# **Unraveling the chemodiversity of halogenated disinfection by-products formed during drinking water treatment using target and non-target screening tools** 4 **C. POSTIGO**  $a,b^*$ , **A. ANDERSSON**<sup>c</sup>, **M. HARIR**<sup>d,e</sup>, **D. BASTVIKEN**<sup>c</sup>, **M.** 5 **GONSIOR**<sup>**f</sup>, <b>P.** SCHMITT-KOPPLIN<sup>c,d</sup>, **P.** GAGO-FERRERO<sup>§</sup>, L. AHRENS<sup>b</sup>,</sup> **L. AHRENS <sup>b</sup> , and K. WIBERG <sup>b</sup>** <sup>a</sup> Water, Environmental, and Food Chemistry Unit (ENFOCHEM), Department of Environmental Chemistry, Institute of Environmental Assessment and Water Research (IDAEA-CSIC), Jordi Girona 18-26, 08034 Barcelona, Spain. <sup>b</sup> Department of Aquatic Sciences and Assessment, Swedish University of Agricultural Sciences (SLU), Box 7050, SE-750 07 Uppsala, Sweden. 13 <sup>c</sup> Linköping University, Department of Thematic Studies-Environmental Change, 581 83 Linköping, Sweden. 15<sup>d</sup> Research Unit Analytical BioGeoChemistry, Department of Environmental Sciences, Helmholtz Zentrum Muenchen, Ingolstaedter Landstrasse 1, D-85764 Neuherberg, Germany. 17 <sup>e</sup> Chair Analyt Food Chem, Technical University Munich, Maximus von Imhof Forum 2, 85354 Freising Weihenstephan, Germany. 19 <sup>f</sup> Chesapeake Biological Laboratory, University of Maryland Center for Environmental Science, 20 Solomons, Maryland 20688, United States. 21 <sup>8</sup> Catalan Institute for Water Research (ICRA), Emili Grahit, 101, Edifici H2O, Parc Científic i Tecnològic de la Universitat de Girona, 17003 Girona, Spain. 24<br>25<br>26<br>27<br>28<br>29<br>30 \*Corresponding author: Cristina Postigo (0000-0002-7344-7044) [cprqam@cid.csic.es](mailto:cprqam@cid.csic.es) Institute of Environmental Assessment and Water Research (IDAEA-CSIC) Department of Environmental Chemistry C/ Jordi Girona 18-26, 08034 Barcelona, Spain. Tel: +34-934-006-100, Fax: +34-932-045-904

## **ABSTRACT**

 To date, there is no analytical approach available that allows the full identification and characterization of highly complex disinfection by-product (DBP) mixtures. This study aimed at investigating the chemodiversity of drinking water halogenated DBPs using diverse analytical tools: measurement of adsorbable organic halogen (AOX) and mass spectrometry (MS)-based target and non-target analytical workflows. Water was sampled before and after chemical disinfection (chlorine or chloramine) at four drinking water treatment plants in Sweden. The target analysis had the highest sensitivity, although it could only partially explain the AOX formed in the disinfected waters. Non- target Fourier transform ion cyclotron resonance (FT-ICR) MS analysis indicated that only up to 19 Cl and/or Br-CHO formulae were common to all disinfected waters. Unexpectedly, a high diversity of halogenated DBPs (presumed halogenated polyphenolic and highly unsaturated compounds) was found in chloraminated surface water, comparable to that found in chlorinated surface water. Overall, up to 86 DBPs (including isobaric species) were tentatively identified using liquid chromatography (LC)-Orbitrap MS. Although further work is needed to confirm their identity and assess their relevance in terms of toxicity, they can be used to design suspect lists to improve the characterization of disinfected water halogenated mixtures.

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 **KEYWORDS:** chlorination; chloramination; non-target analysis; identification workflow; haloacids

## **INTRODUCTION**

 Water disinfection is essential to protect public health from waterborne infectious diseases. Although disinfection can be achieved through physical and chemical methods, adding chemical disinfectants like chlorine or chloramine to the final product is common, as they are cheap, easy to handle, effective, and, most importantly, they provide residual disinfection that prevents pathogen regrowth in the distribution system network. Chemical disinfectants are strong oxidants that react with building blocks or alter the metabolism of pathogenic organisms, eventually killing them as the ultimate consequence [\[1\]](#page-99-0). However, oxidative reactions are not specific to the substrate and thus, all organic and inorganic constituents of water may be involved. As a result, a wide range of disinfection by-products (DBPs) will be unintentionally formed during the process [\[2\]](#page-99-1).

 The scientific community and the drinking water sector has been concerned about the formation of DBPs and their effects since the first discovery of DBPs in chlorinated water in two independent studies conducted in parallel in the mid-1970s [\[3,](#page-99-2) [4\]](#page-99-3). Research in this field has pointed out that many DBPs are highly cytotoxic and genotoxic to mammalian cells [\[5\]](#page-100-0), and a few of the tested DBPs even have all the toxicological characteristics to be classified as carcinogens to human (regulated trihalomethanes (THMs), formaldehyde, acetaldehyde, mutagen X (MX) and N-nitrosodimethylamine (NDMA)) [\[6\]](#page-100-1). Indeed, long-term exposure to THMs has been associated in epidemiological studies to an increased incidence of bladder cancer [\[7\]](#page-100-2) and may also have reproductive and developmental effects (mainly related to growth retardation) [\[7,](#page-100-2) [8\]](#page-100-3). *In vitro* toxicological studies have provided evidence on the different toxic potency of individual DBPs depending on their chemical structure. While nitrogen-containing DBPs are generally more toxic than only carbon-based DBPs [\[9\]](#page-100-4), and halophenolic

 DBPs are generally more toxic than haloaliphatic DBPs [\[10\]](#page-100-5), the toxicity of halogenated DBPs is also related to the halogen present in their structure, and increases in the order 84 chloro-DBPs<<<br/>shomo-DBPs<iodo-DBPs [\[9\]](#page-100-4).

 The chemical composition of the DBP mixtures and their formation from dissolved natural organic matter (NOM) are strongly dependent on the disinfectant used and the disinfection conditions (dose, contact time, water pH and temperature, etc.) and the source water characteristics (type and amount of NOM, inorganic ions such as bromide, iodide, ammonia, etc.) [\[11-14\]](#page-100-6).

 Due to the high chemodiversity of NOM (DBP precursors) and DBP mixtures, their comprehensive understanding and monitoring become a challenge [\[15,](#page-101-0) [16\]](#page-101-1). Furthermore, their full characterization is not possible with a single analytical technique. Pan et al. [\[17\]](#page-101-2) have recently reviewed the approaches used for NOM characterization in drinking water sources. As for DBP mixtures, only regulated DBPs are systematically monitored, and the analytical methods employed for their isolation and concentration and analysis are only capable of identifying and characterizing a specific fraction of the material formed during disinfection processes [\[18\]](#page-101-3). Most of the DBPs known to date (nearly 700 [\[19\]](#page-101-4)), especially those that are usually quantified in disinfected water, belong to the semi- to the highly-volatile fraction of the adsorbable organic halogenated material (AOX) formed during disinfection of water and are amenable to liquid-liquid extraction (LLE) and gas chromatography (GC)-mass spectrometry (MS) analysis [\[20\]](#page-102-0). However, the non-volatile fraction, for which major constituents and characteristics are largely unknown, may be toxicologically more relevant than the volatile portion [\[21\]](#page-102-1). Solid-phase extraction (SPE) approaches and liquid chromatography interfaced with MS (LC-MS) techniques are being applied during the last few years to characterize the unknown AOX fraction [\[19,](#page-101-4) [22\]](#page-102-2).

 Recently, high-resolution mass spectrometry (HRMS) techniques have become more popular using non-target workflows to unveil previously unknown DBPs [\[23-42\]](#page-102-3) and also to discover DBP precursors [\[43,](#page-105-0) [44\]](#page-105-1). However, the results of these studies are based on the use of one analytical technique only, i.e., GC-HRMS [\[23,](#page-102-3) [25\]](#page-102-4), Fourier transform ion cyclotron resonance (FT-ICR) MS [\[26-34\]](#page-102-5) or LC-HRMS [\[35-42\]](#page-104-0), and hence, the characterization of the DBP mixture is limited to one fraction. The majority of the LC-MS-based studies conducted to date in this field focus on discovering the DBPs generated by selected emerging organic contaminants during disinfection processes usually in pure water, using time-trends of features of interest [\[36-39\]](#page-104-1) or developing specific suspect lists [\[35\]](#page-104-0). There are only a few studies on purely LC-MS- based non-target workflows to unveil unknown DBPs in real mixtures [\[40-42\]](#page-104-2), and they focused on the identification of specific groups of compounds such as amino- compounds [\[41\]](#page-105-2), halogenated carboxylic acids [\[40\]](#page-104-2), peptide-DBPs [\[42\]](#page-105-3), or chlorine- [\[45\]](#page-105-4), bromine- [\[46\]](#page-105-5) or iodine-containing DBPs [\[47\]](#page-105-6) through the so-called product ion scan approach [\[22\]](#page-102-2).

 In this context, this study aimed at expanding the knowledge of real DBP mixtures produced by chlorine and chloramine-based disinfection processes at full-scale drinking water treatment plants (DWTPs) by applying different complementary analytical tools for DBP characterization. Target and non-target approaches were combined. GC-MS in combination with various extraction procedures offered a broad (~50) screening for known DBPs in the volatile fraction and HRMS tools, *viz*., FT-ICR MS and LC- Orbitrap MS, and non-target data treatment workflows unveiled the composition and chemodiversity of DBPs in the non-volatile fraction. Furthermore, the results from the aforementioned HRMS tools were compared with each other.

## **MATERIALS AND METHODS**

## *Chemicals*

 In total 47 DBPs were included in the target analysis including 8 THMs, 4 trihalogenated haloacetaldehydes (THALs), 8 haloacetonitriles (HANs), 13 haloacetamides (HACMs), and 14 haloacids (HAAs). The list has been provided in Table 1.

138 Ultrapure water (resistivity of 18.2 M $\Omega$ ·cm at 25 °C; TOC  $\leq$  5 ppb) used to prepare analytical methods blanks and to rinse sampling bottles and labware during the cleaning process was obtained using a Milli-Q Advantage system and aQ-POD dispenser equipped with a Millipack® Express 40 filter (Asymmetric PolyEtherSulfone (PES) 142 membrane, 0.22  $\mu$ m) for particles and bacteria removal, connected in series with a LC- Pack® Point-of-use Polisher cartridge (C18 reverse-phase silica) for trace organics removal (Merck Millipore).

 All reagents and solvents used were of high purity and mostly supplied either by VWR International (Spånga, Sweden) or Merck KgaA (Darmstadt, Germany).

 L(+)-ascorbic acid and sodium thiosulfate pentahydrate used to quench chlorine in water were Normapur® grade and supplied by VWR. Anhydrous sodium sulfate used to increase the ionic strength of the water to improve LLE efficiency and to dry the extracts was also Normapur® grade (VWR). ISOLUTE® Na2SO<sup>4</sup> drying cartridges used to dry extracts for HACMs and HAAs analysis were obtained from Biotage, Sweden.

 As for the acids used, ACS reagent grade formic acid (98-100%) (Emsure®), nitric acid 70%, and hydrochloric acid 30% (Suprapure®) were provided by Merck, whereas sulfuric acid 96% was supplied by VWR.

