

ISME Communications

Air monitoring by nanopore sequencing

--Manuscript Draft--

Manuscript Number:	
Full Title:	Air monitoring by nanopore sequencing
Article Type:	Original Article
Keywords:	air microbiome; bioaerosols; metagenomics; nanopore sequencing; long-read sequencing; shotgun sequencing; de novo assembly; infectious disease; antimicrobial resistance; urban air microbiome
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Additional Information:	
Question	Response
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<p>Has this manuscript been submitted to this journal previously?</p>	<p>Yes</p>
<p>Please indicate the manuscript number for the previous submission. as follow-up to "Has this manuscript been submitted to this journal previously?"</p>	<p>ISMEECOMMUN-D-24-00010</p>



1 **Air monitoring by nanopore sequencing**

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24

25 **Abstract**

26 While the air microbiome and its diversity are essential for
27 human health and ecosystem resilience, comprehensive air
28 microbial diversity monitoring has remained rare, so that little is
29 known about the air microbiome's composition, distribution, or
30 functionality. Here we show that nanopore sequencing-based
31 metagenomics can robustly assess the air microbiome in
32 combination with active air sampling through liquid impingement
33 and tailored computational analysis. We provide fast and
34 portable laboratory and computational approaches for air
35 microbiome profiling, which we leverage to robustly assess the
36 taxonomic composition of the core air microbiome of a controlled
37 greenhouse environment and of a natural outdoor environment.
38 We show that long-read sequencing can resolve species-level
39 annotations and specific ecosystem functions through *de novo*
40 metagenomic assemblies despite the low amount of fragmented
41 DNA used as an input for nanopore sequencing. We then apply
42 our pipeline to assess the diversity and variability of an urban air
43 microbiome, using Barcelona, Spain, as an example; this
44 randomized experiment gives first insights into the presence of
45 highly stable location-specific air microbiomes within the city's
46 boundaries, and showcases the robust microbial assessments
47 that can be achieved through automatable, fast, and portable
48 nanopore sequencing technology.

49

50 **Introduction**

51 The air microbiome encompasses a broad spectrum of
52 bioaerosols, including bacteria, archaea, fungi, viruses,
53 bacterial endotoxins, mycotoxins, and pollen [1]. While its
54 pivotal functions for human health and ecosystem resilience are
55 recognized, little is known about its composition, distribution,
56 and functionality [2]. Past research efforts, particularly those
57 driven by infectious diseases such as COVID-19 and
58 tuberculosis, have shifted the research focus towards potentially
59 pathogenic microbial taxa; however, exposure to a diverse air
60 microbiome has also been increasingly considered as a health-
61 promoting factor, underscoring the need for holistic air microbial
62 diversity monitoring [3].

63 Most genetics-based air microbiome studies have employed
64 targeted sequencing via metabarcoding due to the low biomass
65 of bioaerosols [1, 4]. While metabarcoding increases the
66 sensitivity of taxonomic detection, it is inherently limited by
67 amplification biases and incomplete databases. In contrast,
68 metagenomics, which is based on shotgun sequencing of native
69 DNA, avoids amplification biases and allows for *de novo*
70 reconstructions of microbial genomes for robust species
71 identification and functional annotation. Such metagenomic
72 approaches have also recently been applied for low biomass
73 bioaerosol analysis [5] and have revealed the complex nature
74 and diverse origins of the air microbiome [4], including vertical-

75 altitudinal stratification of microbial abundance and distribution
76 [6], and substantial diurnal, seasonal, temperature-, and
77 humidity-dependent fluctuations [7].

78 These metagenomic assessments of the air microbiome have
79 thus far relied on short-read sequencing technology, which
80 provides accurate sequencing data but hampers *de novo*
81 assemblies, especially of highly repetitive genomic regions, and
82 accurate species- or strain-level identification due to the
83 inherently short sequencing reads; long-read sequencing, on
84 the other hand, has facilitated *de novo* genome assemblies [8]
85 and assessments of highly repetitive genomic regions, including
86 the detection of antimicrobial resistance genes [9], from
87 metagenomic data. Especially recent advances in nanopore
88 sequencing technology have made long-read sequencing
89 increasingly relevant for microbial diversity assessments due to
90 the technology's substantially improving sequencing accuracy
91 [10, 11] while maintaining its long-read sequencing capacity and
92 its automatable [12], fast, and portable deployability for
93 applications in clinical [13] or remote settings [14]. While
94 nanopore sequencing has been used to characterize the
95 microbial diversity of various environments, such as of
96 freshwater [15] and dust [16], no approaches have yet been
97 established to leverage the technology's unique advantages for
98 monitoring the taxonomic and functional diversity of the air
99 microbiome.

100 Here, we established laboratory and computational approaches
101 to enable robust air microbiome profiling through nanopore
102 metagenomics. We first evaluated the suitability of long-read
103 shotgun sequencing for assessing the air microbiome in a
104 controlled indoor environment, and then applied our approaches
105 to an outdoor environment for validation. We showed that
106 nanopore sequencing is a robust tool to describe the
107 composition and diversity of microbial taxa in the air, and to
108 concurrently annotate *de novo* microbial genomes to evaluate
109 potential human health consequences. We finally applied our
110 laboratory and computational approaches to conduct a
111 randomized air sampling campaign in Barcelona, Spain, to
112 robustly describe its urban air microbiome.

113

114 **Materials and Methods**

115 We first conducted preliminary tests to compare standard air
116 sampling and DNA extraction approaches for nanopore
117 sequencing-based air metagenomics; this included the testing
118 of standard quartz filter- and liquid impingement-based air
119 samplers and the optimization of respective DNA extraction
120 approaches for subsequent nanopore shotgun sequencing,
121 which relies on minimum DNA input without nucleotide
122 amplification and is sensitive to native DNA contamination
123 (**Supplementary Information: Air sampling and DNA extraction**
124 *optimizations*).

125 Based on these preliminary tests, we decided to use the Coriolis
126 μ liquid impinger (Bertin Instruments, France; (**Supplementary**
127 **Information: Air sampling and DNA extraction optimizations**) for
128 air sampling, which uses cyclonic forces to concentrate airborne
129 biomass into a collection liquid in a cone. We used 15 mL of
130 ultrapure water with 0.005 % Triton-X (Sigma-Aldrich, Germany)
131 as collection liquid, which functions as a nonionic surfactant to
132 enhance organic compound solubility and surface enlargement
133 due to foam generation. The liquid impinger was positioned at
134 1.5 m above the ground to sample air within the human
135 breathable zone, which ranges from 1.4 to 1.8 m. We operated
136 the liquid impinger at an air flow rate of 300 L min⁻¹ and at a
137 collection liquid refilling rate of 0.8 mL min⁻¹ to counter liquid
138 evaporation during sampling. After sampling, we directly
139 transferred the collected liquid into a sterile 15 mL falcon tube.
140 We then divided the liquid across three 5 mL tubes, centrifuged
141 them at 18 000 x g for 25 min, and collected the pellets. The
142 pellets were resuspended, aggregated, and subsequently
143 centrifuged twice at 18 000 x g for 25 min while discarding the
144 supernatant.

145 We first sampled air in a greenhouse (“Gh”; Helmholtz Munich
146 Environmental Research Unit) as a controlled environment with
147 moderate human activity and continuous air circulation (mean
148 ambient temperature of 23 °C); we sampled air for three
149 consecutive days, either for 1h in three consecutive replicates
150 per day or for 3h with one replicate per day (**Supplementary**
151 **Table 1**). We next sampled air in a natural environment (“Nat”),

152 namely on the Helmholtz Munich campus on the outskirts of
153 Munich (48.220889, 11.597028), which is mainly surrounded by
154 natural grassland. We sampled for six consecutive days,
155 following an alternating pattern of 3h or 6h of air sampling; we
156 here tested 6h as sampling duration since we expected a higher
157 variability in the air microbiome in comparison to the controlled
158 greenhouse setting (**Supplementary Table 1**). The liquid
159 impinger was positioned in a shaded area to avoid significant
160 thermal fluctuations. While the weather remained relatively
161 constant and sunny across the six sampling days (ambient
162 temperature ranged from 21°C to 25°C, and humidity from 42%
163 to 71%), we note that the 6h-sample from day 4 was affected
164 by rain and thunderstorm at the end of the sampling activity. We
165 finally collected urban air samples in Barcelona, Spain, from
166 October 16th to November 3rd, 2023. We sampled five different
167 urban locations: Gracia (“Residential Area”, 41.398861,
168 2.153490), Eixample (“City Center”, 41.385500, 2.155103),
169 Poblenou (“Urban Beach”, 41.404135, 2.206550), Vall d’Hebron
170 (“Outer Belt”, 41.425887, 2.148349), and Observatori Fabra
171 (“Green Belt”, 41.419772, 2.122447). We conducted
172 randomized sampling in terms of timing (morning *versus*
173 afternoon) and across days; each location was sampled three
174 times for 3h using two Coriolis μ air samplers, respectively,
175 resulting in altogether 30 air samples (**Supplementary Table**
176 **1**).

