zones. These
997 trends were consistent with trends in relative abundances between spring (wettest) and fall (driest
998 season) at Kranzberg (lower panel of C and D). Individual OTU exhibited significant interactions
999 between throughfall exclusion and tree root zone (Table S8 and Table S9) but were displayed in
1000 aggregate. Any significant differences between tree root zone were denoted by bars with asterisk
1001 and interactions between throughfall treatment (or season) and tree root zone were denoted by
1002 asterisk.

1003 **Figure 6.** The relative abundance of four major bacterial orders followed changes in soil water
1004 content caused by throughfall exclusion plots, seasonal variation, and across the latitudinal
1005 precipitation gradient. In (A), the soil water content in the upper soil layer was reduced by
1006 throughfall exclusion (box plot), and across seasons (x-axis) and gradient sites (lines). In (B), the
1007 relative abundances of the four bacterial Orders identified in feature selection as predictive of SWC
1008 (Table 1). The same trends were evident in lower layer soils, though less pronounced (Figure S13).
1009 Significant differences between mean SWC or relative abundance between seasons were denoted
1010 by bars with asterisks. Significant interactions between season and site were denoted by colored
1011 asterisks. Lettering denotes significant differences among sites according to TukeyHSD ($p_{adj} <$
1012 0.05).

1013 **Figure 7.** Shifts in the relative abundance of rhizobacteria in response to (A) throughfall exclusion
1014 or (B) seasonal differences between spring and fall revealed the putative partitioning of roots
1015 between soil layers. Spruce-associated rhizobacteria (upper panel) tended to occur at higher
1016 proportions in the upper soil layer, while beech-associated rhizobacteria (lower panel) tended to
1017 be more abundant in the lower layer. Significant differences between soil layer means were
1018 denoted by bars with asterisk and interactions between throughfall treatment (or season) and soil
1019 layer were denoted by asterisk.

Feature	Rank	Imp.	Order	Family	Correlation with SWC	
					r_{upper}	r_{lower}
Environmental	2	21.7	Soil layer		-	-
	4	17.7	Site		-	-
	13	10.5	Season		-	-
	28	6.9	Tree root zone		-	-
	37	6.0	Year		-	-
Microbiome	1	24.2	Rhizobiales		0.39 ***	0.33 ***
				Beijerinckiaceae	0.38 ***	0.19 **
				Xanthobacteraceae	0.25 ***	0.31 ***
				KF-JG30-B3	-0.33 ***	-0.11 *
	6	15.7	Planctomycetales		0.35 ***	0.24 ***
				Gimesiaceae	0.49 ***	0.20 **
	9	11.2	Gaiellales		-0.23 ***	-0.23 ***
	11	10.6	Caulobacterales		-0.32 ***	-0.22 ***
				Caulobacteraceae	-0.33 ***	-0.24 ***
	15	9.6	Tepidisphaerales		-0.35 ***	-0.11 *
WD2101 soil group				-0.34 ***	-0.11 *	

Drought sensitive
 Drought favored

Table 1. The top environmental and microbiome predictors of soil water content were identified, and ranked, using Boruta random forest-based feature selection. The relative abundance pattern of aggregated counts of Rhizobiales was ranked as the best predictor of SWC, followed by soil layer and precipitation gradient site. Fungal and bacterial count data, aggregated by Order, were included in feature selection, but only bacteria had predictive value. Bacterial orders that had significant Pearson's correlations with SWC content in both layers ($r > |0.2|$ and $p_{\text{adj}} < 0.05$) and were among the top 20 most important features are displayed (full list in Table S8). The family Gimesiaceae is historically known as 'Planctomycetaceae.'

Figure 1.