<b>DBP</b> class	<b>Analyte</b>	Acronym	<b>Molecular</b> formula	Mass CAS	(Da)* Number	<b>Supplier</b> (purity, %)
	Dibromochloromethane	<b>DBCM</b>	Br <sub>2</sub> ClCH	206	124-48-1	Sigma $(>99)$
	<b>Bromoform</b>	<b>TBM</b>	Br <sub>3</sub> CH	250	$75 - 25 - 2$	Sigma $(>99)$
	Dichloro-iodomethane	<b>DCIM</b>	Cl <sub>2</sub> ICH	210	594-04-7	CanSyn $(>95)$
Trihalo-	Chloro-bromo-iodomethane	<b>BCIM</b>	<b>BrCIICH</b>	254	34970-00-8	CanSyn $(>95)$
methanes	Dibromo-iodomethane	<b>DBIM</b>	Br <sub>2</sub> ICH	298	593-94-2	CanSyn (90-95)
(THMs)	Chloro-diiodomethane	<b>CDIM</b>	ClI <sub>2</sub> CH	302	638-73-3	CanSyn (90-95)
	Bromo-diiodomethane	<b>BDIM</b>	BrI <sub>2</sub> CH	346	557-95-9	CanSyn (90-95)
	Iodoform	<b>TIM</b>	$I_3CH$	394	$75 - 47 - 8$	Sigma (99)
Trihalo-	Chloral	<b>TCAL</b>	$Cl_3C_2HO$	146	$75 - 87 - 6$	Sigma $($ >98)
acetal-	Bromodichloroacetaldehyde	<b>BDCAL</b>	BrCl <sub>2</sub> C <sub>2</sub> HO	190	34619-29-9	CanSyn (90-95)
dehydes	Dibromochloroacetaldehyde	<b>DBCAL</b>	$Br_2ClC_2HO$	234	64316-11-6	CanSyn (90-95)
(THALs)	<b>Bromal</b>	<b>TBAL</b>	$Br_3C_2HO$	278	$115 - 17 - 3$	Sigma $(>97)$
	Chloroacetonitrile	CAN	$C_2H_2C1N$	75	$107 - 14 - 2$	Sigma $(>99)$
	Bromoacetonitrile	<b>BAN</b>	$C_2H_2BrN$	119	590-17-0	Sigma $($ >97)
	Iodoacetonitrile	<b>IAN</b>	$C_2H_2IN$	167	624-75-9	Sigma $($ >98)
Halo-aceto-	Dichloroacetonitrile	<b>DCAN</b>	$C_2HCl_2N$	109	3018-12-0	Sigma $($ >98)
nitriles	Dibromoacetonitrile	<b>DBAN</b>	$C_2HBr_2N$	197	3252-43-5	Sigma $(>90)$
(HANs)	Bromodichloroacetonitrile	<b>BDCAN</b>	$C_2BrCl_2N$	187	60523-73-1	CanSyn $($ >85 $)$
	Dibromochloroacetonitrile	<b>DBCAN</b>	$C_2Br_2CIN$	231	144772-39-4	CanSyn $($ >85 $)$
	Tribromoacetonitrile	<b>TBAN</b>	$C_2Br_3N$	275	75519-19-6	CanSyn (90-95)
	Chloroacetamide	<b>CACM</b>	ClC <sub>2</sub> H <sub>4</sub> ON	93	79-07-2	Sigma $(>98)$
	Bromoacetamide	<b>BACM</b>	BrC <sub>2</sub> H <sub>4</sub> ON	137	683-57-8	Sigma $($ >98)
	Iodoacetamide	<b>IACM</b>	IC <sub>2</sub> H <sub>4</sub> ON	185	144-48-9	Sigma $($ >98 $)$
	Bromochloroacetamide	<b>BCACM</b>	BrClC <sub>2</sub> H <sub>3</sub> ON	171	62872-24-8	CanSyn $(>99)$
	Dichloroacetamide	<b>DCACM</b>	$Cl2C2H3ON$	127	683-72-7	Sigma $(>99)$
Halo-	Dibromoacetamide	<b>DBACM</b>	$Br_2C_2H_3ON$	215	598-70-9	CanSyn $(>99)$
	acetamides Chloroiodoacetamide	<b>CIACM</b>	ClIC <sub>2</sub> H <sub>3</sub> ON	219	62872-35-9	CanSyn $(>99)$
(HACMs)	Bromoiodoacetamide	<b>BIACM</b>	BrIC <sub>2</sub> H <sub>3</sub> ON	263	62872-36-0	CanSyn $($ >85 $)$
	Diiodoacetamide	<b>DIACM</b>	$I_2C_2H_3ON$	311	5875-23-0	CanSyn $(>99)$
	Trichloroacetamide	<b>TCACM</b>	$Cl_3C_2H_2ON$	161	594-65-0	Sigma $(>99)$
	Bromodichloroacetamide		BDCACM $BrCl2C2H2ON$	205	98137-00-9	CanSyn $(>99)$
Haloacids (HAAs)	Dibromochloroacetamide		DBCACM ClBr <sub>2</sub> C <sub>2</sub> H <sub>2</sub> ON	249	855878-13-6	CanSyn $(>99)$
		<b>TBACM</b>	$Br_3C_2H_2ON$	293	594-47-8	
	Tribromoacetamide Chloroacetic acid	CAA		94	$79 - 11 - 8$	CanSyn $(>99)$
			$ClC2H3O2$			Sigma $(>99)$
	Bromoacetic acid	<b>BAA</b>	$BrC2H3O2$	138	79-08-3	Sigma $(>99)$
	Iodo acetic acid	IAA	$IC2H3O2$	186	64-69-7	Sigma $(98)$
	Chlorobromo acetic acid	<b>BCAA</b>	$BrClC2H2O2$	172	5589-96-8	Sigma $(>99)$
	Dichloroacetic acid	<b>DCAA</b>	$Cl2C2H2O2$	128	$79 - 53 - 6$	Sigma $(>99)$
	Dibromoacetic acid	<b>DBAA</b>	$Br_2C_2H_2O_2$	216	$631 - 64 - 1$	Sigma $(>99)$
	Chloroiodoacetic acid	<b>CIAA</b>	$CIC2H2O2$	220	53715-09-6	CanSyn $(>90)$
	Bromoiodoacetic acid	<b>BIAA</b>	$BrIC2H2O2$	264	71815-43-5	CanSyn $(>85)$
	Diiodoacetic acid	<b>DIAA</b>	$1_2C_2H_2O_2$	312	598-89-00	CanSyn $(>90)$
	Trichloroacetic acid	<b>TCAA</b>	$Cl_3C_2HO_2$	162	76-03-9	Sigma $(>99)$
	Bromodichloroacetic acid	<b>BDCAA</b>	$BrCl2C2HO2$	206	71133-14-7	Sigma $(>99)$
	Dibromochloroacetic acid	<b>DBCAA</b>	$Br_2ClC_2HO_2$	250	5278-95-5	Sigma $(>99)$
	Tribromoacetic acid	<b>TBAA</b>	$Br_3C_2HO_2$	294	75-96-7	Sigma $(>99)$
	Dalapon (2,2- dichloropropanoic acid)	<b>DCPA</b>	$Cl_2C_3H_4O_2$	142	75-99-0	Sigma $(>99)$

156 **Table 1.** Target DBPs, and corresponding acronyms, CAS numbers, purity and 157 provider of the analytical standard, molecular formula, and mass.

158 \*Nominal monoisotopic mass (Da).

 The solvents used for sample extraction and liquid chromatography analysis were: Ethyl 161 acetate (EtAc) for pesticide residue analysis, HPLC-grade water (Chromasolv<sup>TM</sup>), and HPLC-grade methanol (MeOH) (LiChrosolv®) and methyl *tert*-butyl ether (MTBE) (SupraSolv®) were provided by Merck.

 All reagents used in the production of diazomethane (derivatization agent) were supplied by Sigma Aldrich (Merck): diazald® (99%), Aldrich® diazomethane-generator 166 with System  $45<sup>TM</sup>$  compatible connection, diethylene glycol monoethyl ether (carbitol <sup>TM</sup>) (99%) and ACS-grade potassium hydroxide.

## *Sample collection*

 Water samples were collected at four different DWTPs in Sweden in October 2018. A volume of 24 L was grab sampled before (IN) and after (OUT) the final chemical disinfection process in each plant using stainless steel POP-cans (12 L, Sharpsville 173 container/NSF Component®). Additional water volumes were collected in 100 mL and 500 mL polyethylene (PE) bottles for general physical-chemical characterization and AOX measurements. To preserve AOX, sodium thiosulphate was added at a concentration of 5 mg/L and the water pH was lowered below 2 with concentrated nitric acid, following previous studies [\[48\]](#page-106-0) and EN ISO 9562:2004I recommendations [\[49\]](#page-106-1). After collection, samples were transported under cool conditions and stored at 4ºC in the dark until extraction, which took place the next day. Chlorine of water collected in POP-cans was not quenched to prevent potential interferences in the analysis or contamination. Furthermore, this allows mimicking the DBP mixtures to which consumers are exposed to since ~24 hours is the time that the finished water is in contact with the residual disinfectant before reaching the majority of the households in Sweden.

 Once in the lab, ascorbic acid (2.5 mg/L), (freshly prepared in Milli-Q-grade water) was used to remove free chlorine in sample aliquots used for target analysis, as it was reported to be the safest chlorine quenching agent for the analysis of the targeted DBPs 188 [\[50\]](#page-106-2).

 The selection of DWTPs was based on the type of the source water treated (i.e., surface water, or groundwater, and bromide content), and the chemical disinfectant applied (i.e., chlorine or chloramine). Thus, different DBP mixtures were expected to be formed. In all plants except in DWPT1, additional disinfection through UV radiation was conducted, in all cases before the chemical disinfection (Table2). However, the sample collection was designed and performed to examine only the effects of chemical disinfection. The investigated DWTPs have different treatment capacities, with daily 196 treated water volumes in the range of  $10,000 - 200,000$  m<sup>3</sup> (for details see Tables 2 and 3, and Figure 1).

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199 **Table 2.** Characteristics of the water in the four DWTPs sampled

<b>CODE</b>	<b>Type of</b> source water	<b>Disinfection</b> treatment <sup>a</sup> (mg/L) <sup>b</sup> (L/mg-M) <sup>b</sup> (°C) <sup>b</sup> PH <sup>b</sup>	<b>TOC</b>	<b>SUVA</b>			$Br^-$ $(mg/L)^b$	<b>Residual</b> total $Cl2$ (mg Cl <sub>2</sub> /L) <sup>c</sup>
DWTP1	Artificial groundwater (infiltrated river water)	<b>NaOCl</b>	3.7	1.998	10	8.3	0.11	0.50
	DWTP2 Surface water (lake)	$(UV + )$ NH <sub>2</sub> Cl	4.8	1.536	10	7.7	0.064	0.34
	DWTP3 Groundwater	$(UV + )$ NH <sub>2</sub> Cl	2.5	1.958	12	8.6	0.21	0.24
	DWTP4 Surface water (river)	$(UV + )$ NaOCl	4.0	1.399	11	8.6	0.052	0.13

200 **I**<sup>a</sup> IN samples were collected after UV disinfection and OUT samples after chemical disinfection; 201 b measured in the sample collected before disinfection; <sup>c</sup> measured in the sample collected after 202 disinfection



217 **Figure 1.** Scheme of the water treatment trains implemented in the DWTPs 218 investigated**.**







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<sup>224</sup> \*IN samples were collected after UV disinfection and before chemical disinfection; OUT samples were collected after chemical disinfection.<br>225 Iodide ( $\Gamma$ ), and phosphate ( $\text{P-PO}_4^{-3}$ ), chlorite ( $\text{ClO}_2^-$ ) and

225 Iodide (I<sup>-</sup>), and phosphate (P-PO<sub>4</sub><sup>-3</sup>), chlorite (ClO<sub>2</sub><sup>-</sup>) and chlorate (ClO<sub>3</sub><sup>-</sup>) were below the limit of quantification (<LOQ) in all samples (LOQ of I<sup>-</sup>: 0.025 226 mg/L, LOQ of P-PO<sub>4</sub><sup>-3</sup>: 0.003 mg/L, LOQ of ClO<sub>2</sub><sup>-</sup>: 0.005 mg/L and LOQ of ClO<sub>3</sub><sup>-</sup>: 0.011 mg/L.