177 Based on our preliminary tests, we further decided to use the
178 spin-column based PowerSoil Pro Kit (QIAGEN, 2018, Hilden,

179 Germany) for DNA extractions, using 30 μ L of elution buffer
180 (**Supplementary Information: Air sampling and DNA extraction**
181 *optimizations*). Final DNA concentration was measured on a
182 Qubit 4.0 fluorometer (Invitrogen, 2021), using the high-
183 sensitivity DNA kit and 3 μ L of DNA elution as input per sample.
184 We then used the Rapid Barcoding library preparation kit
185 (RBK114-24 V14), R10.4.1 MinION flow cells, and MinKNOW
186 by Oxford Nanopore Technologies (Oxford, UK) to nanopore
187 shotgun-sequence the extracted DNA of the air samples. During
188 library preparation, we used each barcode twice per air sample
189 to increase the DNA input per sample. For sequencing the
190 samples of the controlled and natural environment, we used one
191 R10.4.1 flow cell per sample type (i.e., for all 1h-, 3h-, or 6-
192 samples and replicates, respectively). For sequencing the
193 samples of the urban environment, we pooled all samples from
194 the Outer Belt location onto one flow cell (since they exhibited
195 the lowest DNA concentrations), and the samples of the City
196 Center and Residential Area, as well as of the Green Belt and
197 Urban Beach, onto one flow cell, respectively. The sequencing
198 parameters included a minimum read length of 20 bases, a
199 translocation speed of 400 bases per second, and each
200 sequencing run lasted 24 hours. As we used MinKNOW
201 v23.04.3 for the controlled and natural environment, this
202 sequencing data was generated at a signal measurement
203 frequency of 4 kHz, whereas we used the updated MinKNOW
204 v23.04.5 for the urban environment, which generated
205 sequencing data at 5 kHz.

206 We included negative controls along our entire protocol to
207 identify contamination of the low-biomass air samples. For
208 sampling negative controls, we treated one liquid impinger cone
209 per sampling event the same way that we treated the actual
210 sampling cone, but we only left them in the impinger for a few
211 minutes and did not actively sample air. For the urban
212 environment, negative sampling controls were collected once
213 per sampling day and sampling location. For DNA extraction and
214 sequencing negative controls, we included one sample of 700
215 μ L nuclease-free water (Thermo Fisher Scientific) per DNA
216 extraction and one sample of 20 μ L nuclease-free water per
217 library preparation, respectively. We barcoded all negative
218 controls, i.e. sampling, extraction, and sequencing controls, and
219 included them in the same sequencing library as the respective
220 control samples. We further subjected a positive control of five
221 Gram-positive bacteria, three Gram-negative bacteria, and two
222 fungal species (ZymoBIOMICS Microbial Community Standard,
223 D6300) to our DNA extraction and sequencing protocols to
224 assess any potential biases. The positive control was
225 sequenced on a separate flow cell since the high DNA
226 concentration would have outcompeted the low-biomass air
227 samples.

228 We next used Guppy v6.3.2 (r10.4.1_e8.2_400bps_hac; [17]) in
229 high-accuracy (HAC) mode for basecalling the controlled and
230 natural environment samples, and Dorado v4.3.0
231 (dna_r10.4.1_e8.2_400bps_hac@v4.3.0; [18]) for HAC-
232 basecalling of the urban environment samples. We only

233 processed the data that had passed internal data quality
234 thresholds during sequencing (“passed” sequencing reads).
235 Porechop v0.2.3 [19] was used for removing sequencing
236 adapters and barcodes, and Nanofilt v2.8.0 [20] was applied for
237 filtering reads at a minimum average quality score of 8 and a
238 minimum length of 100 bases for all samples. We then used
239 Kraken2 v2.0.7 [21] with the NCBI nt database (access
240 29.01.2023) for taxonomic classification across all samples, and
241 downsampled them to a specific read count for comparable
242 taxonomic assessments across samples of one sample type: 5k
243 reads for 1h-samples from the controlled environment, 15k
244 reads for the 3h-samples from the controlled environment, 70k
245 reads for the natural environment samples, and 30k reads for
246 the urban environment samples. We performed Principal
247 Coordinate Analysis (PCoA) on the relative abundances of the
248 genera identified in the urban environment samples, which were
249 downsampled to 30k read, using Python v3.9 with Pandas
250 v1.3.3, NumPy v1.21.2, scikit-learn v0.24.2, scikit-bio v0.5.6,
251 SciPy v1.7.1, and Matplotlib v3.5.2.. The 20 most abundant
252 microbial genera at a minimum relative abundance of 1% as well
253 as the PCoA were visualized using matplotlib v3.5.2 in Python
254 v3.9. We additionally benchmarked several additional
255 bioinformatic analysis tool in application to the controlled and
256 natural environment samples, including DIAMOND BLASTX
257 [22] for protein-based taxonomic classifications and the Chan-
258 Zuckerberg (CZID) computational pipeline [23] for hybrid

259 taxonomic classifications (i.e., as a combination of read- and
260 contig-based classification).

261 We generated *de novo* assemblies using metaflye v2.9.1 [24],
262 followed by polishing with minimap2 v2.17 [25] and three rounds
263 of Racon v1.5 [26]. The resulting contigs were then binned into
264 Metagenome-Assembled Genomes (MAGs) using metaWRAP
265 v1.3 [27], which integrates the output of various binning tools.
266 The MAGs were refined and quality-checked using CheckM
267 v1.2.2 [28]. We only maintained MAGs at minimum
268 completeness of 30% and maximum contamination of 10%. For
269 the urban microbiome dataset, we pooled across all samples per
270 sampling location to maximize the number of reads before
271 binning. We finally applied functional annotation to our
272 metagenomic dataset to assess the presence of general
273 metabolic pathways and ecosystem functions (**Supplementary**
274 **Information: Functional annotation**); to identify antimicrobial
275 resistance and virulence genes, we applied AMRFinderPlus
276 v3.12.8 [29] and ABRicate v1.0.1 [30] to the reads, contigs, and
277 bins; for the application to the read level, we converted the fastq
278 files to fasta files using seqkit v2.8.2 [31].

279 To obtain information about the anthropogenic impact on the
280 different urban sampling locations, we obtained remote sensing
281 data (Sentinel-2 L1C orthoimage products from October 24th
282 2023) that provides top-of-atmosphere reflectance, which we
283 used to classify the city of Barcelona into Local Climate Zones
284 (LCZs) on based ten bands with 10 m and 20 m ground

285 sampling distances [32]. We further used the portable aerosol
286 spectrometer Dust Decoder 11-D (GRIMM Aerosol Technik
287 GmbH, Germany) to monitor particle mass fractions (TSP, PM₁₀,
288 and PM_{2.5}; TSP=total suspended particles; PM=particulate
289 matter) as well as temperature and relative humidity
290 measurements in 1-minute intervals during each sampling
291 event. We then summarized and analyzed the resulting data
292 using Python v3.9 and SciPy v1.13.0: We applied the Kruskal-
293 Wallis and post-hoc Dunn's tests to identify significant
294 environmental differences between locations, and conducted
295 regression analyses to assess correlations between particle
296 mass fractions and microbial diversity indices (Shannon,
297 Simpson, and richness of microbial genera).

298

299 **Results**

300 After confirming that Coriolis μ liquid impingement resulted in
301 sufficient high-quality DNA yield for nanopore shotgun
302 sequencing after one hour of sampling (**Materials and**
303 **Methods; Supplementary Information: *Air sampling and DNA***
304 *extraction optimizations*), we conducted a pilot study in a
305 controlled environment to determine the robustness of the
306 metagenomic data and assess the impact of sampling duration
307 (**Materials and Methods**). For the 1h-samples, DNA yields
308 ranged from 17.7 ng to 50.7 ng (0.98 ng/m³ to 2.82 ng/m³), while
309 the 3-hour samples showed DNA yields ranging from 130.2 ng
310 to 179.4 ng (2.41 ng/m³ to 3.32 ng/m³; **Supplementary Table 1**;

311 *pilot_study* sheet). Nanopore shotgun sequencing delivered
312 between 7k and 60k high-quality sequencing read at a median
313 read length of 896 bases (**Figure 1A**), respectively, of which 5k
314 to 35k reads were successfully mapped to the taxonomic genus
315 level using Kraken2 and the NCBI nt database (**Figure 1B-C**;
316 **Supplementary Table 1**; *pilot_study* sheet). After
317 downsampling to the same number of reads per sample type
318 (1h- and 3h-samples, respectively), the taxonomic composition
319 of the 20 most abundant taxa indicated that only the 3-hour
320 sampling duration captured a stable "core" air microbiome
321 across days at the genus level (**Figure 1D-E**). These
322 assessments were consistent for protein-level or hybrid read-
323 and assembly-based methods, both at the taxonomic phylum
324 and genus level (**Supplementary Figures 1-2**). The most
325 abundant genera included soil- and plant-associated bacteria
326 such as *Bradyrhizobium*, *Paracoccus*, *Nocardioides*, *Massilia*,
327 and *Streptomyces* (**Figure 1D-E**; **Materials and Methods**).

328 Based on these results, we conducted a pilot study in a natural
329 environment over six days; we sampled air for either 3h or 6h,
330 assuming that the natural environment might show more
331 variability than the controlled environment and require longer
332 sampling duration. Briefly, while the extended sampling duration
333 increased total DNA yield, it did not consistently increase the
334 amount of biomass per cubic meter of sampled air, suggesting
335 diminishing returns in efficiency with longer durations
336 (**Supplementary Table 1**; *pilot_study* sheet). Nanopore
337 shotgun sequencing resulted 130k to 200k high-quality

338 sequencing reads at a slightly higher median read length than
339 the controlled environment of 1481 (**Figure 1F**), of which 70k to
340 140k reads were successfully mapped to the taxonomic genus
341 level (**Figure 1G-H; Supplementary Table 1; pilot_study**
342 **sheet**). After downsampling all samples to 70k reads, analysis
343 of the relative abundance of 20 most abundant taxa revealed a
344 very similar profile for both 3-hour and 6-hour samples. The
345 taxonomic assignments were again consistent across protein-
346 level or hybrid read- and assembly-based methods, both at the
347 taxonomic phylum and genus level (**Supplementary Figures 1-**
348 **2**). A distinct air microbiome profile was observed in the natural
349 environment in comparison to the controlled settings, with high
350 predominance of *Pseudomonas* and unique detection of
351 microbial taxa such as *Actinoplanes*, *Amycolatopsis*, *Dugnaella*,
352 *Flavobacterium*, *Nocardia*, *Rhodococcus*, and *Variovorax*
353 (**Figure 1I-J; Materials and Methods**).