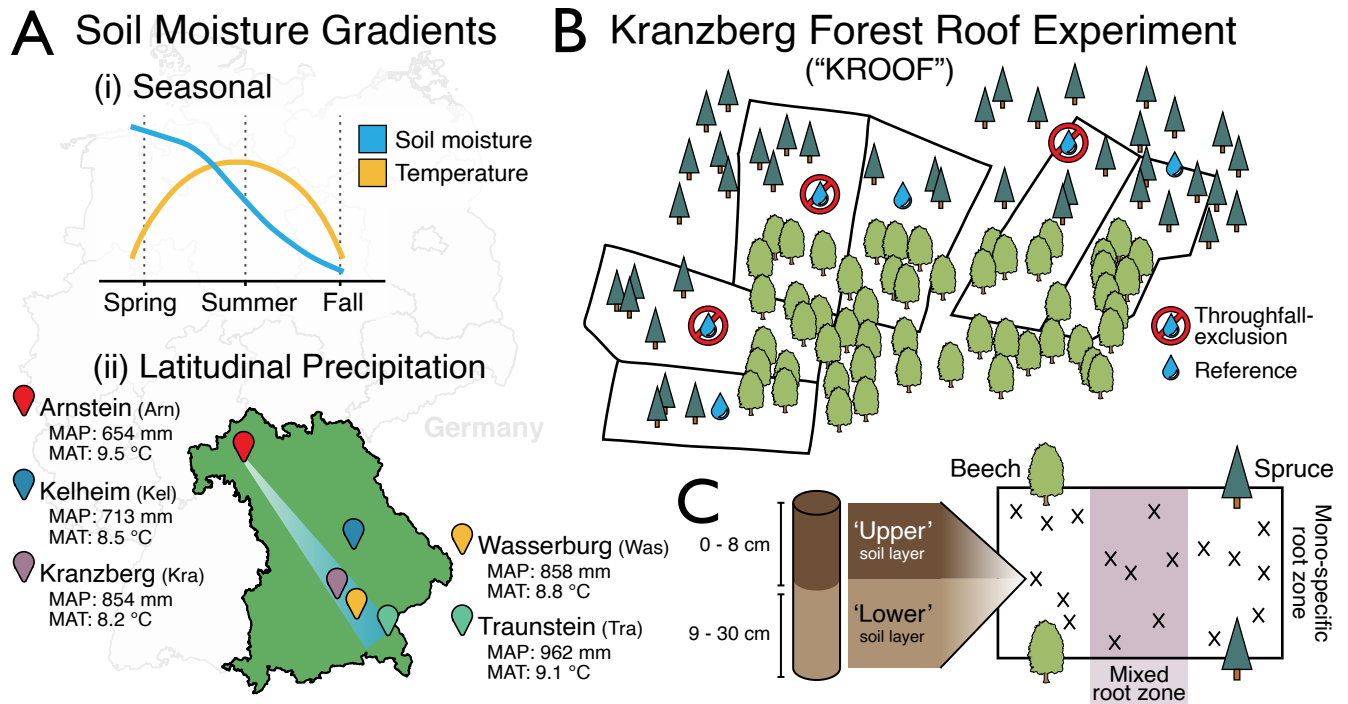


Figure 2.