*Sample extraction for target analysis*

 LLE was used for the target analysis of 47 DBPs. The LLE approaches used to extract targeted DBPs from water samples were based on the US Environmental Protection Agency (USEPA) method for the analysis of DBPs in drinking water (Hodgeson and Cohen 1990). All samples were extracted in duplicate.

 Ascorbic acid (2.5 mg/L) freshly prepared in Milli-Q-grade water was used to 234 quench residual free chlorine  $(< 0.5$  mg/L) in the samples and preserve the target DBPs.

 For extraction of THMs, THALs, and THANs, 100 mL of water was acidified to 236 pH<0.5 with 5 mL of concentrated  $H_2SO_4$  and then 30 g of dried granular  $Na_2SO_4$  was added to increase the ionic strength of the water and favor the partition of the analytes 238 into the extracting solvent (MTBE). After dissolution, the internal standard (IS) (100 µL 239 x 1 µg/mL of 1,2-dibromopropane (Sigma Aldrich) in MTBE) was added and mixed in the solution. Finally, the extracting solvent (2.5 mL of MTBE) was added. Samples were agitated with a mechanical shaker at 500 rpm for 30 min. After settling for 5 min, 242 the MTBE, laying on the top of the sample, was collected and dried using a  $Na<sub>2</sub>SO<sub>4</sub>$ column, and stored in a 2-mL vial at -20ºC in the dark until GC-MS analysis.

 To extract HAAs, a similar procedure was followed, using 50 mL and 245 proportional amounts of H<sub>2</sub>SO<sub>4</sub> (2.5 mL) and Na<sub>2</sub>SO<sub>4</sub> (15 g). After dissolution, the internal standard (IS) (100 µL x 1 µg/mL of 2.,3-dibromopropanoic acid (Sigma Aldrich) in MTBE). Then, 5 mL of the extracting solvent (MTBE) was added and the sample was vigorously manually shaken for 2 min. After settling for 5 min, the MTBE, laying on the top of the sample, was collected and transferred to 20 mL vial. The extraction step with 5 mL of MTBE was repeated twice, and finally, the combined MTBE extract was dried using ISOLUTE® Na2SO<sup>4</sup> drying cartridges (Biotage,

252 Sweden) and concentrated under  $N_2$  to a volume of 0.4 mL in a graduated test tube. The HAA extract (0.4 mL) was derivatized for one hour at room temperature with 0.2 mL of freshly prepared diazomethane. During the derivatization process, the methyl esters of HAAs were formed. These compounds are more volatile than HAAs and thus, amenable to GC-MS analysis. Diazomethane was produced in small (~3 mL) amounts from diazald using a diazomethane generator (Sigma Aldrich, Merck), following the manufacturer indications. After derivatization, the extract was transferred to a 2-mL vial 259 with 300  $\mu$ L insert for GC-MS analysis.

 The extraction of HACMs was conducted following the same steps as aforementioned for the extraction of HAAs. However, three main differences in the extraction protocol have to be highlighted: i) the water pH was lowered only to 5 with 263 diluted H<sub>2</sub>SO<sub>4</sub> to avoid HACMs degradation, ii) the use of <sup>13</sup>C-bromoacetamide dissolved in EtAc as IS, and iii) the use of EtAc as extracting solvent. The combined extract of EtAc obtained after three extraction cycles was dried using ISOLUTE® 266 Na<sub>2</sub>SO<sub>4</sub> drying cartridges and concentrated under N<sub>2</sub> to a volume of 0.2 mL in a graduated test tube. Finally, the concentrated extract was transferred to a 2-mL vial with 268 300 µL insert for GC-MS analysis.

## *Sample extraction for non- target analysis*

 For non-target analysis, the water samples were concentrated in triplicate using an automated SPE-DEX 4790 system (Horizon Technology Inc, Lake Forest, CA). The extraction approach used was based on previous works conducted for NOM characterization [\[26,](#page-102-5) [51\]](#page-106-3). Briefly, 5 L of water was acidified to pH 2.5 with 3 M hydrogen chloride (HCl) and then passed through an Atlantic hydrophilic-lipophilic balance (HLB)-H disk (Horizon Technology) previously conditioned with LC-grade

 methanol (MeOH) (2 x 30 s soak) and acidified LC-grade water (pH=2.5) (2 x 30 s soak). After sample load, the disk was washed using 0.1% formic acid aqueous solution 279 (2 x 10 s soak followed by 10 s  $N_2$  dry) to remove chloride and other ions that may potentially interfere in the FT-ICR MS analysis (e.g., adduct formation). The disk was 281 eluted with MeOH (2 x 90 s soak followed by 30 s N<sub>2</sub> flow). One-third (~10 mL) of the final extract (~160-fold concentrated water) was weighted and stored at -20ºC in the dark until FT-ICR MS analysis. This portion of the extract was further diluted with 284 MeOH to a DOC concentration of  $\sim$ 20  $\mu$ g/mL to prevent the negative effects of ion overload or space charging within the ICR cell. The other portion of the final extract (~20 mL) was reserved for LC-Orbitrap MS analysis. After evaporating its MeOH fraction, the aqueous extract was further processed using SPE with an Oasis HLB (200 mg) cartridge, using the same conditioning and elution solvents as for SPE-DEX extraction, to remove excess water (~36 %) and pre-concentrate to a final volume of 290 200 µL (ultimately a ~16,500-fold concentrated water sample). This extraction protocol did not allow capturing volatile-DBPs. Although a recovery study was not performed in this work, a previous study has reported a fairly good capacity of Oasis HLB sorbent to retain adsorbable organic chlorine and adsorbable organic bromine under acidic pH [21].

## *Physical-chemical characterization*

 Major ions were measured in all samples collected using ion chromatography coupled either to UV detection (230 nm, for iodide and iodate) or conductivity detection (for the remaining ions). Iodide and iodate were measured with an in house validated procedure, whereas the analysis of major ions was performed following ISO 10304-1:2007 [\[52\]](#page-106-4) and ISO 14911:1998 [\[53\]](#page-106-5).

 Total organic carbon (TOC) content was measured in triplicate in non-disinfected water 303 samples as the non-purgeable organic carbon (NPOC) fraction using a  $TOC-V<sub>CPH/CPN</sub>$  (Shimadzu, Japan) and the high-temperature combustion method (Standard method 5310B) [\[54\]](#page-106-6). Samples were first acidified to pH 2 with HCl to convert inorganic carbon 306 species (e.g., carbonates) to  $CO<sub>2</sub>$  that is removed by volatilization after sparging with synthetic air. Note that some volatile organic compounds are likely to be (partially) lost during this process. Finally, the sample was injected onto a heated column where 309 organic compounds are oxidized to  $CO<sub>2</sub>$  and the evolved  $CO<sub>2</sub>$  is measured with a non-dispersive infrared gas detector.

 Specific ultraviolet absorbance (SUVA) of non-disinfected waters was calculated after triplicate measurements of their UV absorbance at 254 nm with a UV-VIS spectrophotometer Lambda 365 (Perkin Elmer) following standard method 5910 [\[55\]](#page-106-7). Data acquisition was managed with the UV Winlab software 6.4.0.971 (Perkin Elmer).

 Temperature and pH of the water and residual free chlorine in disinfected water samples were obtained from on-line probes installed at the DWTPs.

## *Target analysis of halogenated DBPs*

## *GC-EI-MS analysis of THMs, THALs, HANs, and HACMs*

 The target analysis of the selected THMs (all except TCM and BDCM), THALs, HANs, and HACMs was conducted with GC-electron ionization (EI)-MS using a 7890B GC connected in series with a 5977A MSD (Agilent Technologies). One µL of the extract was injected in splitless mode using a 7693 automated autosampler equipped with a multimode inlet (split flow=50 mL/min, splitless time=1.5 min). The temperature of the injector was maintained at 200ºC for 0.1 s and rapidly increased to 300 ºC (600ºC/min).

 GC separation was achieved with a capillary GC column Rtx-200 MS (30 m x 0.25 mm x 0.25 µm) (Restek, Teknokroma, Barcelona), 1.2 mL/min of constant Helium flow, and a temperature gradient. For the analysis of THMs, THALs, and HANs the temperature 329 gradient started at 30°C (held for 5 min), and ramped at a rate of 9°C/min to 165 °C, and then at a rate of 20 ºC/min to 300ºC (held for 5 min). In the case of HACMs, the temperature gradient started at 50ºC (held for 3 min) and then, ramped at a rate of 9ºC/min to 165ºC and a rate of 25ºC/min to 285 ºC (held for 10 min). During both analytical runs, the temperatures of the GC-MS transfer line, and the MS source were 334 set to 280 °C and 200 °C, respectively.

 The analyzer was operated in selected ion monitoring (SIM) mode. A minimum of four ions was registered per analyte (see Table 4). Figures 2 and 3 show the total ion chromatogram obtained after analysis of calibration standard solutions at a concentration of 10 µg/mL. Mass acquisition and data analysis were performed using Mass Hunter B.07.00 software (Agilent Technologies).





Figure 2. Total ion chromatogram (TIC) obtained after GC-EI-MS analysis of MilliQ 343 water fortified with the target THMs, THALs, and HANs at a concentration of 10 µg/L.

- The THMs chloroform (TCM) and dichlorobromomethane (DCBM) were not
- captured with the analytical conditions used as they eluted in the solvent peak front and



therefore, they had to be excluded from the analysis.



 **Figure 3.** Total ion chromatogram (TIC) obtained after GC-EI-MS analysis of MilliQ 349 water fortified with the target HACMs at a concentration of 10 µg/L. 

## *GC-EI-MS/MS analysis of HAAs.*

 Analytical determination of methyl esters of HAAs was performed using a 7890B GC connected in series to a 7000C triple quadrupole (Agilent Technologies). Ionization was carried out in the electron ionization mode. One µL of the derivatized extract was injected in splitless mode using a7638B automated injector (split flow=50 mL/min, splitless time=1.5 min). GC separation of the analytes was achieved using a capillary GC column Rtx-200 MS (30 m x 0.25 mm x 0.25 µm) (Restek, Teknokroma, Barcelona), 1.2 mL/min constant flow of Helium and a temperature gradient (40 ºC held for 2 min, then increased at 10 ºC/min to 65 ºC and held for 2 min, and further increased at 10 ºC/min to 110ºC and at 20ºC/min to 285 ºC and held for 15 min. The temperatures

362 of the injector, the GC/MS transfer line, and the MS source were set to 250 °C, 280 °C, and 200 ºC, respectively.

 The analyzer was operated in selected reaction monitoring (SRM) mode, using nitrogen (1.5 mL/min) as the collision gas. A minimum of two SRM transitions was acquired per analyte (see Table 4). Figure 4 shows the total ion chromatogram obtained after the analysis of a standard calibration solution at a concentration of 10 µg/mL. Mass acquisition was performed using MSD ChemStation and data analysis was done with Mass Hunter B.08 (Agilent Technologies).



 **Figure 4.** Total ion chromatogram (TIC) obtained after GC-EI-MS/MS analysis of MilliQ water fortified with the target HAAs at a concentration of 10 µg/L.

## *Performance of target methods*

 The performance of the targeted methods was evaluated in terms of linearity, sensitivity, and accuracy (analyte recovery) and method repeatability. The results are summarized in Table 4. Quantification was performed by the internal standard method. For this, calibration curves were constructed by plotting the ratio of the analyte and the internal standard peak areas obtained in the different standard calibration solutions  (Milli Q water fortified at different concentration levels with the mixture of the target DBPs). A minimum of five calibration data points (exceptionally four in the case of few DBPs) in the range 0.1-50 µg/L in the case of THMs, THALs, and HANs, and 0.05-50 µg/L in the case of HACMs, and HAAs were used to construct these calibration curves. Quantitation of each analyte in the investigated samples was done according to the least-squares linear regression model obtained after the linear fitting of its calibration curve. The linearity range observed for each analyte and the coefficient of determination obtained for the corresponding model are summarized in Table 4.