354 All negative controls resulted in low DNA yields (of <0.1 ng) from
355 typical contaminant species such as *Escherichia*, *Salmonella*,
356 *Shigella*, *Francisella*, and *Pseudomonas* (**Supplementary**
357 **Figure 3A-B; Material and Methods**) [33]. This demonstrates
358 that no external contamination had influenced our assessment
359 of air as a low-biomass ecosystem, thus underscoring the
360 reliability of the presented results. The application of our
361 protocol to a well-defined mock community further showed that
362 all bacterial and fungal species could be detected with
363 approximately correct abundance estimates. Although the
364 fungal taxa and Gram-positive *Bacillus subtilis*, in particular,

365 were underrepresented (**Supplementary Figure 3C; Material**
366 **and Methods**).

367 We finally applied our optimized laboratory and computational
368 approaches to assess an exemplary urban microbiome using
369 nanopore metagenomics (**Figure 2A; left, Materials and**
370 **Methods**). Our remote-sensing-based LCZ classification
371 (**Figure 2A; right**) indicated that most of our sampling locations
372 (City Center, Residential Area, and Urban Beach) were of the
373 compact low-rise category, a typical feature of central urban
374 environments. The Outer Belt location was classified as
375 compact mid-rise category, which features taller buildings at the
376 outskirts of the city. The Green Belt location was classified as
377 scattered trees category, featuring more natural elements. In
378 terms of air pollution assessed through particle mass fractions
379 (**Supplementary Table 2; Materials and Methods**), we found
380 significant differences in TSP, PM₁₀ and PM_{2.5}, between our
381 sampling locations (**Supplementary Figure 4**). The total air
382 pollution measured by TSP was highest in the three compact
383 low-rise sampling locations, while TSP was lowest in the Outer
384 Belt. The relatively medium levels of TSP in the Green Belt were
385 dominated by relatively high levels of PM₁₀ (**Supplementary**
386 **Figure 4**).

387 Nanopore shotgun sequencing delivered between 33k and 422k
388 high-quality sequencing read at a median read length of
389 between 598 and 2358 bases (**Figure 2B**), respectively, of
390 which 21k to 312k reads were successfully mapped to the

391 taxonomic genus level using Kraken2 and the NCBI nt database
392 (**Figure 2C; Supplementary Table 1**; *urban_study* sheet). The
393 City Center exhibited the longest DNA fragments, and the Outer
394 Belt location the shortest DNA fragments (**Figure 2B**). The
395 relatively high fragmentation in the Outer Belt coincided with
396 generally low DNA yields across all the location's samples and
397 replicates (**Supplementary Table 1**; *urban_study* sheet).

398 For taxonomic comparisons across replicates and samples, we
399 again downsampled the number of reads (here to 30k reads per
400 sample) and compared the relative distribution of the 20 most
401 abundant microbial genera per location at a minimum relative
402 abundance cutoff of 1% displaying (**Materials and Methods**).

403 We observed that the microbial compositions were highly
404 location-specific across all six samples per location, including
405 across the three randomized sampling events and the two
406 respective sampling replicates (**Figure 2D; Materials and**
407 **Methods**). The core urban air microbiome consisted of
408 microbial genera such as *Streptomyces*, *Sphingomonas*,
409 *Pseudomonas*, *Nocardioides*, and *Microbacterium*, which were
410 detected across all samples. Specifically the Green Belt was
411 characterized by the presence of several unique taxa such as
412 *Rubrobacter*, *Gemmatirosa*, *Capillimicrobium*, and
413 *Amycolatopsis*, whereas dominant "urban" taxa such as
414 *Paracoccus*, *Kocuria*, *Deinococcus*, and *Cellulomonas* were not
415 detected at all (**Figure 2D**). Principal Coordinate Analysis
416 (PCoA) clearly distinguishes the five different urban locations,
417 with the first PCoA axis separating the Green Belt and City

418 Center locations from the remaining ones; the second PCoA
419 axis then further delineates the individual sampling locations
420 (**Figure 2E**).

421 Despite the location-specific differences in air microbial
422 composition (**Figure 2A**), in LCZ-based land usage (**Figure 2D**)
423 and in air pollution measured by particle mass fractions
424 (**Supplementary Figure 4**), we found no significant correlations
425 between any environmental variable and microbial diversity
426 measurements (**Materials and Methods**).

427 To next obtain as highly contiguous *de novo* genome
428 assemblies as possible, we pooled all samples per location
429 before contig assembly and binning (**Materials and Methods**).

430 Taxonomic classification of these bins showed that only the
431 most abundant taxa could be assembled (**Table 1**). Functional
432 annotation of the reads, contigs, and bins detected typical
433 microbial metabolic functions (**Supplementary Information:**
434 *Functional annotation*). We next focused on the annotation of
435 antimicrobial resistance and virulence genes with potential
436 human health consequences (**Supplementary Table 3;**
437 **Materials and Methods**). One of the most frequently detected
438 genes was the *VanR-O* gene, which is responsible for
439 vancomycin resistance. When comparing resistance gene
440 prevalence across urban locations, the Urban Beach location
441 exhibited the highest density of resistance genes; the *blaCARB-*
442 *8* and *blaCARB-16* genes, which confer beta-lactam resistance,
443 and the *blaOXA-17* gene, which confers oxacillin resistance,

444 were detected at the read level. Additionally, *blaL1*, which
445 confers to a broad range of beta-lactam antibiotics, the *blaOXY*
446 gene, which confers oxacillin resistance, and the *blaPSZ* gene,
447 which confers resistance to penicillins and cephalosporins, were
448 identified at the contig level (**Supplementary Table 3**).

449 **Discussion**

450 Metagenomic approaches have provided unprecedented
451 insights into the nature, origin, and complexity of the air
452 microbiome [4, 6, 7, 34]. While past studies have relied on
453 traditional short-read sequencing, we here describe the first
454 long-read nanopore sequencing technology-based approaches
455 to robustly assess the air microbiome. Although nanopore
456 sequencing has been applied to various environmental
457 samples, such as water and soil [15, 35, 36], its applicability to
458 air samples was expected to pose a particular challenge due to
459 the ultra-low biomass of air and the amplification-free nature of
460 nanopore sequencing [5]. We here showed that nanopore
461 shotgun sequencing in combination with active air sampling
462 through liquid impingement and tailored computational analyses
463 can reproducibly describe the air microbiome of different
464 environments (**Figure 1**) while leveraging the latest nanopore
465 chemistry improvements which offer high sequencing accuracy
466 and reduced minimum DNA input requirements [10, 11].

467 We further showed that only three hours of active air sampling
468 resulted in robust air microbiome assessments in a controlled
469 and natural environment, with consecutive application of our

470 laboratory and computational approaches to the urban air
471 microbiome in Barcelona, Spain, revealing surprisingly stable
472 location-specific signatures of microbial composition and
473 diversity (**Figure 2**). These stable signatures could importantly
474 be identified across replicates (using two air samplers per
475 sampling event) and despite stringent randomization across
476 sampling days and morning and afternoon sampling events.
477 Several microbial taxa such as *Sphingomonas* and
478 *Streptomyces*, which are known for their evolutionary
479 adaptability, were nevertheless present in all air microbiomes,
480 and could potentially be part of the stable air microbiome of this
481 urban environment. Ordination of the taxonomic composition
482 was able to capture the majority of variance in this
483 multidimensional data (>80%; **Figure 2E**) and nicely visualizes
484 the distinct clusters that separate each urban location and
485 specifically the Green Belt and City Center locations from the
486 remaining ones. The relative similarity of Green Belt and City
487 Center samples might be attributable to the phenomenon of
488 orographic uplift, where air masses ascend from lower regions
489 (here the Barcelona City Center) to higher elevated areas (here
490 the closeby Green Belt). As a result of this upward movement,
491 certain airborne particles and microorganisms might have been
492 transported from the City Center to the Green Belt location [37,
493 38].

494 The individual samples of the Green Belt location cluster
495 together most tightly (**Figure 2E**). be because of several
496 microbial taxa that were uniquely detected at this location, which

497 represents the only natural environment in our study according
498 to our remote-sensing-based assessments; those unique taxa
499 are known to be associated with soil or have been frequently
500 found in forests and green spaces [39]. Besides this finding, we
501 however found no evidence of correlation of the urban air
502 microbiome with measurements of anthropogenic impact (as
503 assessed through the remote-sensing-based Local Climate
504 Zones, LCSz; **Figure 2A**) or of air pollution (as assessed
505 through particle mass fraction measurements; **Supplementary**
506 **Figure 4**). This might be due to complex interactions between
507 air microbiomes, as exemplified by our hypothesis of the impact
508 of orographic uplift, or because of lack of depth when describing
509 our environmental variables. For example, air pollution by TSP
510 was higher in the Green Belt than in the Outer Belt, which would
511 have not been expected according to the remote-sensing-based
512 anthropogenic impact inferences. However, these elevated
513 levels of TSP in the Green Belt might have originated from
514 natural air components such as pollen, which would require
515 more in-depth environmental monitoring to dissect.

516 The annotation of antimicrobial resistance and virulence genes
517 in our metagenomic data shows that we can use the same
518 dataset to assess potential anthropogenic impacts on microbial
519 diversity while concurrently understanding potential public
520 health consequences [40]. We detected evidence of
521 antimicrobial resistance across all sampled environments
522 (**Supplementary Table 3**), but especially the detections of
523 clinically relevant beta-lactamases such as *blaCARB-8*,

524 *blaOXA-1*, and *blaI-1*, and of genes conferring resistance to
525 other antibiotics such as carbenicillin and oxacillin [41], in
526 Barcelona's urban air microbiome underscore the possibility of
527 monitoring airborne virulence dissemination using nanopore-
528 based metagenomics.