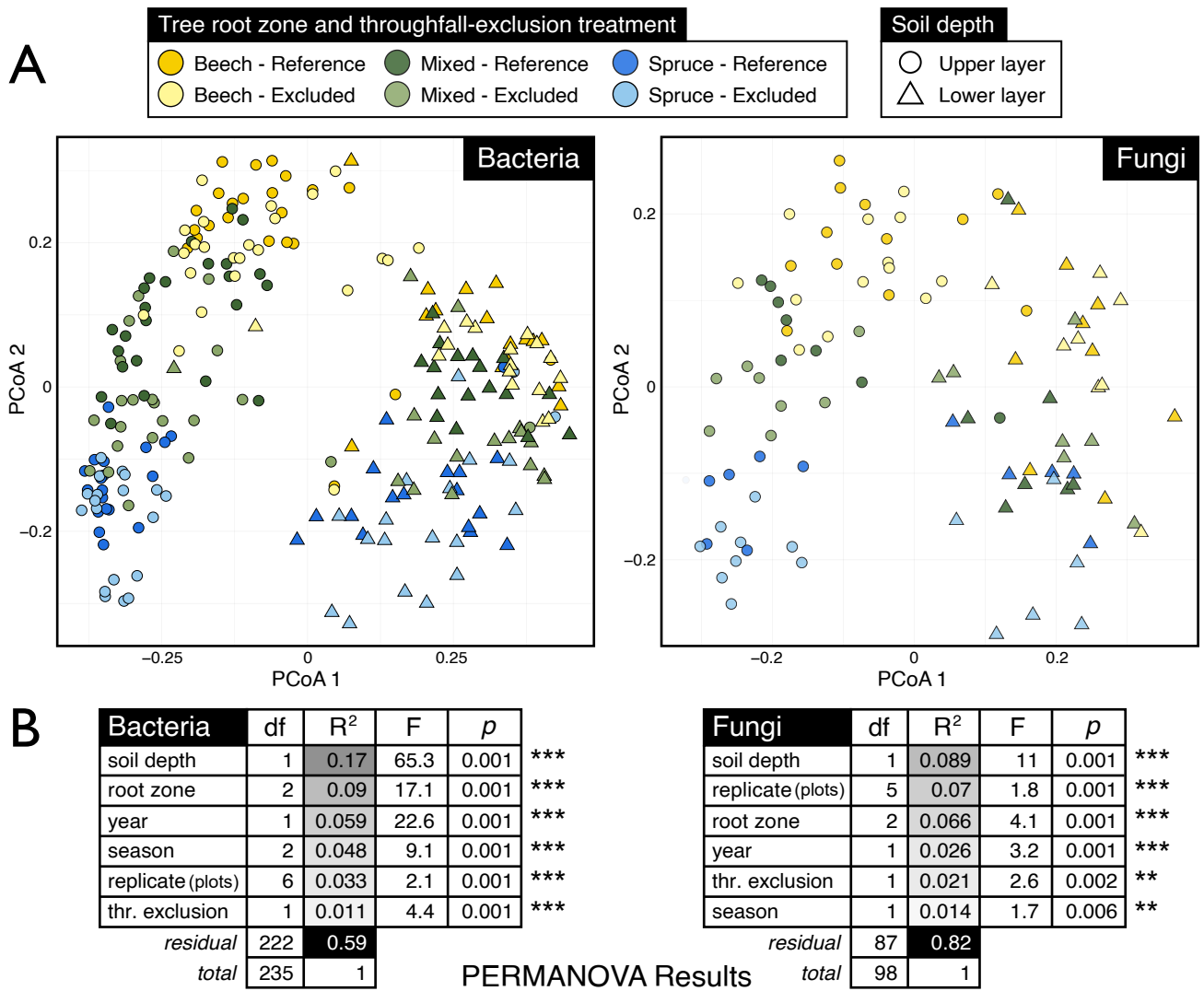


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