 Method sensitivity was estimated from the analyte signal observed at the lowest calibration solutions. Method reporting limits corresponded with the analyte concentration that provided a signal-to-noise ratio of 10, and concentrations below the MRL with an S/N ratio of 3 were provided as detected but could not be quantified.

 Analyte absolute recoveries and method repeatability were evaluated through a recovery study. For this, LC-grade waters were fortified with the target DBPs at 0.5 µg/L (*n*=4) or higher concentration in the case of regulated THMs, DBCM, and TBM, and the trihalo-HANs BDCAN, DBCAN and TBAN, (1 µg/L, *n*=3) or those DBPs with MRL≥2.5 µg/L (DIACM, DBCACM, BDCAA, DBCAA, and TBAA) (5 µg/L, *n*=4) and extracted following the analytical protocols described. To calculate analyte absolute recoveries and repeatability the peak areas obtained in fortified water samples and standard solutions at equivalent concentrations were compared. The lowest recoveries were found for HACMs, which affects the sensitivity of the method for these compounds. However, analyte losses during the extraction were automatically corrected in the quantification process because calibration solutions were prepared by fortifying LC-grade water at different concentrations and processing these solutions as if they were samples.

405 **Table 4.** Retention time and ions/SRM transitions monitored for GC-MS analysis of 406 the target DBPs. The quantification ion/SRM transition is highlighted in bold.



<sup>4</sup>Average absolute recoveries observed at 0.5  $\mu$ g/L ( $n=4$ ) and relative standard deviation (RSD). In the case of regulated THMs, DBCM, and TBM, and the trihalo-HANs BDCAN, DBCAN, and TBAN, 408 case of regulated THMs, DBCM, and TBM, and the trihalo-HANs BDCAN, DBCAN, and TBAN, recoveries were investigated at 1  $\mu g/L$  ( $n=3$ ). For those analytes with MRL $\geq 2.5 \mu g/L$  (DIACM,

409 recoveries were investigated at 1 µg/L (*n*=3). For those analytes with MRL≥2.5 µg/L (DIACM,

410 DBCACM, BDCAA, DBCAA, and TBAA, average absolute recoveries were studied at a concentration<br>411 level of 5  $\mu g/L$  (*n*=4). NR: Analyte not properly recovered (RSD>100 and absolute recovery <30).

411 level of 5  $\mu$ g/L (*n*=4). NR: Analyte not properly recovered (RSD>100 and absolute recovery <30).<br>412 \*\* A minimum of 5 calibration points (exceptionally four) in the range 0.1-50  $\mu$ g/L in the case of TI

<sup>412</sup> \*\*A minimum of 5 calibration points (exceptionally four) in the range  $0.1-50 \mu g/L$  in the case of THMs,<br><sup>413</sup> THALs, and HANs, and  $0.05-50 \mu g/L$  in the case of HACMs, and HAAs were used to construct 413 THALs, and HANs, and 0.05-50 µg/L in the case of HACMs, and HAAs were used to construct

calibration curves.

## *AOX analysis*

 AOX was determined to assess the bulk of halogenated compounds present in the water. AOX analyses were conducted in all samples in triplicate, according to ISO standard 9562:2004 [\[49\]](#page-106-1). Briefly, 100 mL of water was transferred to an Erlenmeyer flask, 419 followed by pH adjustment to  $\neg$ pH 2 using concentrated HNO<sub>3</sub> and the addition of 5 mL 420 acidic nitrate solution (0.02 M HNO<sub>3</sub>, 0.2 M KNO<sub>3</sub>) and 50 mg ( $\pm$ 3 mg) activated carbon. The flask was shaken for 60 min at 180 rpm. The samples were then filtered to retain the activated carbon with the adsorbed organic compounds (polycarbonate material, 0.4 µm) (GE Healthcare Life Sciences, Uppsala, Sweden). Remaining halides were washed out from the filter using sequentially 2x10 mL of an acid nitrate solution 425 (1 mM HNO<sub>3</sub>, 10 mM KNO<sub>3</sub>) and 2x10 mL of acidified Milli-Q water (pH 2, after 426 addition of concentrated HNO<sub>3</sub>). The adsorbed organic compounds were combusted at 427 1000 °C in  $O_2$  atmosphere and the halides (HX) released in the process were determined by on-line microcoulometric titration (ECS 3000, Thermo Fisher Scientific).

## *Non-target FT-ICR MS analysis of halogenated DBPs*

 Non-target analysis of halogenated DBPs in SPE-DEX extracts was performed using a Bruker SolariX 12 Tesla FT-ICR MS and an APPOLO II ionization source, operating in negative electrospray ionization (ESI(-)) mode. The analysis was performed with a 434 spray current of -3.6 kV and a flow rate of 2  $\mu$ L min<sup>-1</sup>. A source heater temperature of 200°C was maintained to ensure rapid desolvation in the ionized droplets. The spectra were acquired with a time-domain of 4 megawords, and 300 scans were accumulated for each mass spectrum over the mass range *m/z* 147.4 to 1000. Injection lines were washed with a mixture of MeOH:water (8:2, v/v) between each sample, and MeOH solvents were run to control cross-contamination and carry-over.

 The non-target approach used is suitable to investigate non-volatile, medium to low polarity, and oxygen-containing compounds, e.g., molecules with carboxyl and/or hydroxyl moieties (amenable to ESI(-)).

 For data processing, unique molecular formulae were assigned to *m/z* ions present in the mass spectra using in-house software, developed at the Helmholtz Center for Environmental Health, Munich (Germany). Element constraints for the molecular 446 formulae assignments were <sup>12</sup>*C*: 0–100, <sup>1</sup>*H*: 0–∞, <sup>16</sup>*O*: 0–80, <sup>14</sup>*N*: 0–3, <sup>32</sup>*S*: 0–2, <sup>35</sup>*Cl*: 0– and <sup>79</sup>Br: 0–5. As a first data filter, only molecular formulas with a total ion count 448 (TIC) intensity >3,000,000,  $m/z \le 800$  Da, a mass error  $\le 0.2$  ppm, and in agreement with the nitrogen rule (i.e., N containing ions with even mass contain an odd number of N atoms) and containing Cl and Br atoms were further processed to identify and verify chlorinated and brominated DBPs, according to the approach followed in a previous study [\[26\]](#page-102-5). Iodine was not considered in formula assignment because an initial search using in-house developed software did not detect iodine-containing compounds.

 Furthermore, unrealistic formulae were also discarded according to their C, H and O 455 proportions, so that only those with C, O and H >0, O/C  $\leq$ 1, H/C $\leq$ 2.5, N and S $\leq$ 1 and double bond equivalents (DBE)≥0 were considered. Remaining formulae were verified as halogenated DBPs after evaluation of their predictable isotopic patterns, i.e., the presence of *m/z* ions expected to occur due to the different combinations of chlorine and 459 bromine stable isotopes  $({}^{35}Cl^{37}Cl$  and  ${}^{79}Br^{81}Br$ ). Verified formulae containing nitrogen or sulfur atoms were very few (Tables 7 and 8). Moreover, in the case of S-containing formulae, the majority was present in non-disinfected and disinfected waters at comparable intensity, and therefore they were excluded for data analysis.

 Only verified formulae with CHO and Br and/or Cl occurring in all three sample replicates of the water samples were further evaluated and interpreted. Finally, the

 formulae detected after disinfection while not being detected at the point before disinfection were considered as DBPs. Hence, presence and absence, rather than differences in relative intensities of individual formulae, was used to define the DBPs formed. Procedural blanks were used as quality controls because the very few peaks present in blank samples are usually not seen in real samples due to the suppression effects caused by the sample matrix components that compete for ionization.

 Visualization of non-target data was undertaken through three-dimensional van Krevelen diagrams (H/C vs O/C) that provide information about the degrees of saturation (y-axis) and oxygenation (x-axis) of the verified formulae [\[56\]](#page-107-0) and their mass distribution. Modified Kendrick mass defect (-KMD/z\*) plots were also created to show homologous series of molecules according to increasing number of methylene (-CH2) 476 units in the x-axis and the nominal exchange of  $CH_4$  against O along the y-axis and  $H_2$  along diagonals, since heteroatoms in the verified formulae are limited to oxygens [\[57\]](#page-107-1). 478 Diagrams showing DBE [\[58\]](#page-107-2), modified aromaticity index (AI<sub>mod</sub>) [58], and average 479 oxidation state of the carbon  $(C_{OS})$  [\[26\]](#page-102-5) against the number of carbons of the verified formulae were also constructed to evaluate and detect changes in DBP mixtures.

481 The differences in mass distribution, O/C, O/H, DBE,  $AI_{mod}$ ,  $C_{OS}$ , Cl/C and Br/C of the halogenated mixtures observed in each plant before and after disinfection (after removing overlapping features) were statistically evaluated using non-parametric tests (Mann-Withney U test) with a significance level of 0.05. To evaluate significant differences among all investigated disinfected waters, the Kruskal-Wallis test, and Dunn's pairwise *post hoc* tests were applied. Statistics were done using IBM SigmaPlot 12.5.

## *Non-target LC-ESI(-)-Orbitrap MS analysis of halogenated DBPs*

 The SPE-concentrated fraction was analyzed using an Acquity UPLC system (Waters) coupled to an Orbitrap mass spectrometer (QExactive, Thermo Scientific). Chromatographic separation was achieved with a Purospher® STAR RP-18 endcapped column (2 µm particle size, 150x2.1 mm) and a linear organic gradient of a mobile phase consisting of water and MeOH both with 0.1% formic acid at a constant flow rate of 0.2 mL/min.

 ESI was performed in the negative mode (ESI(-)) for the comparability of FT-ICR MS results. HRMS acquisition was conducted in data-dependent scan mode. This included a full scan over the *m/z* range 35- 650 at full width at half maximum (FWHM) resolution 500 of 70,000, and a data-dependent-MS<sup>2</sup> scan at a resolution of 35,000 on the top 10 ions 501 above an intensity threshold of  $1e^5$ .

 HRMS data were processed using the Compound Discoverer 3.1 software. Element 503 constraints for the molecular formulae assignments were <sup>12</sup>*C*: 0–90, <sup>1</sup>*H*: 190–∞, <sup>16</sup>*O*: 0– 504 15, <sup>14</sup>N: 0–10, <sup>32</sup>S: 0–5,<sup>31</sup>P: 0–3, <sup>23</sup>Na: 0-2, <sup>35</sup>Cl: 0–4, <sup>79</sup>Br: 0–4, <sup>127</sup>I: 0–3, and mass error 505 was set to  $\pm$  5 ppm. The number of oxygen atoms for elemental composition prediction was restricted to 15 according to the findings of FT-ICR MS data (halogenated formulae with a maximum of 12 oxygen atoms, Figure 11). Only features above a TIC intensity of 100,000 were considered. The thousands of peaks found were prioritized for further identification tasks according to their exclusive occurrence in all three replicates of disinfected water samples and absence in non-disinfected and blank samples and to the presence of halogens (i.e., Cl, Br or I) in their structure.

 Orbitrap MS has a lower FWHM mass resolution (70,000 at *m/z* 200) than FT-ICR MS 513 (400,000 at  $m/z$  400), which results in a higher mass error ( $\leq$ 5 ppm vs  $\leq$ 0.2 ppm). Such a mass error in Orbitrap MS generally leads to more than one logical elemental

 composition containing CHO, N, S, I, Br, and/or Cl. Therefore, the isotopic pattern of the parent ion was used to restrict the number of Br and Cl atoms in the elemental 517 composition, so that the isotopic cluster includes all ions with an  $m/z$  defect of  $\pm 1.997$ . Once the elemental composition was established, the MS2 fragmentation of the prioritized DBP was compared with *in silico* fragmentation of molecules with the same elemental composition contained in the PubChem database using MetFrag [\(https://msbi.ipb-halle.de/MetFrag/\)](https://msbi.ipb-halle.de/MetFrag/) [\[59\]](#page-107-3). The one with the highest score was provided as a tentative candidate, and its identity was only confirmed after the comparison of its retention time and MS2 fragmentation with an analytical standard (when available). The scoring terms selected were i) fragments match after *in silico* fragmentation and ii) spectral similarity of structure candidates (Figure 5). This workflow is illustrated in Figure 5, using halogenated derivatives of hydroxypiranones as an example. The main limitation of this approach is that structure candidates are limited to the database content.