529 Genome assembly and binning of the long nanopore reads
530 further allows us to be more confident in the presence of specific
531 microbial species and of their pathogenic potential through the
532 identification of Metagenome-Assembled Genomes (MAGs)
533 (**Table 1**). We obtained high-quality genome assemblies
534 (**Materials and Methods**) of the pathogenic species
535 *Stenotrophomonas maltophilia* and *Salmonella enterica* from
536 the urban microbiome data (**Table 1**). The *Stenotrophomonas*
537 species is known as an emerging difficult-to-treat human
538 pathogen [42] and many of the *Salmonella enterica* serovars can
539 cause disease in humans through zoonotic or foodborne
540 transmission [43]. While we require good coverage of a
541 microbial genome to create such assemblies for taxonomic
542 species or strain identification, also just the presence of
543 individual pathogen-associated sequencing reads might be
544 used for obtaining first information on the potential presence of
545 microorganisms of public health concern. For example, given
546 the presence of sequencing reads of the *Brucella* genus, an
547 animal pathogen that can affect dogs, in several of our urban air
548 samples, we further analyzed our taxonomic annotation, which
549 was based on the entire NCBI nt database, and were indeed
550 able to detect the presence of *Canis lupus familiaris* in the same

551 air samples [44]. While this might point to a potential impact of
552 animal domestication and specifically frequent dog walking in
553 Barcelona on public health [45], such complex
554 interdependencies can only be confirmed in a controlled and/or
555 experimental setting.

556 While we were able to build *de novo* assemblies from our
557 nanopore-based air metagenomic data, most of the MAGs were
558 incomplete (<30%) and/or showed high levels of contamination
559 (>10%) (**Table 1**). Given the low amount of DNA input and
560 therefore relatively small size of the resulting metagenomic
561 datasets in combination with the expectedly high fragmentation
562 of DNA in air samples, this might just be an inherent
563 shortcoming when it comes to assessing the air microbiome
564 – albeit applying long-read sequencing technology. We here
565 found a particular small median DNA fragment and sequencing
566 read length for the Outer Belt location (**Figure 2B**), which might
567 point towards the impact of environmental conditions or specific
568 taxonomic compositions (and variables such as the
569 microorganisms' genome size and cell wall composition) on the
570 final fragment and read length distribution. It is further expected
571 that non-viable microorganisms, which might significantly
572 contribute to the air microbiome, result in more fragmented DNA
573 in the air samples; this means that substantial differences in
574 read lengths between microbial taxa might also be attributed to
575 their differential viability in the air environment – a hypothesis
576 that we might be able to resolve in the future using viability-
577 resolved metagenomic approaches [46].

578 We emphasize that our sampling, laboratory and computational
579 approaches constitute one feasible and reproducible way of
580 using nanopore shotgun sequencing to profile the air
581 microbiome. While we tested some additional established air
582 sampling and DNA extraction methodology, we have not
583 conducted an extensive study of all possible approaches. We
584 specifically emphasize that the detection of fungi and Gram-
585 positive bacteria could be improved when using different sample
586 processing and DNA extraction techniques. This is also
587 reflected by the application of our approaches to a positive
588 control, which shows that fungal taxa and Gram-positive
589 *Bacillus subtilis*, in particular, were underrepresented. As
590 sturdier cell walls would require more aggressive DNA
591 extraction approaches, this would, however, also lead to
592 increased DNA fragmentation, especially in Gram-negative
593 bacteria, and therefore more difficult downstream analyses. A
594 good trade-off could be the sequencing of several, differently
595 processed DNA extracts and subsequent data pooling to assess
596 the microbial diversity of any air sample more holistically.

597 In conclusion, our study establishes a robust framework for air
598 microbiome assessments using nanopore metagenomics. We
599 envision that nanopore sequencing for air monitoring can
600 provide a basis for fast, robust, and automated characterizations
601 of the air microbiome in both urbanized and remote settings.
602 This characterization importantly extends beyond taxonomic
603 composition to include functions related to human and
604 ecosystem health, such as pathogen and drug resistance and

605 virulence gene detection, which can enhance our understanding
606 of infectious disease transmission patterns and their relationship
607 with exerted anthropogenic pressures.

608

609 **Data and code availability**

610 All raw data and MAGs have been made publicly available via
611 ENA (study accession number: PRJEB76446). All code has
612 been made publicly available via Github:
613 https://github.com/ttmgr/Air_Metagenomics.

614 **Acknowledgements**

615 **Author contributions**

616 T.R, S.P., S.B. and L.U. designed the experiment. T.R, S.P.
617 conducted the sampling, processing, and data analysis under
618 L.U.'s supervision. T.R, S.P., and L.U. wrote the manuscript with
619 the contribution of all authors who revised and approved the
620 manuscript.

621 **Financial Disclosure**

622 This study was funded by a Helmholtz Principal Investigator
623 Grant awarded to L.U.

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784

785 **List of Tables**

786 **Table 1.** *De novo* genome assembly results across all air samples from the
 787 controlled (Gh), natural (Nat) environment, and the urban microbiome dataset.
 788 Contigs were assembled and then used to identify metagenome-assemblies
 789 (MAGs), their taxonomic origin, completeness, and contamination.

Sample	# contigs (mean)	N50 contigs (mean)	# MAGs	Species	Completeness [%]	Contamination [%]
Gh1h	21	5 928	/	/	/	/
GH3h	121	15 330	2	<i>Paracoccus aerius</i>	64.59	2.94
				<i>Paracoccus denitrificans</i>	63.41	1.46
Nat3h	204	7 401	/	/	/	/
Nat6h	117	7 282	/	/	/	/
City center	1170	23 151	/	/	/	/
Residential area	470	11 098	/	/	/	/
Green belt	1171	15 215	1	<i>Burkholderia sp.</i>	36.66	2.38
Urban beach	7732	21 049	1	<i>Stenotrophomonas maltophilia</i>	48.33	6.69
Outer belt	1874	10 282	1	<i>Salmonella enterica</i>	41.72	10.59

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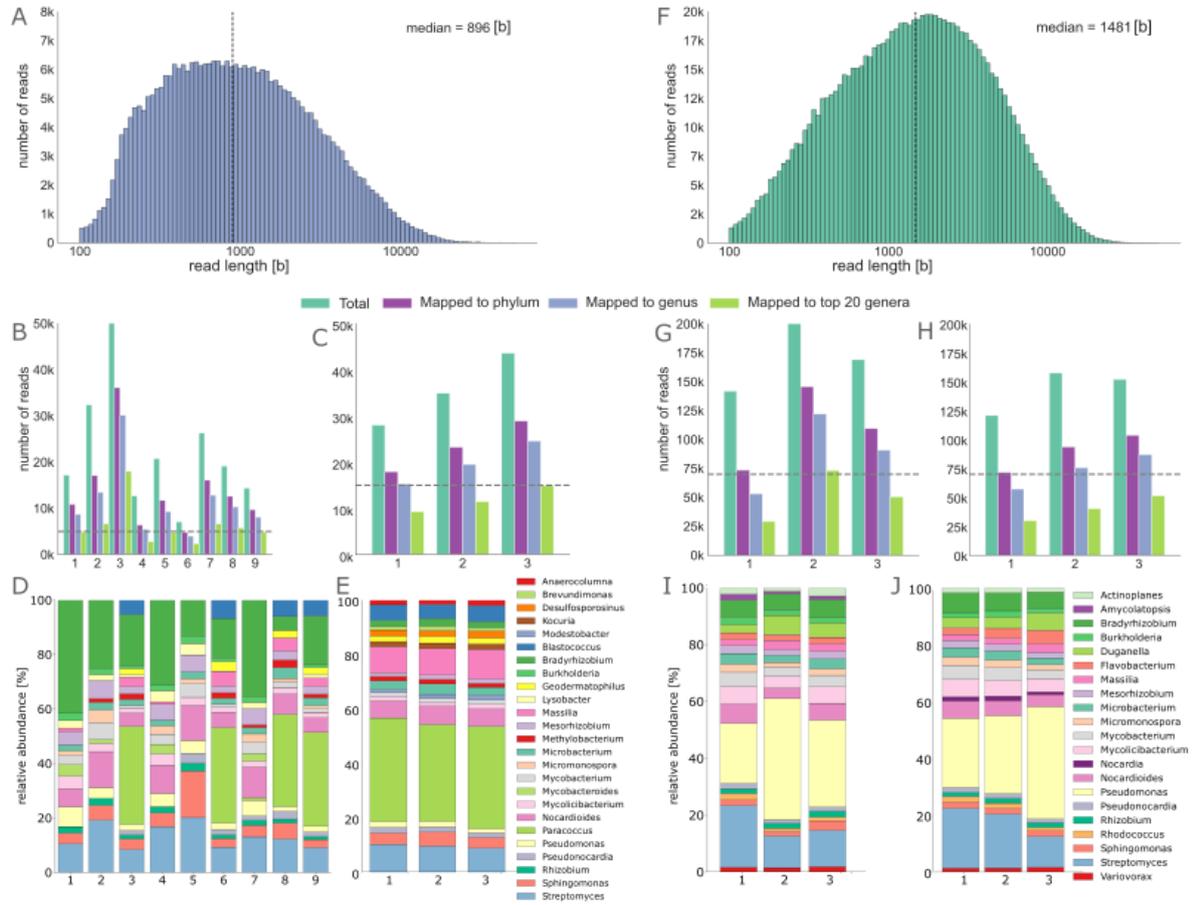
791 **List of Figures**

792 **Figure 1.** Robust air microbiome assessments of a controlled (*left, A-E*) and
793 natural (*right, F-J*) environment through nanopore shotgun sequencing. **A.**
794 Nanopore sequencing read length distribution across 1h- and 3-Gh samples.
795 **B-C.** Number of total sequencing reads, and of reads mapping to taxonomic
796 phylum and genus level as well as to the 20 most dominant genera using
797 Kraken2 (Material and methods) across the samples; the downsampling
798 threshold across samples is indicated by the dashed horizontal line.
799 Taxonomic composition of the **D.** 1h- and **E.** 3h-samples after downsampling
800 based on the 20 most dominant genera across samples. **F.** Nanopore
801 sequencing read length distribution across 3h- and 6h-Nat samples. **G-H.**
802 Number of total sequencing reads, and of reads mapping to taxonomic phylum
803 and genus level as well as to the 20 most dominant genera using Kraken2
804 across the samples; the downsampling threshold across samples is indicated
805 by the dashed horizontal line. Taxonomic composition of the **I.** 3h- and **J.** 6h-
806 samples after downsampling based on the 20 most dominant genera across
807 samples.