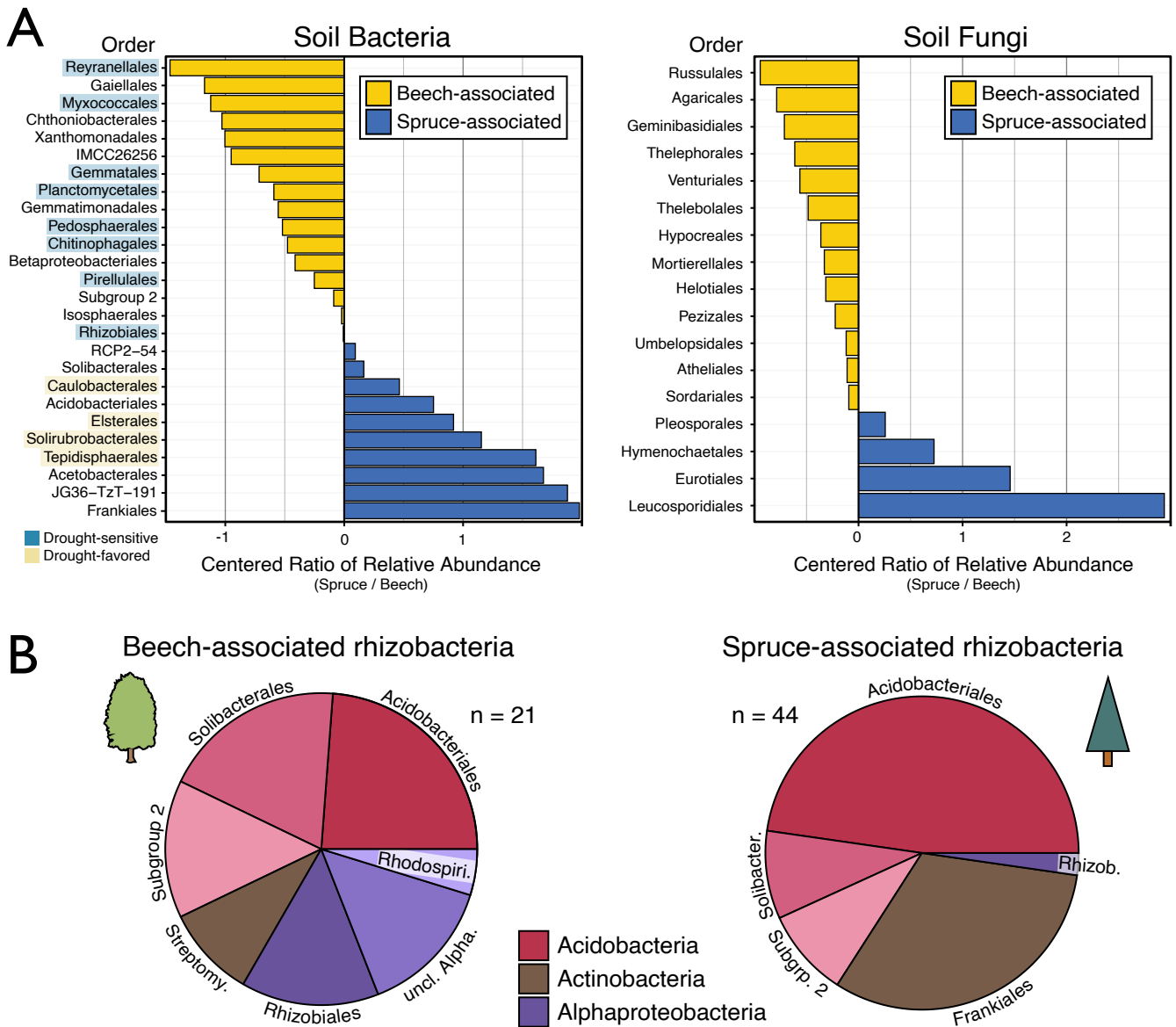
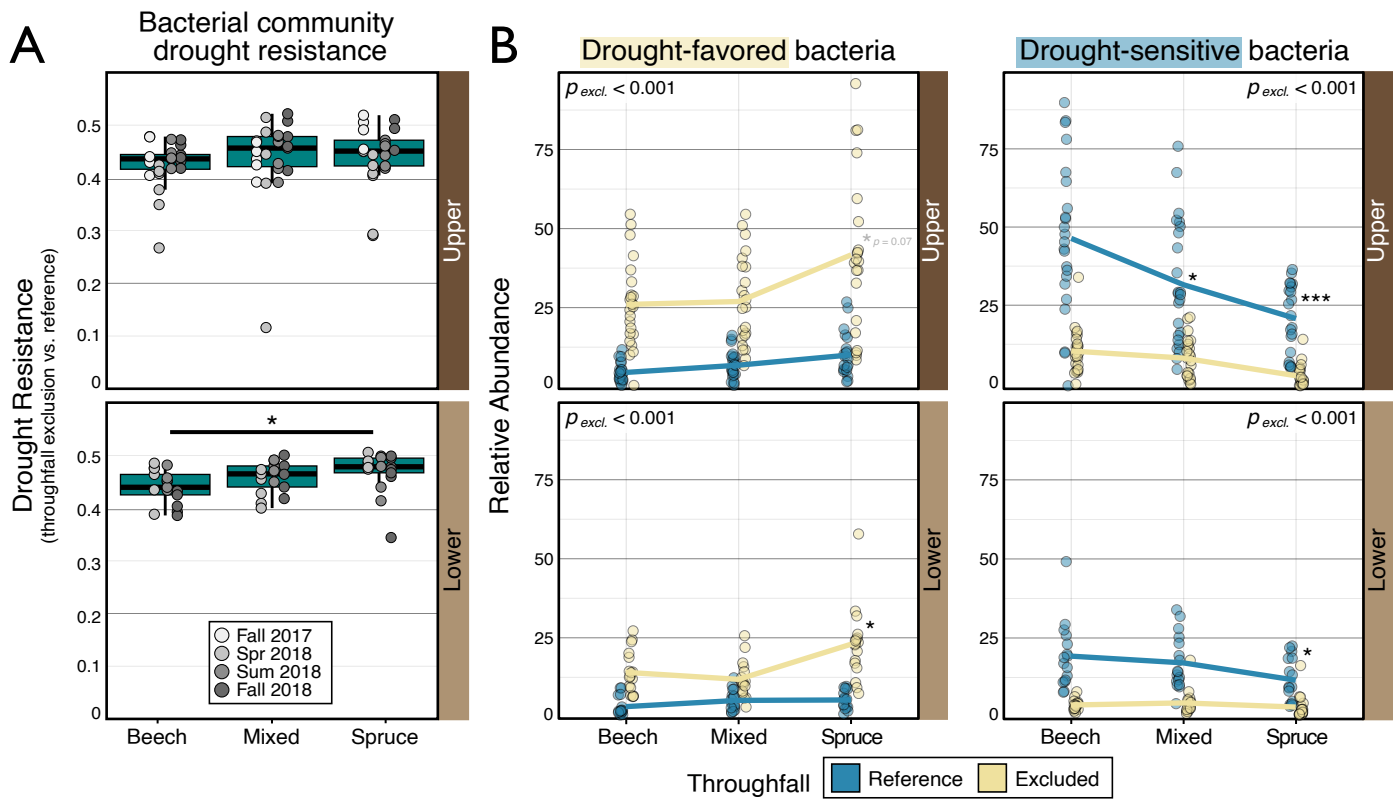