## *Tranformation of DBP concentrations into Cl-equivalent concentrations*

 To convert DBP concentrations into Cl-eq concentrations, the following formula was applied, in which the same atomic weight (35.45 Da) is assigned to all halogens present in the molecule (chlorine, bromine, and iodine) [\[60\]](#page-107-4):

$$
\frac{\mu g \text{ of } DBP \text{ as } Cl - eq}{L} = \frac{DBP \text{ conc } (\frac{\mu g}{L})}{DBP \text{ M. W.} (\frac{g}{mol})} * (No. halogen atoms) * 35.45
$$

 where *DBP conc* is the concentration of the DBP (in µg/L) and *DBP M.W.* is the molecular weight of the DBP (in g/mol).



 **Figure 5.** Workflow for the elucidation of the molecular structure of halogenated DBPs with an example of two candidates.

#### **RESULTS AND DISCUSSION**

## *Target analysis of halogenated DBPs*

 Levels of selected halogenated DBP classes in disinfected water samples are summarized in Figure 6, whereas individual concentrations measured for each compound are provided in Table 5. Before chemical disinfection, only trace levels of a few HAAs, namely, dichloroacetic acid (DCAA), trichloroacetic acid (TCAA), dibromoacetic acid (DBAA), bromodichloroacetic acid (BDCAA), and dalapon were present in the DWTPs with concentrations generally below 0.2 µg/L, except for DWTP4, where DCAA, TCAA, and DCBAA were detected between 1.0 and 1.7 µg/L.

553 **Table 5.** Concentration (in ng/L) of the target halogenated DBPs measured in the 554 disinfected water samples (n.d.= not detected). MRL: method reporting limit.





 **Figure 6**. a) Concentrations of the target halogenated DBP classes investigated in chemically disinfected waters and b) proportion (%) of iodine-, bromine- and only chlorine-containing DBPs to target ΣDBP concentrations.

 After chemical disinfection, the highest total concentrations of selected DBPs (ΣDBP) were found in chlorinated waters, DWTP1 (36 µg/L), and DWTP4 (50 µg/L). On the contrary, ΣDBP in chloraminated waters was always <2 µg/L. Overall, the halogenated DBP classes THMs and HAAs were the dominant groups, with a joint average contribution of 92 % to the ΣDBP. It is well known that the use of chlorine enhances the formation of THMs and HAAs as compared to chloramine-based treatments [\[14,](#page-101-5) [61,](#page-107-5) [62\]](#page-107-6). The fact that TCM and BDCM, the main THM species formed during chlorination of waters with low bromide content [\[63,](#page-107-7) [64\]](#page-108-0), were not covered in our study may explain that HAAs contributed more than THMs to ΣDBP in DWTP4, where bromide concentration in source water was quite low (0.054 mg/L), and therefore, high concentrations of TCM and BDCM could form. According to the measurements of regulated THMs conducted by the DWTPs in that period of the year, TCM and BDCM may contribute with 73% and 94% to the total THM concentrations present in DWTP1 and DWTP4, respectively.

 According to the target approach, the formation of iodine-containing DBPs of the investigated waters was in general low (< 8%), which is in agreement with the low iodide levels of the source waters (<limit of quantification (LOQ) of 25 µg/L). As for the potential of the waters to form bromine-containing DBPs (those compounds with at least one bromine atom in their structure, excluding iodo-DBPs that contain also bromine), the highest concentrations were found in waters from DWTP1, where 75% of the DBP mass found was formed by Br-DBPs, followed by waters from DWTP4, with 43% of Br-DBPs. This can be related to the amounts of bromide present in the corresponding non-chemically disinfected waters (0.11 mg/L in DWTP1 and 0.054 mg/L in DWTP4). Note that these contributions of Br-DBPs are higher than real due to failure in capturing TCM and BDCM with the GC-MS conditions used. Because of the high Brˉ levels of DWTP3 source water (0.21 mg/L), the bromine incorporation into NOM during chloramination could also be expected, although it was not reflected in the target analysis, i.e., no Br-DBPs were found. Low bromine incorporation factors into NOM in the presence of chloramine have been consistently reported in the literature [\[65-67\]](#page-108-1). This could be attributed to the negligible formation of HOBr in the presence of 591 chloramine via bromamine formation (NH<sub>2</sub>Cl + Br<sup>−</sup> → NH<sub>2</sub>Br + Cl<sup>−</sup> with k=1.4 × 10<sup>-2</sup>  $M^{-1}s^{-1}$  and NH<sub>2</sub>Br + H<sub>2</sub>O → HOBr + NH<sub>3</sub> with k=1.5 × 10<sup>-3</sup> M<sup>-1</sup>s<sup>-1</sup>) [\[68\]](#page-108-2), or 593 chloramine hydrolysis (NH<sub>2</sub>Cl + H<sub>2</sub>O  $\rightarrow$  HOCl + NH<sub>3</sub>, with a reaction rate constant 594 k=3.0 ×  $10^{-5}$  M<sup>-1</sup>s<sup>-1</sup>) and subsequent reaction of HOCl with bromine [\[68-70\]](#page-108-2), and/or the low stability of bromamines in the solution compared to chloramines [\[70\]](#page-108-3). Moreover, the brominated-DBPs formed during chloramination processes may have not been targeted with our analytical approaches (e.g., amine compounds). This could be the case of bromochloramine, a reaction product of monochloramine with bromine whose

 formation exceeds its decay after 24 h chloramine contact time in a typical drinking water distribution system, as predicted by Liu and Mariñas [\[71\]](#page-108-4).

601 Among the HANs and HACMs, only trace levels  $\langle$  ( $\angle$ 1 µg/L) of dihalogenated species (HANs: dichloroacetonitrile (DCAN) and dibromoacetonitrile (DBAN); HACMs: dichloroacetamide (DCACM) and bromochloroacetamide (BCACM)) were formed after chemical disinfection in all plants, except in DWTP3 (Table 5). THALs were only present in chlorinated waters.

#### *AOX as a surrogate measurement of halogenated DBP mixtures*

 AOX is a bulk measurement of known and unknown DBPs in a sample [\[72,](#page-108-5) [73\]](#page-108-6). AOX concentrations of disinfected waters (Figure 7) were in line with the total target DBP concentrations measured, with decreasing levels in the order DWTP4 > DWTP1 > DWTP2 > DWTP3. The AOX level of water collected at DWTP2 was about half of the levels observed in DWTP1 and DWTP4, in spite a higher TOC content (Table 2). This is consistent with previous studies reporting that chloramine (used at DWTP2) has a lower reactivity towards NOM and hence, results in the formation of lower DBP levels, as compared to chlorine (used at DWTP1 and DWTP4) [\[14,](#page-101-5) [73,](#page-108-6) [74\]](#page-109-0). In the present 616 study, chlorination increased background AOX levels  $\langle$ <14 µg Cl-eq/L) by a factor of 6 and 10 in DWTP1 and DWTP4, respectively, while the AOX increase was only a factor of 3 (DWTP3) or 4 (DWTP2) during chloramination (Figure 7).

 After transforming the concentration of targeted DBPs present in waters into µg Cl- eq/L, it can be concluded that only 27% of the halogenated material formed during the chemical disinfection processes can be explained by the target DBP analysis (in the best-case scenario; DWTP4). This value is similar to, or below the percentage of AOX  explained by known DBPs reported in chlorinated waters in the peer-reviewed literature [\[14,](#page-101-5) [75-77\]](#page-109-1). Note that the inclusion of TCM and BDCM in the list of targeted DBPs would slightly increase the proportion of AOX explained by targeted approaches in chlorinated waters. Considering the contribution of each THM species to total THM concentrations in each plant in that time of the year (data provided by the DWTPs) and the levels of TBM and DBCM measured in our study, the percentage of AOX explained by known DBPs might increase to 74% and 48% in DWTP1 and DWTP4, respectively. The AOX formed in chloraminated DBP mixtures was poorly explained by targeted DBPs, being the Cl-eq DBP concentrations <1% of the AOX. Our results are in agreement with previous studies that reported a larger unknown fraction of AOX in chloraminated than in chlorinated waters [\[14\]](#page-101-5).



 **Figure 7***.* AOX concentrations (µg Cl-eq/L) in waters before and after chemical disinfection. The fraction of AOX explained by target DBP analysis is indicated with the black bars (note that TCM and DCBM were not included in the target analysis). SUVA and TOC levels of source waters are also indicated.

## 639 *Non-target FT-ICR MS analysis of halogenated DBP mixtures*

640 The molecular composition of the DBPs detected by FT-ICR MS in the investigated

641 disinfected waters is summarized in Table 6, and in Figures 8-11. DBP formulae have

642 been listed in Tables 7-22.

643

 **Table 6.** Counts and average neutral mass, elemental proportion, aromaticity, and oxidation degree, weighted by the relative abundance of each verified DBP present in disinfected waters as computed from ESI(-)-FT-ICR mass spectra for singly charged ions. Computations are based on formulae in neutral form and are restricted to formulae present in three technical replicates.

	<b>DWTP1</b>	<b>DWTP2</b>	DWPT3	DWPT4				
# of verified formulae	95	349	151	335				
<b>Neutral mass</b>	388.0	349.0	375.4	376.5				
(Da)	$(288.9 - 660.1)$	$(244.0 - 572.2)$	$(284.0 - 500.0)$	$(256.1 - 600.1)$				
Element proportion in formulae								
C[%]	37.9	37.0	38.3	37.3				
$H[\%]$	41.4	42.7	42.2	41.0				
$O[N_0]$	18.1	17.5	17.2	18.5				
Cl [%]	2.5	2.7	2.0	3.2				
Br [%]	0.1	0.1	0.3	$\overline{0}$				
H/C	1.09	1.13	1.09	1.09				
	$(0.56 - 1.53)$	$(0.20-2.00)$	$(0.64 - 1.50)$	$(0.20-2.00)$				
O/C	0.49	0.48	0.45	0.50				
	$(0.18 - 0.67)$	$(0.11 - 0.75)$	$(0.29 - 0.62)$	$(0.16 - 0.79)$				
Cl/C	0.07	0.08	0.05	0.09				
	$(0-0.15)$	$(0-0.40)$	$(0-0.08)$	$(0.04 - 0.40)$				
Br/C	0.004	0.005	0.008	0.002				
	$(0-0.11)$	$(0-0.20)$	$(0-0.08)$	$(0-0.20)$				
Aromaticity and oxidation degree <sup>a</sup>								
<b>DBE</b>	8.0	6.7	7.9	7.3				
	$(4-18)$	$(0-11)$	$(4-11)$	$(0-18)$				
DBE/C	0.48	0.47	0.48	0.48				
	$(0-0.78)$	$(0-0.80)$	$(0.27 - 0.71)$	$(0-0.80)$				
$AI_{mod}$	0.36	0.35	0.37	0.37				
	$(0.13 - 0.75)$	$(-0.07-1.14)$	$(-0.11 - 0.68)$	$(-0.07-1.14)$				
$\bf{C}_{OS}$	$-0.05$	$-0.089$	$-0.13$	$-0.02$				
	$(-0.7-0.89)$	$(-1.68-1.60)$	$(-0.82 - 0.46)$	$(-1.58-1.60)$				

<sup>649</sup> BBE/C: double bond equivalent relative to the number of carbon atoms,  $AI_{mod}$ :

650 modified aromaticity index;  $C_{OS}$ : carbon oxidation state.



 **Figure 8***.* Molecular composition of the DBP mixtures according to ESI(-)-FT-ICR MS analysis visualized by van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect plots (right panel). Only formulae present in all three replicates are shown.