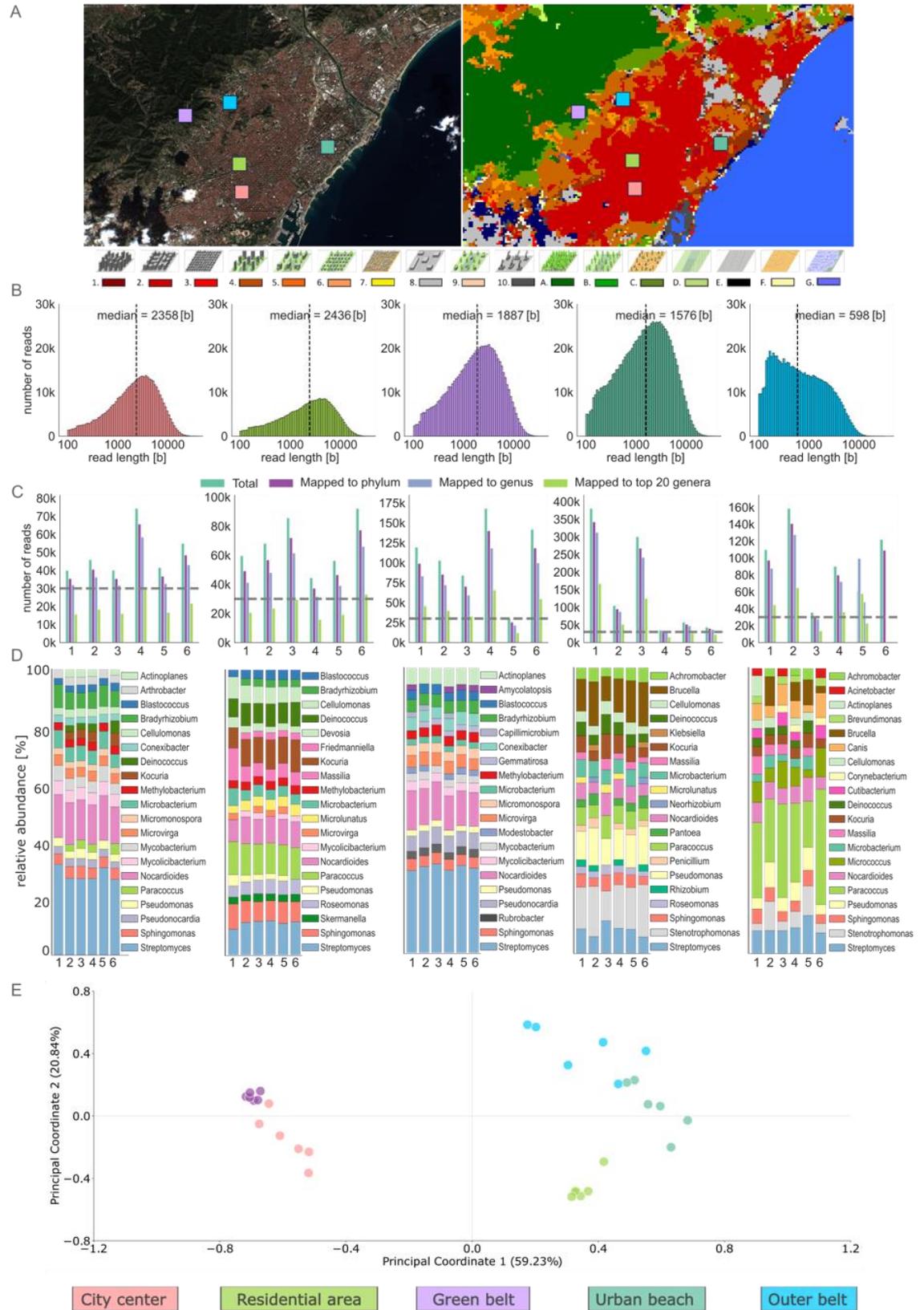
808 **Figure 2.** Local climate zone (LCZ) classification and metagenomic analysis
809 of Barcelona using satellite images and nanopore shotgun sequencing data.
810 **A.** Sentinel-2 image and LCZs classification map of Barcelona on 24.10.2023,
811 with a legend of LCZs classes at the bottom. The colored squares indicate the
812 five sampling locations. The legend at the bottom depicts various LCZs
813 represented by 3D models and their corresponding colors. (1-Compact high-
814 rise; 2-Compact mid-rise; 3-Compact low-rise; 4-Open high-rise; 5-Open mid-
815 rise; 6-Open low-rise; 7-Lightweight low-rise; 8-Large low-rise; 9-Sparsely
816 built; 10-Heavy industry; A-Dense trees; B-Scattered trees; C-Bush scrub; D-
817 Low plants; E-Bare rock or paved; F-Bare soil or sand; G-Water). **B.**
818 Histograms showing the distribution of read lengths [b] for each sampling site,
819 with the median read length indicated on the top of each histogram. **C.** Bar
820 plots displaying the number of reads mapped at various taxonomic levels for
821 each sample site, with the downsampling threshold indicated by the dashed
822 horizontal line. **D.** Relative abundance of the top 20 most abundant bacterial

823 genera at the read level, downsampled to 30K reads before taxonomic
 824 classification using Kraken2. E. Principal Coordinates Analysis (PCoA) of the
 825 relative abundances of the bacterial genera identified at the five sampling
 826 locations.