Figure 4.



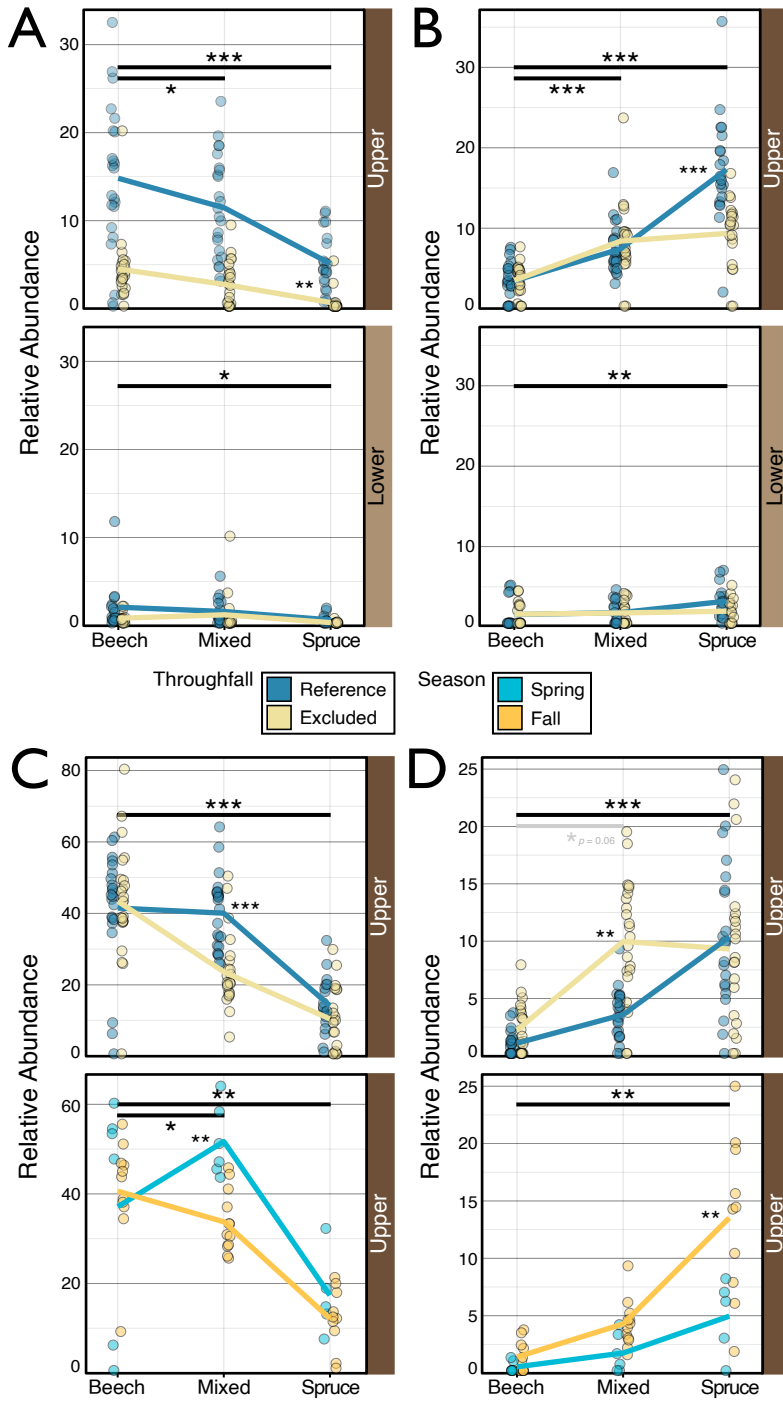


Figure 5.