 **Figure 9.** Plots showing DBE, AImod, and C*OS* versus the number of carbon for verified DBPs (*m/z* ions only present in disinfected water) according to negative ESI-FT-ICR

MS analysis.



Figure 10. Box plots showing the properties of verified formulas in IN and OUT

samples, after FT-ICR MS analysis.





 **Figure 11.** The number of verified chlorinated and brominated DBPs (CHO-type) in the investigated DBP mixtures against the number of oxygen atoms of each DBP composition according to negative ESI-FT-ICR MS analysis.
685 **Table 7**.Nitrogen-containing formulae in the investigated samples after search and

686 formula filtration<sup>a</sup>.

687

Sample code		No. of formulas in the 3 sample replicates	No. of verified formulas (in all replicates)	<b>Theoretical</b> mass (Da) $[M-H]$	<b>Molecular</b> formula [M]	<b>DBE</b>
DWTP1 IN		19		326.09725	$C_{12}H_{26}O_4BrN$	$\overline{0}$
	<b>OUT</b>	24	0			
DWTP2 IN		28	$\overline{0}$			
	<b>OUT</b>	28	3	288.02804	$C_{11}H_{12}O_6ClN$	6
				302.04369	$C_{12}H_{14}O_6ClN$	6
				314.04369	$C_{13}H_{14}O_6ClN$	7
DWTP3	IN	24	$\overline{0}$			
	<b>OUT</b>	22	0			
DWTP4 IN		50	3	610.14959	$C_{30}H_{40}O_3Cl_2BrN$	10
				638.21727	$C_{33}H_{48}O_2Cl_2BrN$	9
				652.19654	$C_{33}H_{46}O_3Cl_2BrN$	10
	<b>OUT</b>	67	5	300.02804	$C_{12}H_{12}O_6Cl_1N$	7
				302.04369	$C_{12}H_{14}O_6Cl_1N$	6
				312.02804	$C_{13}H_{12}O_6Cl_1N$	8
				314.04369	$C_{13}H_{14}O_6Cl_1N$	7
				316.05934	$C_{13}H_{16}O_6Cl_1N$	6
					<sup>a</sup> Masses with equal intensity in IN and OUT are highlighted in italics and grey	

688 *Masses with equal intensity in IN and OUT are highlighted in italics and grey.*

689

691 **Table 8**.Sulfur-containing formulae in the investigated samples after search and

692 formulae filtration<sup>a</sup>.

Sample code		No. of formulas in the 3 sample replicates	No. of verified formulas (in all replicates)	<b>Theoretical</b> mass (Da) $[M-H]$	<b>Molecular</b> formula [M]	<b>DBE</b>
DWTP1 IN		30	$\overline{5}$	413.0922	$C_{17}H_{32}O_2ClBrS$	$\mathbf{1}$
				425.0922	$C_{18}H_{32}O_2ClBrS$	$\mathbf{2}$
				427.10787	$C_{18}H_{34}O_2ClBrS$	$\mathcal{I}$
				427.14425	$C_{19}H_{38}OClBrS$	$\theta$
				453.15990	$C_{2I}H_{40}OClBrS$	$\cal I$
	<b>OUT</b>	30	5	346.73820	$C_5H_3OBr_3S$	3
				427.10787	$C_{18}H_{34}O_2ClBrS$	1
				427.14425	$C_{19}H_{38}OClBrS$	$\theta$
				439.14425	$C_{20}H_{38}OClBrS$	$\mathbf{1}$
				453.1599	$C_{2I}H_{40}OClBrS$	$\boldsymbol{l}$
DWTP2 IN		22	6	411.11295	$C_{18}H_{34}OClBrS$	$\overline{l}$
				413.09222	$C_{17}H_{32}O_2ClBrS$	$\overline{I}$
				413.12860	$C_{18}H_{36}OClBrS$	$\mathcal{O}$
				425.09222	$C_{18}H_{32}O_2ClBrS$	$\mathbf{2}$
				427.10787	$C_{18}H_{34}O_2ClBrS$	$\mathbf{1}$
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\theta$
	<b>OUT</b>	55	5	346.73820	$C_5H_3OBr_3S$	3
				411.11295	$C_{18}H_{34}OClBrS$	$\cal I$
				413.09222	$C_{17}H_{32}O_2ClBrS$	1
				413.12860	$C_{18}H_{36}OClBrS$	$\theta$
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\mathcal{O}$
DWTP3 IN		23	5	425.09222	$C_{18}H_{32}O_2ClBrS$	$\overline{2}$
				427.10787	$C_{18}H_{34}O_2ClBrS$	1
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\theta$
				451.14425	$C_{2I}H_{40}OClBrS$	$\sqrt{2}$
				453.1599	$C_{2I}H_{40}OClBrS$	$\cal I$
	<b>OUT</b>	59	8	413.09222	$C_{17}H_{32}O_2ClBrS$	$\mathbf{1}$
				427.10787	$C_{18}H_{34}O_2ClBrS$	1
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\theta$
				439.14425	$C_{20}H_{38}OClBrS$	1
				451.14425	$C_{2I}H_{40}OClBrS$	$\mathbf{2}$
				453.1599	$C_{2I}H_{40}OClBrS$	1
				477.00526	$C_{21}H_{15}O_9ClS$	14
				507.01582	$C_{22}H_{17}O_{10}CIS$	14
DWTP4 IN		18	5	413.09222	$C_{17}H_{32}O_2ClBrS$	$\mathcal{I}$
				413.12860	$C_{18}H_{36}OClBrS$	$\theta$
				425.09222	$C_{18}H_{32}O_2ClBrS$	$\mathfrak{2}% \left( \mathfrak{2}\right) ^{2}$
				427.10787	$C_{18}H_{34}O_2ClBrS$	$\boldsymbol{I}$
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\mathcal O$
	<b>OUT</b>	24	$\overline{4}$	413.09222	$C_{17}H_{32}O_2ClBrS$	$\overline{I}$
				413.12860	$C_{18}H_{36}OClBrS$	$\theta$
				427.10787	$C_{18}H_{34}O_2ClBrS$	$\boldsymbol{I}$
				427.14425	$C_{18}H_{34}O_2ClBrS$	$\theta$



694 **Table 9**. List of verified formulae of the 19 DBPs common to all four disinfected water 695 samples according to negative ESI-FT-ICR MS analysis (only present in all three 696 replicates of disinfected water).

697



699 **Table 10**. List of verified formulae of the 23 DBPs unique to DWTP1 according to 700 negative ESI-FT-ICR MS analysis (only present in all three replicates of disinfected 701 water).

702 703



704 705

707 **Table 11**. List of verified formulae of the 124 DBPs unique to DWTP2 according to

708 negative ESI-FT-ICR MS analysis (only present in all three replicates of disinfected

709 water).





## 710 **Table 11. (cont.)** 711





713 **Table 12.** List of verified formulae of the 44 DBPs unique to DWTP3 according to

714 negative ESI-FT-ICR MS analysis (only present in all three replicates of disinfected

### 715 water).





717 **Table 13**. List of verified formulae of the 121 DBPs unique to DWTP4 according to

718 negative ESI-FT-ICR MS analysis (only present in all three replicates of disinfected

719 water).



721 **Table 13.** (cont).





- 724 **Table 14**. List of verified formulae of the 49 DBPs common to DWTP1 and DWTP2
- 725 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of
- 726 disinfected water).





728 **Table 15**. List of verified formulae of the 48 DBPs common to DWTP1 and DWTP3 729 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of 730 disinfected water).

731





- 733 **Table 16**. List of verified formulae of the 47 DBPs common to DWTP1 and DWTP4
- 734 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of
- 735 disinfected water).





737 **Table 17**. List of verified formulae of the 80 DBPs common to DWTP2 and DWTP3

738 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of disinfected water).



741 **Table 18**. List of verified formulae of the 190 DBPs common to DWTP2 and DWTP4

742 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of

743 disinfected water).



745 **Table 18.** (cont.)

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748 **Table 18.** (cont.)



- 751 **Table 19**. List of verified formulae of the 61 DBPs common to DWTP3 and DWTP4
- 752 according to negative ESI-FT-ICR MS analysis (only present in all three replicates of
- 753 disinfected water).



C16 H17 O7 Cl1 355.05901 C15 H15 O8 Cl1 357.03827 C15 H19 O8 Cl1 361.06957 C17 H15 O7 Cl1 365.04336 C18 H19 O6 Cl1 365.07974 C17 H17 O7 Cl1 367.05901

- 755 **Table 20.** List of verified formulae of the 33 DBPs common to DWTP1, DWTP2, and
- 756 DWTP3 according to negative ESI-FT-ICR MS analysis (only present in all three 757 replicates of disinfected water).



- 760 **Table 21**. List of verified formulae of the 23 DBPs common to DWTP1, DWTP3, and
- 761 DWTP4 according to negative ESI-FT-ICR MS analysis (only present in all three 762 replicates of disinfected water).



765 **Table 22.** List of verified formulae of the 45 DBPs common to DWTP2, DWTP3, and 766 DWTP4 according to negative ESI-FT-ICR MS analysis (only present in all three 767 replicates of disinfected water).

768





#### *Changes of the molecular composition of halogenated NOM during disinfection*

 In the case of DWTP1, 5% of the substances verified in disinfected water were also identified in the water before disinfection. In the case of the other investigated DWTPs the formulae overlay in non-disinfected and disinfected waters ranged between 20 and 22% (Figure 12).



 **Figure 12.** Venn diagrams showing the number of molecular formulae unique and common to non-disinfected (IN) and disinfected water (OUT) in each investigated drinking water treatment plant, after non-target FT-ICR MS analysis. 

 The weighted average molecular mass (weighted against relative intensities) of Cl- and Br-containing substances decreased during the chemical disinfection of water. However, 782 this decrease was statistically significant only in DWTP2 and DWTP3 (p<0.05, Table 23). Verified halogenated formulae in chemically-disinfected water samples had a lower 784 H/C ratio and a higher O/C ratio,  $AI_{mod}$  and  $C_{OS}$  than those in non-disinfected waters.

 These differences were statistically significant in most cases (except for H/C ratio, 786 DBE, and  $AI_{mod}$  in DWTP2 and DBE and  $AI_{mod}$  in DWTP4) with a confidence level of 95% (p<0.05, Tables 24-28). These differences can be explained by the specific reactivity of the chemical disinfectants with NOM, forming aromatic halogenated compounds with high C-C double bond density and DBE (Figure 10).

790 **Table 23**. Statistics for comparison of the molecular mass of verified Cl and Br 791 formulae in the investigated samples, after negative ESI-FT-ICR MS analysis**.**

792



793 \*When p-value <0.05, the molecular mass of the Cl and Br-formulae before and after 794 disinfection are significantly different with a significance level of 5%. Overlapping features 795 between IN and OUT were removed.

796  $*$ When p-value <0.05, the molecular mass of the Cl and Br-formulae in disinfected water of 797 the different DWTPs are significantly different with a significance level of 5%. Pairwise 797 the different DWTPs are significantly different with a significance level of 5%. Pairwise<br>798 comparison with a posthoc Dunn's test allows identification of the differences. comparison with a posthoc Dunn's test allows identification of the differences.

801 **Table 24**. Statistics for comparison of the H/C content of verified Cl and Br formulae in 802 the investigated samples, after negative ESI-FT-ICR MS analysis**.** 

803



804 \*When p-value <0.05, the H/C content of verified Cl and Br-formulae before and after 805 disinfection are significantly different with a significance level of 5%. Overlapping features 806 between IN and OUT were removed.

807  $*$ When p-value <0.05, the H/C content of verified DBPs in disinfected water of the different 808 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a 808 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a 809 posthoc Dunn's test allows identification of the differences.

- 810
- 811

812 **Table 25**. Statistics for comparison of the O/C content of verified Cl and Br formulae in 813 the investigated samples, after negative ESI-FT-ICR MS analysis**.**

814



815 \*When p-value  $\langle 0.05 \rangle$ , the O/C content of verified Cl and Br-formulae before and after 816 disinfection are significantly different with a significance level of 5%. Overlapping features disinfection are significantly different with a significance level of 5%. Overlapping features 817 between IN and OUT were removed.