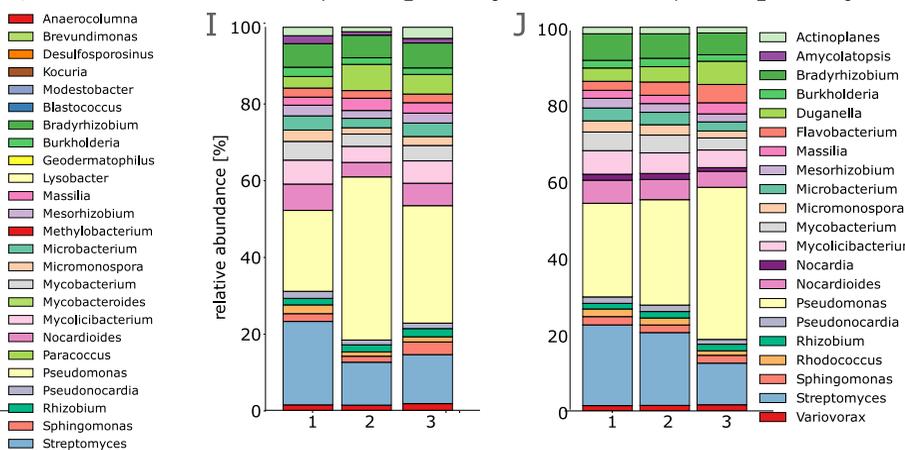
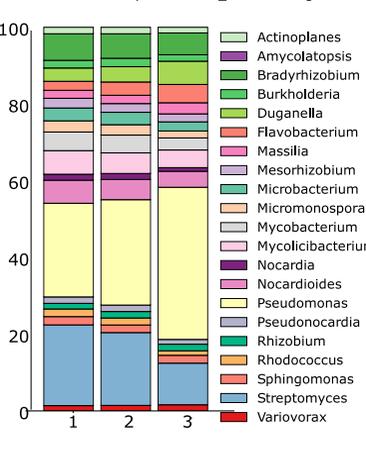
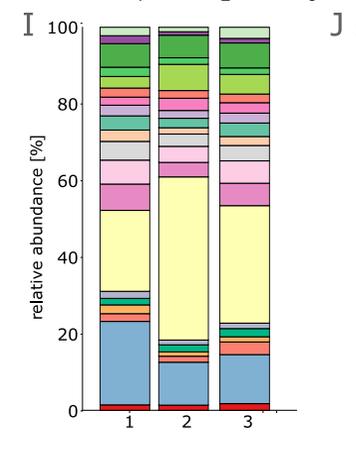
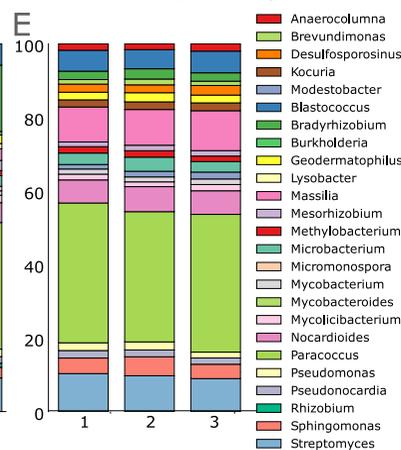
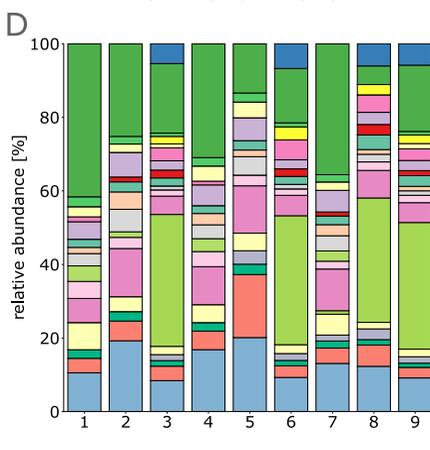
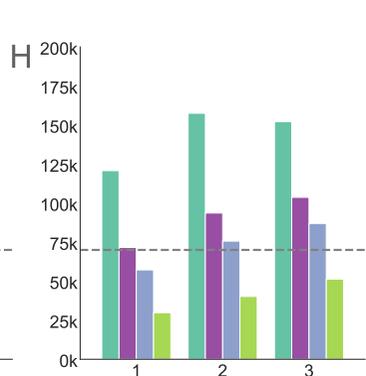
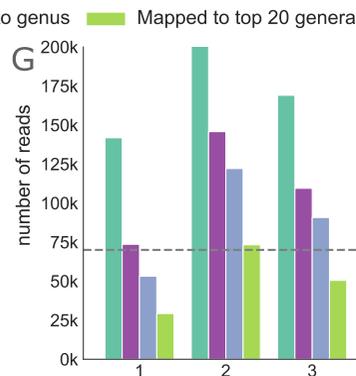
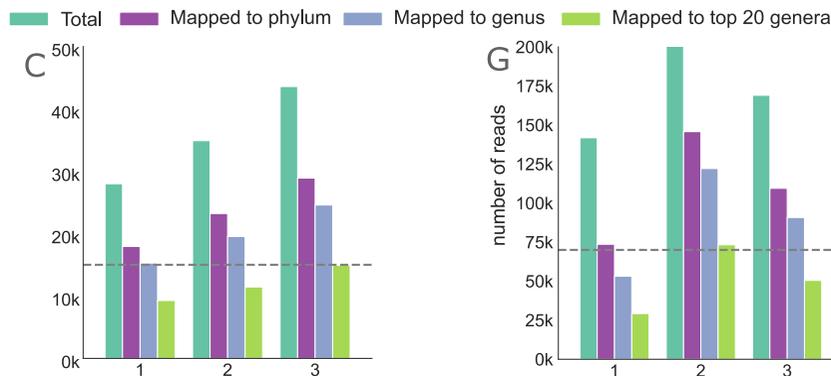
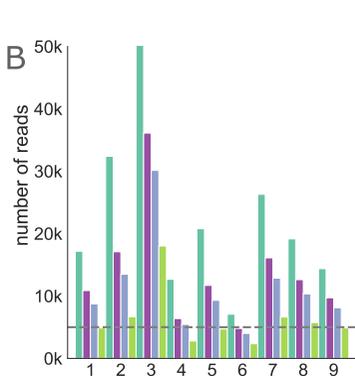
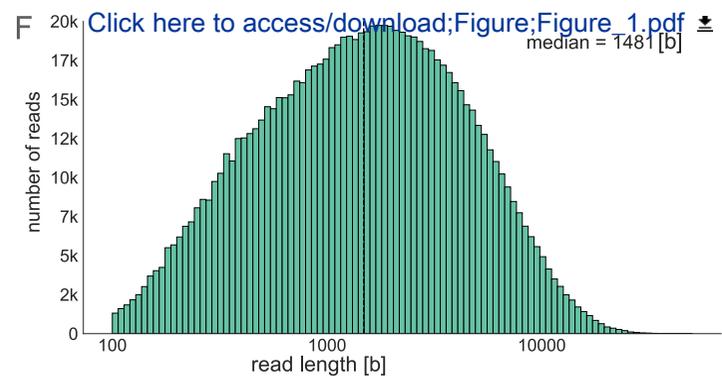
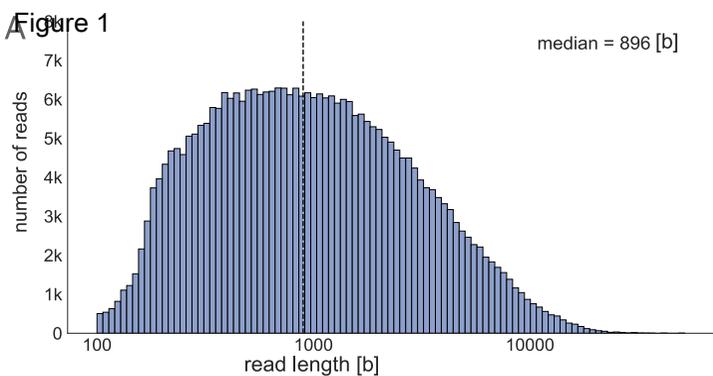
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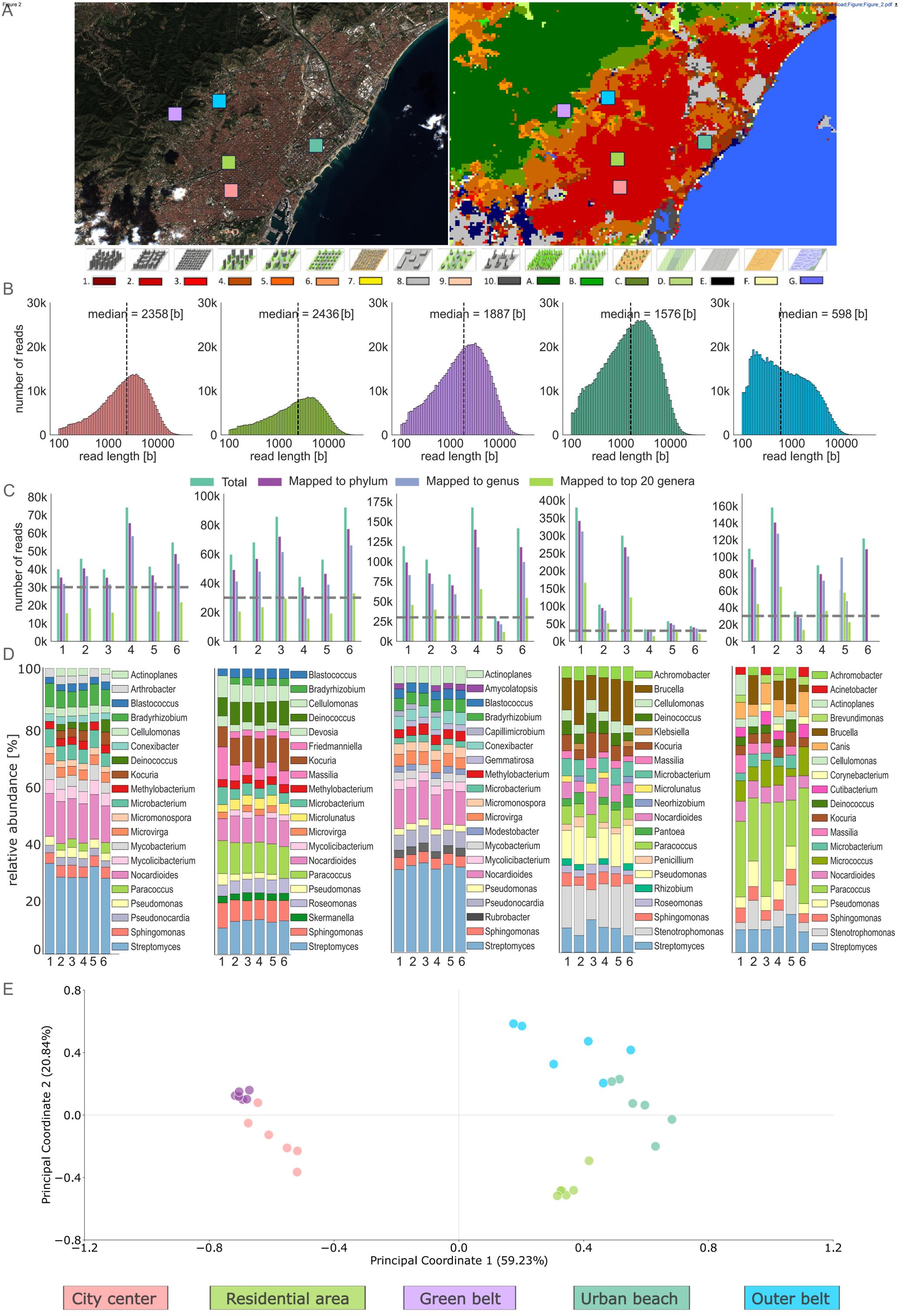


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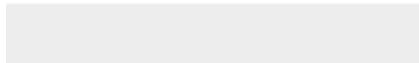
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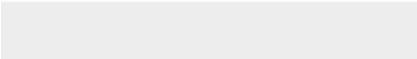
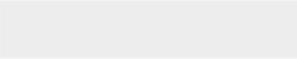


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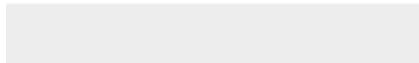


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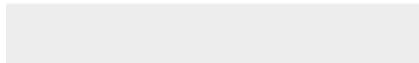


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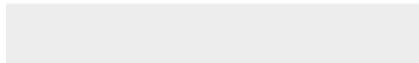


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Munich, June 14th 2024

Dear Marnix, dear Janet, dear ISME Communications Editorial Board, dear reviewers,

We thank the editors and reviewers of our manuscript **Air monitoring by nanopore sequencing** for their constructive feedback. We herewith resubmit an extensively revised version of our manuscript, which should address all of the reviewers' comments and importantly contains completely new data which we specifically generated and analyzed to demonstrate the robustness of our laboratory and computational approaches. We specifically applied our approaches to robustly characterize the urban air microbiome of the city of Barcelona, Spain, which is known to suffer from high air pollution and where one of our research centers (ISGlobal is based). We further provide more detail on our assessments of potential human health consequences (please refer to our reply to reviewer #1's 2nd comment). Please find all our detailed responses below.