Figure 6.

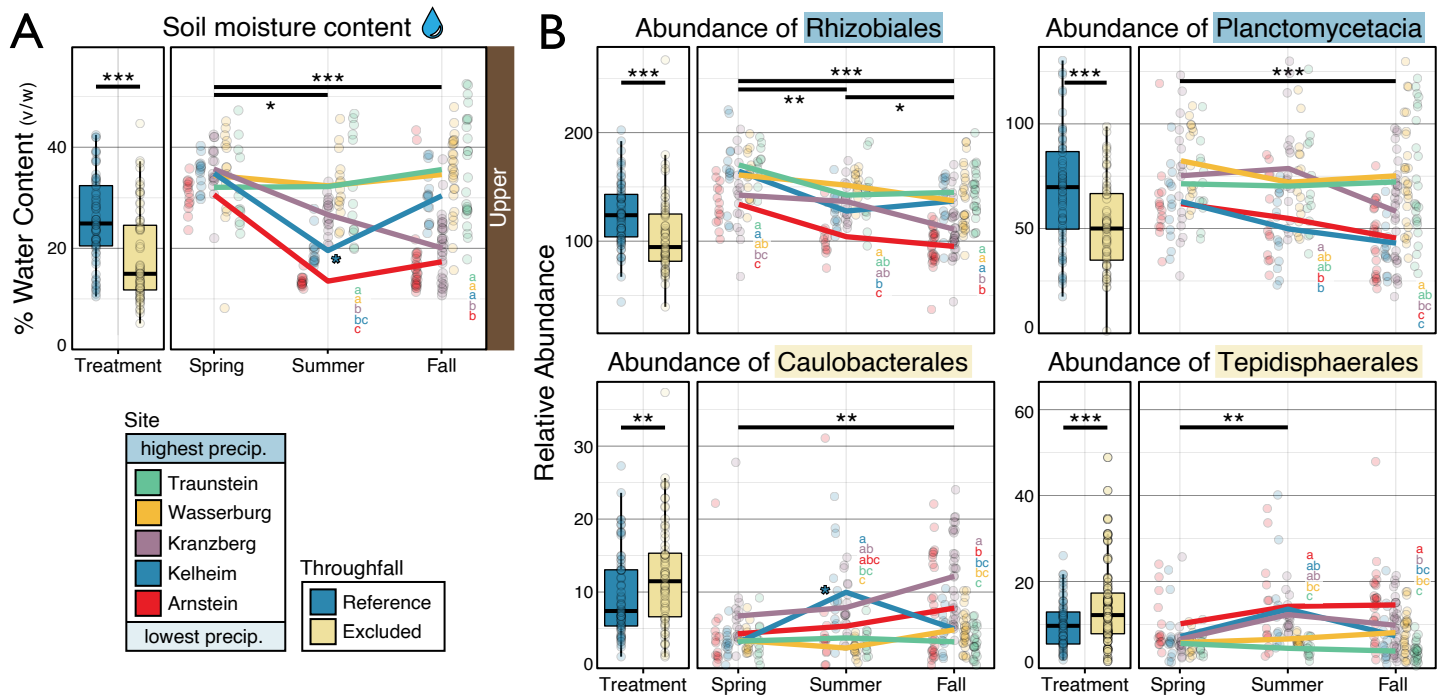
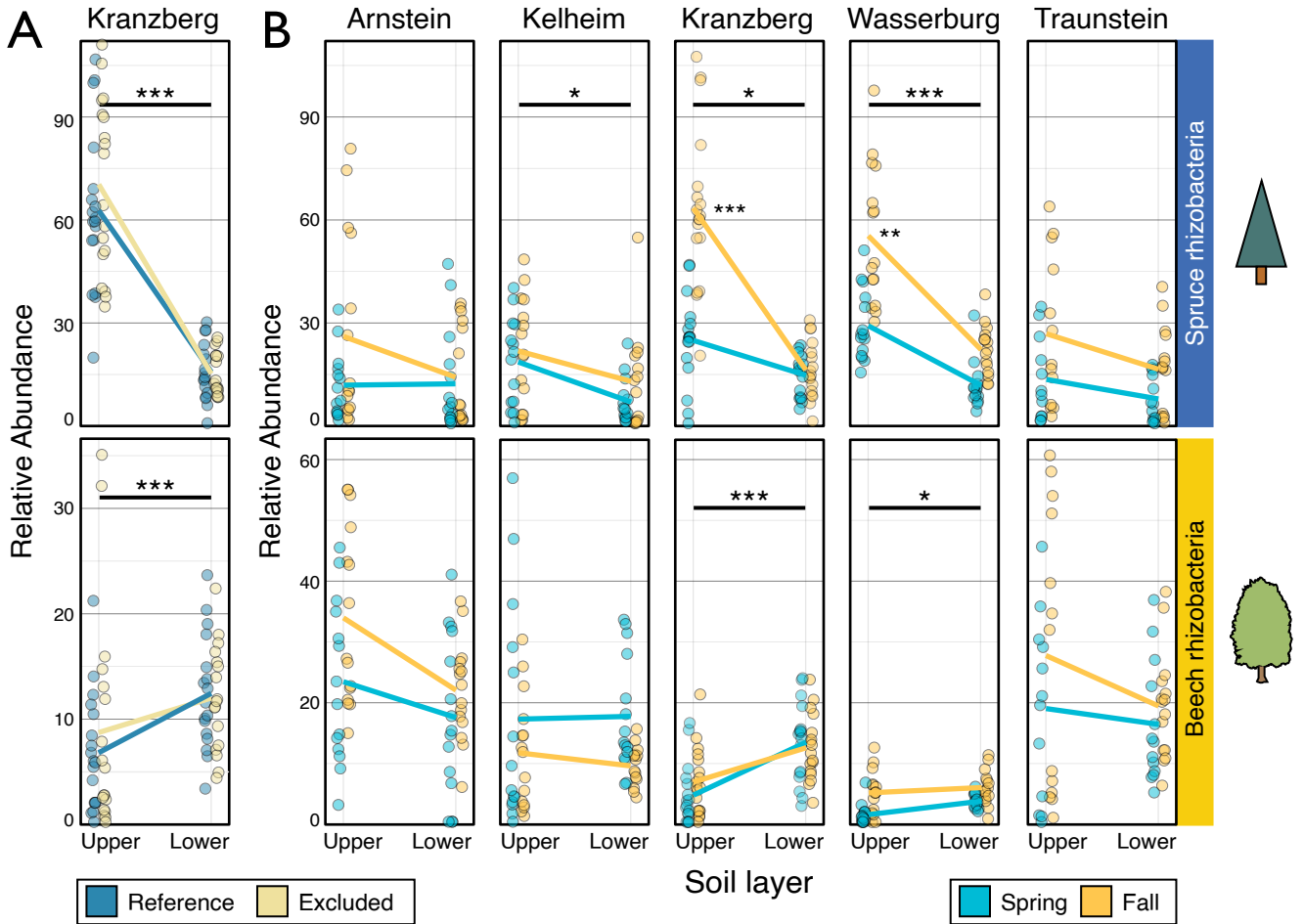
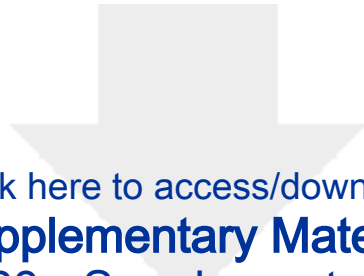


Figure 7.

Signatures of root partitioning in rhizobacteria between soil layers





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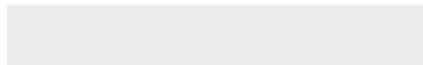




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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Author Contributions

RCW performed the data analysis, research, and writing. JMU contributed to data analysis and research. Field sampling was performed by JMU, TLB, MG, FW, and KP and sample processing (soil parameters, root biomass) by JMU and FW. Amplicon sequencing libraries were prepared by RCW, JMU, and FW. TLB, KP, and DHB guided all research efforts, including analyses and writing.