818 \*\*When p-value <0.05, the O/C content of verified DBPs in disinfected water of the different

819 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a

820 posthoc Dunn's test allows identification of the differences.

821

822

823

825 **Table 26.** Statistics for comparison of the AI<sub>mod</sub> of verified Cl and Br formulae in the 826 investigated samples, after negative ESI-FT-ICR MS analysis**.**

827



828 \*When p-value <0.05, the AImod content of verified Cl and Br-formulae before and after 829 disinfection are significantly different with a significance level of 5%. Overlapping features

830 between IN and OUT were removed.

831 \*\*When p-value <0.05, the AImod content of verified DBPs in disinfected water of the different

832 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a

833 posthoc Dunn's test allows identification of the differences.

834 835

836 Table 27. Statistics for comparison of the C<sub>OS</sub> of verified Cl and Br formulae in the 837 investigated samples, after negative ESI-FT-ICR MS analysis**.**

838



839 \*When p-value  $\sqrt{0.05}$ , the  $\text{C}_{OS}$  content of verified Cl and Br-formulae before and after 840 disinfection are significantly different with a significance level of 5%. Overlapping features 841 between IN and OUT were removed.

842 \*\*When p-value <0.05, the  $C_{OS}$  content of verified DBPs in disinfected water of the different 843 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a 844 posthoc Dunn's test allows identification of the differences.

845

848 **Table 28.** Statistics for comparison of the DBE of verified Cl and Br formulae in the 849 investigated samples, after negative ESI-FT-ICR MS analysis**.**

850



851 \*When p-value <0.05, the DBE content of verified Cl and Br-formulae before and after 852 disinfection are significantly different with a significance level of 5%. Overlapping features 853 between IN and OUT were removed.<br>854 \*\*When n-value < 0.05, the DBE cor

 $*$ When p-value <0.05, the DBE content of verified DBPs in disinfected water of the different 855 DWTPs are significantly different with a significance level of 5%. Pairwise comparison with a 856 posthoc Dunn's test allows identification of the differences.

857

#### 858 *Cl- and Br- compounds in DBP mixtures*

 The contribution of different groups of halogenated substances to the DBP mixture chemodiversity in all disinfected waters is summarized in Figure 13. Monochlorinated compounds (CHOCl) contributed the most to the total DBP mixture chemodiversity in 862 all disinfected waters (65-75%) except in DWTP4, where both CHOCl and CHOCl<sub>2</sub> were equally relevant (49% each). Bromine incorporation into NOM led to the formation of monobrominated substances (CHOBr) in the order DWTP3 (23%) > 865 DWTP1 (14%) > DWTP2 (9%) > DWTP4 (2%). Even higher bromination rates were expected to occur in DWTP3, according to the results of a previous study, where chloramination of source waters with a slightly higher concentration of bromide (0.28 mg/L) than in DWTP3 (0.22 mg/L) resulted mainly in the formation of CHOBr compounds [\[26\]](#page-102-0). This finding, which has also been confirmed by target analyses in this study, could be associated with a dominant presence of aromatic DBP precursors in  DWTP3 source water (as indicated by SUVA measurements, Table 2). Thus, this could result in low incorporation of bromine into NOM during chloramination, as reported elsewhere [\[78\]](#page-109-0).

874 Dichlorinated compounds (CHOCl<sub>2</sub>) were not present in DWTP3, but were the second most abundant group formed in the remaining DWTPs. As previously mentioned, 876 CHOCl<sub>2</sub> make up 49% of the formulae found in DWTP4, where chlorination of the source water with the lowest amount of bromide (0.05 mg/L) occurred, but accounted for less than 24% of the halogenated substances found in DWTP1 and DWTP2 disinfected waters (Figure 13). In this regard, the halogenated chemical space covered by the FT-ICR MS analysis in DWTP4 gives evidence that substances highly substituted with chlorine are formed during the chlorination of waters with low bromide content in agreement with previous studies [\[64,](#page-108-0) [79\]](#page-109-1). This is also confirmed by the DBPs found in DWTP4 waters with the target approach (Figure 6 and Table 5).





 **Figure 13**. Contribution of each group of halogenated compounds to the chemodiversity of the investigated disinfected waters, after FT-ICR MS analysis. Y-axis shows the percent of verified molecular formulae.

 Finally, a small group of halogenated DBPs containing one Br and one Cl atom (CHOClBr) was also found in DWTP2 and DWTP4, and constituted 3% and 1.5%, respectively, of the total formulae verified in these samples. In DWTP4, these formulae 893 have DBE between 0 and 1, very low  $AI_{mod}$  (<0), high H/C ratio (1.8-2), and low O/C ratio (0.2-0.3). Consequently, they correspond to aliphatic compounds (Figures 8 and 10). In DWTP2, most of the verified CHOClBr DBPs have an aromatic character (DBE 896 of 5 or 6, AI<sub>mod</sub> between 0.38 and 0.57, H/C ratio  $\leq$  1, and relatively high O/C ratio of 0.6-0.7).

 The investigated DBP mixtures contained only two highly halogenated formulae, *viz*.  $C_{33}H_{28}O_6Cl_4$  in DWTP1 and DWTP4 and  $C_5HO_3Cl_2Br$  in DWTP2. The CHOCl<sub>4</sub> formula corresponded with the verified DBP of the highest molecular weight. This and the non-detection of additional highly halogenated formulae may suggest that these type of compounds are unstable or intermediate DBPs that may rapidly alter via hydrolysis to smaller compounds; or that the specific precursors of this type of DBPs, required for their formation, were not abundant in these source waters [\[80\]](#page-109-2).

#### *Specific molecular composition of DBP mixtures of each water treatment plant*

 In total, 19 formulae, all of them corresponding with monochlorinated compounds were observed to occur in all disinfected waters; whereas 23, 124, 44, and 121 were unique to DWTP1, DWTP2, DWTP3, and DWTP4, respectively (Figure 14, and Tables 7-22). The molecular composition of the common DBPs and DBPs unique to each water treatment plant is summarized in Figures 15-18. In the case of DWTP3, unique DBPs were mainly CHOBr compounds, whereas in DWTP4 unique DBPs were dominated by  $CHOCl<sub>2</sub>$  formulae.



 **Figure 14.** Venn diagram showing the chemodiversity of the investigated DBP mixtures according to ESI(-)-FT-ICR MS analysis.

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- 



 **Figure 15.** Molecular composition of the DBPs formed in all DWTPs according to ESI(-)-FT-ICR MS analysis visualized by van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect (right panel). Only formulae present in all three replicates are shown.



927 **Figure 16**. Plots showing DBE, AI<sub>mod</sub>, and C<sub>OS</sub> versus the number of carbon for verified 928 DBPs ( $m/z$  ions only present in disinfected water) common to all DWTPs according to negative ESI-FT-ICR MS analysis.

 The weighted average molecular mass of DBPs was very similar in all disinfected waters, being all distributed within the mass range of 244 - 660 Da (Table 6). However, 933 the distribution of the molecular mass of the  $m/z$  ions in DWTP2 was slightly lower than 934 that observed in the other plants  $(p<0.001)$ . This was also true for the distribution of DBE in the DBPs identified in DWTP2. DWTP2, together with DWTP4, presented the highest diversity of bromine and chlorine-containing features identified as DBPs (Figures 8 and 17). Thus, heterogeneity of the mixture seems to be associated to the properties and amount of NOM in the source water rather than the disinfectant applied. It is worthy to highlight that the number of molecular formulae verified in one sample is subject to a very conservative verification approach (i.e., the *m/z* ion should appear above an established threshold in all three replicates). In this regard, samples yielding more verified formulae are more representative of the true chemodiversity of the mixture than samples with fewer formulae. However, this does not necessarily translate 944 into a higher mixture heterogeneity, since the intensity of an  $m/z$  ion in a sample is highly depending on matrix effects and intensities of other formulae in the sample, and as a result, the number of formulae present in the sample may be underestimated.



 **Figure 17.** Molecular composition of the DBPs unique to each DWTP according to ESI(-)-FT-ICR MS analysis visualized by van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect plots (right panel). Only formulae present in all three replicates are shown.





966 **Figure 18**. Plots showing DBE, AI<sub>mod</sub>, and C<sub>OS</sub> versus the number of carbon for unique verified DBPs (*m/z* ions only present in disinfected water) according to negative ESI-FT-ICR MS analysis.

 While mono and diCl–DBPs were dominant in DWTP4 disinfected water, the DBP mixture in DWTP3 was dominated by monohalogenated Br– and Cl–DBPs. Besides  $C_{33}H_{28}O_6Cl_4$  and  $C_{17}H_{34}O_4ClBr$ , no other di- or higher halogenated formulae were present in DWTP3 disinfected water. The computed weighted average O/C ratio and  $C_{OS}$  of the formulae verified in DWTP3 were significantly different (lower) compared 975 to the formulae verified in the other investigated DBP mixtures  $(p<0.001)$ . This could be partially attributed to the use of chloramine for disinfection that has a lower oxidation potential than chlorine.

 The –KMD/z\* diagrams revealed two major groups of DBPs in each DBP mixture 979 (Figure 8); one group located in the lower region of the diagram  $(-KMD/z^* < 0.12)$ , characterized by unsaturated compounds, and one group, located in the upper region of 981 the diagram ( $-KMD/z^* > 0.12$ ), mainly formed by highly oxygenated and unsaturated 982 compounds. Specifically, CHCl<sub>2</sub> formulae distribute in three regions in DWTP2 and

983 DWTP4  $(-KMD/z^* < 0.05, -KMD/z^*$  around 0.10 and  $-KMD/z^* > 0.12$ , which could indicate that each of these groups arises from different precursors.

 The average Cl/C ratio of the verified formulae decreased in the order DWTP4>DWTP2>DWTP1>DWTP3, while the average Br/C ratio decreased as follows DWTP3>DWTP2>DWTP1>DWTP4 (Table 2).

# *Non-target LC-ESI(-)-Orbitrap MS analysis to identify DBPs in halogenated DBP mixtures*

 Using LC-ESI(-)-Orbitrap MS, a total of 81, 129, 54, and 116 newly formed halogenated and non-halogenated features with abundances above 100,000 counts were found in all three triplicate samples of DWTP1, DWTP2, DWTP3, and DWTP4. The halogenated features were compared to those detected by FT-ICR MS. Only few formulae were detected using both techniques (i.e., 286.91968; 243.00658, 259.00149; 300.96761; 335.04586; 255.04296; 256.84134; 318.94179). The low percentage of agreement between the halogenated features detected with both techniques could be attributed to: *(i)* the chromatographic column including retention factor, selectivity and/or efficiency, and *(ii)* the incompatibility of some DBPs with the mobile phase used in the LC-ESI(-)-Orbitrap MS approach, *(iii)* the loss of some DBPs during the second SPE preconcentration process for Orbitrap MS analysis, *(iv)* interference problems related to the ion suppression phenomenon (that may vary between the ESI ion source configurations used, and reduce after chromatographic separation of sample components), *(v)* the use of different data processing tools (e.g., the algorithm used for peak deconvolution of LC-Orbitrap MS data) [\[81\]](#page-110-0), or *(vi)* a mixture of all these factors. Besides, the DBP with the lowest *m/z* confirmed with FT-ICR MS had a nominal *m/z* of 243, whereas many of the DBPs detected with Orbitrap MS were below this value. This could be attributed to on the one hand the higher mass cutoff set in FT-ICR MS compared to LC-Orbitrap MS, and also the low capability of the direct infusion approach to detect ions in the low m/z range. Direct infusion is highly affected by ion suppression effects as all matrix components are analyzed at once, and this may

 condition the detection of low m/z ions. The implementation of LC before FT-ICR MS is limited by the acquisition speed of the ICR cell operated.