Reviewer #1: Comments to the authors: This is a well-presented and performed study detailing the use of nanopore sequencing for characterising the air microbiome. It includes suitable positive and negative controls and provides a useful guide to those working in a similar field that would like to employ nanopore sequencing in their own studies. However, there are some improvements that could be made:

1. The authors refer to supplementary information when stating that several methods were tested - however, it would be useful to briefly include the outcome of method comparisons and outcomes (ie which methods were tried and which final one was selected and why) in the main text (referring to more detail in supplementary information). Note the specific section within the supplementary info needs to be stated so the reader can more easily locate the relevant section of the supplementary information. This is the case in other parts of the manuscript - a specific part of the supplementary information should be referenced - not just (Supplementary Information).

Thank you very much for your positive feedback! We followed your suggestion and now provide a short overview of the methods and results that we have so far only mentioned in the Supplementary Information. We now specifically have two text sections in the Supplementary Information, (1) "Air sampling and DNA extraction optimizations" (p. 5, ll 56-86.) and (2) "Functional annotation" (p. 6-7, ll 86-121) which provide some information for interested researchers, but which (1) did not result in enough data to make robust conclusions, or (2) allowed us to make hypotheses about microbial ecosystem functions which we will, however, not be able to follow up in this study.

Materials and Methods:

p. 5, ll. 115-124:

"We first conducted preliminary tests to compare standard air sampling and DNA extraction approaches for nanopore sequencing-based air metagenomics; this included the testing of standard quartz filter- and liquid impingement-based air samplers and the optimization of respective DNA extraction approaches for subsequent nanopore shotgun sequencing, which relies on minimum DNA

input without nucleotide amplification and is sensitive to native DNA contamination (Supplementary Information: Air sampling and DNA extraction optimizations).”

p. 11, ll. 271-274

“We finally applied functional annotation to our metagenomic dataset to assess the presence of general metabolic pathways and ecosystem functions (Supplementary Information: Functional annotation);”

Results:

p. 12, ll. 300-304

“After confirming that Coriolis μ liquid impingement resulted in sufficient high-quality DNA yield for nanopore shotgun sequencing after one hour of sampling (Materials and Methods; Supplementary Information: Air sampling and DNA extraction optimizations),”

p. 17, ll. 431-434

“Functional annotation of the reads, contigs, and bins detected typical microbial metabolic functions (Supplementary Information: Functional annotation).”

2. The authors perform functional annotation, but do not really explore these findings within the context of the air microbiome. Table 1 is referred to as having information on putative biodegradation mechanisms - but no such information is clearly included - above the genes identified. It would be useful for the reader to include brief descriptions of the functional roles of these genes/groups of genes in this or another table. While the authors do elaborate on the functional roles of some of the identified genes in the text, more could be included on how these are related to functions of the air microbiome and in particular putative biodegradation mechanisms that they highlight numerous times. They particularly state that this method can be used to describe functions related to/evaluate human and environmental health consequences, and targeted applications such as infectious disease transmission. However, they do not go into any examples here from their own results.

Expanding on this would improve the impact of the paper, really demonstrating the use and benefits of the functional annotation.

Thanks to your feedback, we discussed the general gene annotation results in more detail, especially with the microbiologist (Michael Schloter) on our team, and we concluded that our statements on biodegradation functions can only be of hypothetical nature and that we currently don't have means to follow up on them experimentally. As the statements on such microbial functions were not the focus of the manuscript, we now shifted them to the Supplementary Information (p. 6-7, ll. 87-121), where we discuss findings in our new urban air microbiome in a more suggestive manner.

Importantly, we also analyzed our data for evidence of human and environmental consequences. First, we now importantly include “real-world” microbial data by monitoring the urban air in Barcelona, Spain (p. 7, ll. 164-176); this allows us to assess if we can indeed find relevant pathogens and

antimicrobial resistance and virulence genes in public spaces, which would be relevant for public health assessments.

We next included Supplementary Table 3 that lists all the resistance and virulence genes that we detected in all our air samples. We further included the input of a medical doctor (Ela Sauerborn) who pinpointed resistance genes that are known to have human health consequences; she specifically found many beta-lactamases:

p. 20-21, ll 516-528.

“The annotation of antimicrobial resistance and virulence genes in our metagenomic data shows that we can use the same dataset to assess potential anthropogenic impacts on microbial diversity while concurrently understanding potential public health consequences. We detected evidence of antimicrobial resistance across all sampled environments (Supplementary Table 3), but especially the detections of clinically relevant beta-lactamases such as blaCARB-8, blaOXA-1, and blaI-1, and of genes conferring resistance to other antibiotics such as carbenicillin and oxacillin, in Barcelona’s urban air microbiome underscore the possibility of monitoring airborne virulence dissemination using nanopore-based metagenomics.”

We next focused on the detection of specific pathogenic taxa. We hereby obtained high-quality assemblies of *Stenotrophomonas maltophilia* and *Salmonella enterica*, which were both detected in the urban air microbiome (Table 1).

p. 21-22, ll 529-555.

“Genome assembly and binning of the long nanopore reads further allows us to be more confident in the presence of specific microbial species and of their pathogenic potential through the identification of Metagenome-Assembled Genomes (MAGs) (Table 1). We obtained high-quality genome assemblies (Materials and Methods) of the pathogenic species *Stenotrophomonas maltophilia* and *Salmonella enterica* from the urban microbiome data (Table 1). The *Stenotrophomonas* species is known as an emerging difficult-to-treat human pathogen, and many of the *Salmonella enterica* serovars can cause disease in humans through zoonotic or foodborne transmission. While we require good coverage of a microbial genome to create such assemblies for taxonomic species or strain identification, also just the presence of individual pathogen-associated sequencing reads might be used for obtaining first information on the potential presence of microorganisms of public health concern. For example, given the presence of sequencing reads of the *Brucella* genus, a typical canine pathogen, in several of our urban air samples, we further analyzed our taxonomic annotation, which was based on the entire NCBI nt database, and were indeed able to detect the presence of *Canis lupus familiaris* in the same air samples. While this might point to a potential impact of animal domestication and specifically frequent dog walking in Barcelona on public health, such interdependencies would have to be investigated in a controlled and/or experimental setting.”

3. The authors could also comment on how this form of sequencing compares to others in terms of detecting microbes within air.

As we now discuss in more detail in the Introduction, the most commonly used method for characterizing air microbiomes is the amplicon-based approach followed by short-read Illumina

sequencing (p. 4, ll. 78-83), which copes better with the ultra-low biomass nature of the air but introduces well-known biases (amplification bias, database completeness, taxonomic resolution, etc.). We now also discuss the advantage of metagenomics and long-read shotgun sequencing in comparison to such amplicon-based approaches in more detail the Introduction (p. 4, ll. 83-93).

4. The authors refer to the "stable core microbiome" and that at least 3hour sampling was required to detect a "stable core microbiome". This suggests to the non-expert reader that a 3h plus sampling time is the best to use for consistency. However, it is likely that the microbiome constantly changes at local levels (within the breathing range of humans) - perhaps on a minute by minute basis - and so sampling for longer periods of time reduces temporal resolution - which may be important for certain environments, particularly when measuring/analysing anthropogenic sources of bioaerosols. It would be interesting to know the detection limits for this method - ie the shortest sampling time that still generates a robust signal, and indeed if/how the microbiome changes over short periods of time. The authors themselves state how the air microbiome can have diurnal patterns. I appreciate the more stable results demonstrate the reproducibility of the method - but the reader/potential user of such a technique may need to consider what sampling time is appropriate to them depending on their research question. This could be clarified in the manuscript.

Thank you for this important input. Our maximal temporal resolution in this study was 1h since even when sampling for an hour, we still obtained very little DNA, which was however just sufficient to create interpretable nanopore sequencing data. We showed in our pilot study that 1h-sampling resulted in a stable microbiome across sampling events per day and sampling events across days in a controlled environment (greenhouse; Figure 1D), but that the across-day 3h-sampling results seemed to be even more stable (Figure 1E). We subsequently showed that 3h-samples are also sufficient to capture similar microbial signatures across days in a natural environment (i.e. outdoors; Figure 1I; Figure 1J for 6h-samples).

Having said all this, we now most importantly include real-time data from different locations in Barcelona to describe the city's air microbiome. Here, we show that – across two technical replicates and randomized across days over a time period of two weeks and randomized between morning and afternoon sampling – we can still assess highly location-specific air microbiomes using 3h-sampling and our approaches (Figure 2D). As such robust assessments of air microbiomes was our primary goal here – to subsequently think about human and environmental consequences that we might draw from such data –, we believe that the longer sampling time is warranted – while we absolutely agree that these assessments accumulate across any smaller-scale variability, which we are not interested in this case and would also not be able to measure given minimum DNA inputs.

5. Minor comments:

Abstract - "we here show" to "Here we show"

Response: Thank you, corrected. (p. 2, ll. 30-33)

Abstract - "assemblies from the long sequencing" - remove "the"

Response: This section is not anymore present in the new version of our manuscript.

Intro - "can create such long sequencing reads" to "can sequence long reads"

Response: This section is not anymore present in the new version of our manuscript.

Intro - "that would allow to leverage" to "that would leverage"

Response: This section is not anymore present in the new version of our manuscript.

Intro - "to allow for robust" to "to enable robust"

Response: Thank you, corrected. (p. 5, ll. 100-102)

Intro - "air, and concurrently provides" to "air, providing"

Response: Could you please specify where you found it?