 Contrary to FT-ICR MS instruments, the Q-Exactive, due to its hybrid nature (Quadrupole-Orbitrap MS) provides structural information of the different ions in the mixture. Thus, it allows assigning a molecular structure for most of the halogenated DBPs present in the investigated disinfected samples.

- Despite that the iodo-acids found in the target approach are indeed amenable to ESI(-) [\[82,](#page-110-1) [83\]](#page-110-2), iodo-DBPs were not detected in the samples using LC-Orbitrap MS. This can be attributed to the fact that their concentrations were below the limit of detection of the technique, or they were not captured with the extraction method used (water pH during extraction was equal to the highest pKa of iodo-acids that were detected).
- The workflow used (Figure 5) allowed identifying in total 86 DBPs (including isobaric species), which corresponded with 43% (DWTP1) - 70% (DWTP3) of the newly formed features. Most of the identifications were obtained with a confidence level of 3, according to Schimanski's scale [\[84\]](#page-110-3), i.e., there were identification pieces of evidence from MS2 data for proposing a specific molecular structure, but this could not be confirmed. A confidence level of 1 was achieved for 4 compounds, specifically, for 4 HAAs after injection of an extract aliquot fortified with pure analytical standards, and comparison of their retention time and fragmentation pattern.
- The DBPs tentatively identified are listed in Table 29. According to the structures proposed, most DBPs identified are highly unsaturated and phenolic compounds, which is in agreement with their properties, summarized in Table 30 and Figures 19-21.

## 1034 **Table 29.** DBPs identified after LC-ESI(-)MS/MS analyses with QExactive MS.






- The highest score in MetFrag





























1036  $\overline{X}$  presence not confirmed,  $\sqrt{1/\sqrt{\pi}}$  peak area $>10e^8$ ,  $\sqrt{\sqrt{\pi}}$  peak area $>10e^7$ ,  $\sqrt{\pi}$  peak area  $>10e^6$ , T: Trace amounts. 1036<br>1037

 **Table 30.** Counts and average neutral mass, elemental proportion, aromaticity, and oxidation degree, weighted by the relative abundance of each DBP identified in disinfected waters as computed from LC-ESI(-)-Orbitrap mass spectra for singly charged ions. Computations are based on formulae in neutral form and are restricted to formulae present in three technical replicates.



1045 <sup>a</sup> DBE/C: double bond equivalent relative to the number of carbon atoms, AI<sub>mod</sub>: 1046 modified aromaticity index;  $C_{OS}$ : carbon oxidation state.

1047 \*Only those halogenated DBPs for which a unique molecular formula could be assigned

1048 were considered in the calculations.

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 **Figure 19**. Contribution of each group of halogenated compounds to the chemodiversity of the investigated disinfected waters, after LC-ESI(-)-Orbitrap MS analysis. Y-axis shows the percent of confirmed or tentatively identified structures.



 **Figure 20**. Molecular composition of the DBPs of each DWTP according to LC-ESI(-)- Orbitrap MS analysis, visualized by van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect plots (right panel). Only formulae present in all three replicates are shown.



 (*m/z* ions only present in disinfected water) according to LC-ESI(-)-Orbitrap MS analysis.

 In agreement with ESI(-) FT-ICR MS results, monochlorinated compounds (CHOCl) contributed the most to the total DBP mixture (31-59%), followed by dichlorinated compounds (CHOCl2) (11-24%) (Figure 19). LC-ESI(-)-Orbitrap MS revealed the presence of dibrominated (2-15% of total DBPs) and trihalogenated species (3-13% of total DBPs) in the mixture. However, the formation of highly substituted (3 halogens) was overall minor. As for brominated features, more Br-DBPs were detected with LC- Orbitrap MS than with FT-ICR MS. They decreased in the order DWTP1 (40%) > 1091 DWTP3 (30%) > DWTP2 (25%) > DWTP4 (18%). This finding could be attributed to both the bromide content of source waters (DWTP3 > DWTP1 > DWTP2 > DWTP4, Table 2), and the disinfection treatment applied, where chlorination (DWTP1) is likely to form more Br-DBPs than chloramination (DWTP3).

 LC-Orbitrap MS analysis confirmed that substances highly substituted with chlorine (dichloro- and trichloro-DBPs) are formed during the chlorination of water with low bromide content, as previously published [\[64,](#page-108-0) [79\]](#page-109-0) and indicated by target analysis (Figure 6 and Table 5) and FT-ICR MS analysis (Figure 13).

 Halogenated DBPs containing one Br and one Cl atom (CHOClBr) were detected in all samples and constituted between 6% and 10% of the total DBPs identified in these samples.

## *Specific molecular composition of DBP mixtures of each water treatment plant*

In total, 18 formulae were observed to occur in all disinfected waters; whereas 4, 17, 2,

and 6 were unique to DWTP1, DWTP2, DWTP3, and DWTP4, respectively (Figure 22

and Table 29). The molecular composition of the common DBPs and DBPs unique to

 each DWTP is summarized in Figures 23-26. Common DBPs included mostly monochlorinated and dichlorinated compounds, but also the confirmed HAAs dibromoacetic acid and bromochloroacetic acid, and 4-bromo-5-methoxy-benzene-1,3- diol (*m/z* 216.0505). Unique DBPs in DWTP1 were mostly dibrominated compounds, whereas exclusive monochlorinated compounds were mainly formed in DWTP2 and DWTP4.



1113<br>1114 **Figure 22.** Venn diagram showing the chemodiversity of the investigated DBP mixtures according to LC-ESI(-)-Orbitrap MS analysis.

 The weighted average molecular mass of DBPs was very similar in all disinfected waters, being all distributed within the mass range of 94 - 346 Da (Table 30) (no statistically significant differences were found). Although the scan range was comparable, the average mass of the DBPs identified with LC-Orbitrap MS was about 100 Da lower than that of DBPs characterized using FT-ICR MS. In line with FT-ICR MS analysis, DWTP2 and DWTP4 were the mixtures with the highest heterogeneity of Cl- and Br-DBPs; however, the approach used to process LC-Orbitrap MS data is also

- conservative and the analytical technique also affected by matrix effects, which means
- that the heterogeneity of the other DBP mixtures may be underestimated.
- 



1128<br>1129 Figure 23. Molecular composition of the DBPs formed in all DWTPs according to LC- ESI(-)-Orbitrap MS analysis. van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect plots (right panel) of the compounds present in the disinfected samples. Only formulae present in the three replicates are shown.

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DBPs common to all DWTPs



 

1141 **Figure 24.** Plots showing DBE,  $AI_{mod}$ , and  $C_{OS}$  versus the number of carbon for verified DBPs (*m/z* ions only present in disinfected water) common to all DWTPs according to LC-ESI(-)-Orbitrap MS analysis.



 **Figure 25**. Molecular composition of unique DBPs according to LC-ESI(-)-Orbitrap MS analysis. van Krevelen diagrams (left panel), mass edited H/C ratios (middle panel), and modified Kendrick mass defect plots(right panel) of the compounds present in the disinfected samples. Only formulas present in the three replicates are shown.

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1165 **Figure 26**. Plots showing DBE, AI<sub>mod</sub>, and C<sub>OS</sub> versus the number of carbon for verified DBPs (*m/z* ions only present in disinfected water) in unique DBPs according to LC-ESI(-)-Orbitrap MS analysis.

1168 The weighted average O/C ratio and  $C_{OS}$  of the formulae identified in chloraminated mixtures (DWTP2 and DWTP3) were lower than that in chlorinated mixtures (although differences were not statistically significant). This could be partially attributed to the use of chloramine for disinfection that has a lower oxidation potential than chlorine.

 Similarly as in FT-ICR MS results, the –KMD/z\* diagrams revealed two groups of DBPs in each DBP mixture (Figure 20). One group, containing most of the features, is 1174 characterized by less unsaturated compounds  $(-KMD/z^* < 0.12)$ , and the other group, containing only a few features, is mainly formed by highly oxygenated and thus, highly unsaturated compounds (–KMD/z\* >0.12).

 The average Cl/C ratio of the verified formulae decreased in the order DWTP4>DWTP2>DWTP1>DWTP3, while the average Br/C ratio decreased as follows DWTP1>DWTP2>DWTP3>DWTP4 (Table 30). Thus, this finding confirms that the disinfection of low bromide containing waters results in formulae with high chlorine incorporation ratios, whereas the chlorination of high bromide containing waters favors bromide incorporation into NOM.

 Comparing the characteristics of the DBPs verified/identified with the two non-target approaches employed in this study (Tables 6 and 30), it can be concluded that different DBP groups in the mixture were captured with each approach, despite the use of the same ionization source. Overall, halogenated DBPs identified with LC-Orbitrap MS presented on average a higher bromine incorporation factor, a higher DBE per number 1189 of carbon atoms, a higher carbon  $C_{OS}$ , and  $AI_{mod}$  than those detected with FT-ICR MS. Furthermore, a very small overlap was obtained among the DBPs identified with the different approaches used (4 HAAs between the target GC-MS screening and LC-Orbitrap MS and 8 compounds between LC-Orbitrap MS and FT-ICR MS). Thus, this  study demonstrates the relevance of employing different analytical techniques to unravel the chemodiversity of DBP mixtures.

## **CONCLUSIONS**

 Target screening of DBPs at four Swedish DWTPs accounted partially for the halogenated material formed during disinfection processes. The non-target analysis evidenced a wide diversity of the halogenated DBP mixtures formed. The large differences observed in the DBP mixture composition among the investigated DWTPs indicate that DBP formation is highly dependent on local conditions (disinfection treatment and water source characteristics). This makes the development of models to predict DBP formation extremely complicated. Furthermore, the regulated volatile DBPs routinely monitored (THMs) may not adequately reflect the local DBP composition, and efforts to monitor an extended set of DBPs such as in this study should be applied at each particular case. For the evaluation of the DBP mixture chemodiversity, the use of complementary analytical tools is recommended, as evidenced in this work.

 Although only a few of the DBPs detected using HRMS analyses were confirmed with pure analytical standards, tentative identified DBPs indicate that they are highly polyunsaturated and polyphenolic compounds. These 86 DBPs identified can be used to design suspect lists that improve the characterization of halogenated compounds in waters disinfected with chlorine-based agents. Efforts should be made in the future to confirm the identity of these DBPs as well as to assess the relevance of their concentrations.

 One of the main limitations of non-targeted approaches for exploring DBP mixtures is the impossibility of extracting all DBPs formed and capturing all with a single analytical technique, due to the different nature of these compounds. While purging and trapping procedures aimed at extracting volatile DBPs (e.g., THMs), the use of solid- phase extraction techniques is directed for retaining a wide range of hydrophobic to hydrophilic compounds. Like in this study, generic-purpose sorbents are commonly employed for non-target screening of DBPs. However, the characterization of the most polar fraction of the DBP mixture could be also possible with the use of ion-exchange cartridges. For this, hydrophilic interaction liquid chromatography (HILIC) coupled to HRMS may play a relevant role.

 Based on this, the non-target screening approach used in this study covered only medium to low polar compounds amenable to ESI(-). Thus, highly polar compounds and volatile compounds were excluded. Because of the ionization technique used, the identification is limited mainly to compounds containing carboxylic, carbonyl, and alcohol moieties, and ion suppression further drastically favors carboxylic acids over carbonyl and alcohols. The use of different ionization methods (e.g., positive ESI, photoionization), and the development of highly sensitive and specific data processing workflows that allow capturing DBPs present at low concentrations could contribute to unveil the remaining unknown fraction of AOX.

 The AOX fraction not (un)covered in our approach may include halogenated polyunsaturated and polyphenolic compounds (like the ones found in this study but present at levels below the method detection limit), nitrogen-containing DBPs with different heteroatoms (amines or amides, not hydrolyzed under the acidic conditions of the extraction procedure and thus, amenable to positive ESI), and high molecular weight halogenated fulvic acid molecules little fragmented, as suggested elsewhere [\[85\]](#page-110-0). Thus,

 efforts should be conducted in the future to characterize this unknown fraction and evaluate its bioactivity.

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