Intro "genomes annotations" to "genome annotations"

Response: Could you please specify where you found it?

Figure 1 legend - remove "across the C. 1h- and D. 3-Gh samples" or change to just "across samples"

Response: Thank you, corrected. (p. 33, ll. 795-799)

Figure 1 legend - remove "across the I. 3h- and J. 6-Gh samples" or change to just "across samples"

Response: Thank you, corrected. (p. 33, ll. 801-805)

Results and discussion - "while especially the fungal taxa and Gram-positive Bacillus subtilis were underrepresented by our genomic data" change to "although the fungal taxa and Gram-positive Bacillus subtilis, in particular, were underrepresented...". End sentence after "our genomic data" and start new sentence with "This might reflect"

Response: Thank you, corrected. (p. 14-15, ll. 363-366)

Within methods - authors refer to supplementary information - this needs to specify which part of supplementary info they are referring to specifically. There are also a couple of examples of this throughout the manuscript - ensure when referring to supplementary info - it is clear which specific section

Response: Thank you, we improved the text and reference to the methods and supplementary information accordingly.

Reviewer #2:

This study is well conducted and written up, I have no hesitation in recommending it for publication with a few minor corrections and queries to the authors. The application of shotgun sequencing using the nanopore platform is novel and challenging. In addition to the data published in the main paper, there is a wealth of methodological optimization included in the supplementary information that will be extremely valuable to the field.

I Have a couple of questions for the authors:

1) Do they think fungi are underrepresented in the data? To me there are fewer fungi that I would expect. What was the proportion of reads assigned to fungi (and other groups like arthropods and plants which are also collected with Coriolois)? The plots are for the most dominant taxa only, so maybe the fungi are there but missed on the plots, can they make a plot that filters for fungi only?

Do you think there is a bias for bacteria? Is there a reason for a bias towards bacteria (extraction, bioinformatics/QC of data?). It would have been good to have a comparison with meta-barcoding, or shotgun sequencing on an Illumina platform.

Thank you very much for your very positive feedback on our manuscript! Indeed, we believe that fungal taxa (and possible others with study cell walls) are underrepresented in our data due to choices we had to make with respect to sample processing and DNA extraction. We now detail this in much more detail in the Discussion:

p. 23 ll. 578-596

“We emphasize that our sampling, laboratory and computational approaches constitute one feasible and reproducible way of using nanopore shotgun sequencing to profile the air microbiome. While we tested some additional established air sampling and DNA extraction methodology, we have not conducted an extensive study of all possible approaches. We specifically emphasize that the detection of fungi and Gram-positive bacteria could be improved when using different sample processing and DNA extraction techniques. This is also reflected by the application of our approaches to a positive control, which shows that fungal taxa and Gram-positive *Bacillus subtilis*, in particular, were underrepresented. As sturdier cell walls would require more aggressive DNA extraction approaches, this would, however, also lead to increased DNA fragmentation, especially in Gram-negative bacteria, and therefore more difficult downstream analyses. A good trade-off could be the sequencing of several, differently processed DNA extracts and subsequent data pooling to assess the microbial diversity of any air sample more holistically.”

We here importantly emphasize the results of the application of our protocols to a mock community that contains fungal as well as Gram-positive and -negative taxa. We believe that the results of this mock community analyses very well reflect the biases in our data, and discuss a trade-off for more holistic microbial assessments in the future. As other approaches lead to their own biases in similar mock communities (e.g. strong overamplification of Enterobacteriaceae via 16S rRNA-based metabarcoding approaches, see e.g. Urban et al. (2021), eLife: <https://elifesciences.org/articles/61504#fig2>), we believe that the applications of our approaches to

the mock community are most meaningful to point towards certain biases in our data. With respect to direct comparisons between Illumina- and nanopore-based shotgun sequencing, we believe that past research, also on mock communities (e.g., Liu et al. (2022) BMC Microbiome: <https://microbiomejournal.biomedcentral.com/articles/10.1186/s40168-022-01415-8>), has already shown the improved performance of nanopore sequencing for certain metagenomic approaches.

2) The MAGs have low completeness (table 1 max is ~65%). Contamination does not seem to be a problem (indeed looks very good in comparison). Still but the best MAGs are "medium" quality (<https://doi.org/10.1038/nbt.3893>) and I guess the majority you found are very poor and therefore not in the table at all?

Why do you think this is? Too little coverage? is more biomass needed? Where the reads too fragmented, as the bioinformatics seem robust to me. Contamination is low maybe because of the long reads. Would greater completion of the MAGs be a benefit of shotgun sequencing on Illumina here? Please comment on this as a possible limitation of long- v short-read shotgun methods in the discussion.

Thank you; we now discuss this in much more detail in the Discussion as well (please see below). However, we would not expect any greater completion of MAGs with short-read sequencing data, but the exact opposite. Many other studies have shown that long-read sequencing data substantially improved de novo assemblies (e.g., Liu et al. (2022) BMC Microbiome: <https://microbiomejournal.biomedcentral.com/articles/10.1186/s40168-022-01415-8>).

p. 22, ll 556-577.:

“While we were able to build de novo assemblies from our nanopore-based air metagenomic data, most of the MAGs were incomplete (<30%) and/or showed high levels of contamination (>10%) (Table 1). Given the low amount of DNA input and therefore relatively small size of the resulting metagenomic datasets in combination with the expectedly high fragmentation of DNA in air samples, this might just be an inherent shortcoming when it comes to assessing the air microbiome – albeit applying long-read sequencing technology. We here found a particular small median DNA fragment and sequencing read length for the Outer Belt location (Figure 2B), which might point towards the impact of environmental conditions or specific taxonomic compositions (and variables such as the microorganisms’ genome size and cell wall composition) on the final fragment and read length distribution. It is further expected that non-viable microorganisms, which might significantly contribute to the air microbiome, result in more fragmented DNA in the air samples; this means that substantial differences in read lengths between microbial taxa might also be attributed to their differential viability in the air environment – a hypothesis that we might be able to resolve in the future using viability-resolved metagenomic approaches.”

3) Other points:

> Not clear in the abstract what natural means to me. It is a vague term, I get what you are trying to say from the context and as a contrast to a greenhouse, but maybe be more specific about your natural and controlled environments (specify them in brackets or similar) for the abstract.

Response: Thank you. The corresponding changes have been made throughout the manuscript.

> Background paragraph 2. "... due to the air's ultra-low biomass (1, 4)." The air its self cannot have low biomass, rephrase please.

Response: Thank you, corresponding changes have been made p. 3 ll. 63-65

> Background paragraph 2. Would be good to include some of the limitations of shotgun sequencing for characterizing community composition and diversity, particularly when paired with long-read. For example, bias to taxa with large genomes, sequencing of non-target DNA (human etc.) taking up read depth and reducing coverage, lack of standardized unit of diversity (e.g. 16s gene) for robust estimation of alpha-diversity.

Response: Thank you for your comment. We have incorporated this information into the introduction, highlighting the limitations of shotgun sequencing line p. 4 ll. 78-83.

> Will the accession numbers for data be provided, and will the draft MAG genomes be available also? accession number

Response: The data will be available under the accession number: PRJEB76446 including the draft MAG genomes.

Reviewer #3:

In this letter article, Tim et al. claimed to establish an experimental and bioinformatic analysis protocol for monitoring air microbiome with metagenomic approach using nanopore sequencing. However, the article's focus is unclear, as neither sample handling nor bioinformatics analysis is elaborated upon in depth. The sampling method appears to be a standard approach for aerosol collection, not specifically optimized to match the rapid sequencing capabilities of nanopore technology. Additionally, it does not address the challenge of low biomass in aerosol samples, which is a critical factor challenging effective nanopore library preparation for aerosol samples. Moreover, the paper does not present any novel analytical algorithms, nor does it attempt to streamline existing tools into a user-friendly package or enhance databases. From their data, I infer a precarious sequencing endeavor characterized by exceedingly short reads and low throughput. Indeed, long reads can be annotated down to the species level and aid in metagenome-assembled genome (MAG) recovery, but these are general advantages of long-read sequencing, not unique to aerosol samples. The article fails to address the real challenges in aerosol microbiome monitoring, such as developing protocols to collect sufficient biomass or methods to enrich DNA from low-biomass samples. In summary, the research presented lacks innovation and is unlikely to make a significant contribution to the advancement of the field.

Thank you for taking the time to read our manuscript. We, however, have to disagree with the reviewer's basic assumption that our newly established laboratory and computational approaches are not novel. We believe that reviewer #1's and #2's comments on the other hand, show how we have established a completely novel end-to-end protocol to use nanopore sequencing for air metagenomics that previously did not exist and has brought many challenges (see, e.g., reviewer #2's comment: "The application of shotgun sequencing using the nanopore platform is novel and challenging."). We have now made our advances clearer throughout the manuscript, and also expanded on the background in our Introduction section.

We now discuss in more detail that our approach is obviously not the only one to assess air monitoring by nanopore sequencing (e.g. p. 23, ll 578-581.: "We emphasize that our sampling, laboratory and computational approaches constitute one feasible and reproducible way of using nanopore shotgun sequencing to profile the air microbiome."), but we highlight the specific challenges of establishing such an approach throughout the manuscript (e.g. low biomass, fragmented DNA for MAG generation, DNA extraction without nanopore blocking given the low biomass, computational taxonomic assignment, etc.) and provide solutions to allow for air monitoring by nanopore sequencing.

We finally hope that the additional very robust results of the first "real-world" application of our protocol clearly showcase how the replicable assessments of an exemplary urban air microbiome can inform future studies and potential public health consequences. We also believe that alone the finding of surprisingly location-specific air microbiomes within a city's boundary will be of substantial interest to the research community.